

DAUGAVPILS UNIVERSITY
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**LICHENS AND ALLIED FUNGI IN LATVIA,
WITH EMPHASIS
ON DEAD WOOD-DWELLING SPECIES IN POST-HARVEST DRY PINE
FORESTS**

**LATVIJAS ĶĒRPJU UN TIEM RADNIECĪGU SĒŅU BIOTAS IZPĒTE, AR
EPIKSĪLO ĶĒRPJU PIEMĒRU SAUSU PRIEŽU MEŽU JAUNAUDZĒS**

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LIST OF ORIGINAL PAPERS / PUBLIKĀCIJU SARAKSTS

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- II.** Motiejūnaitē J., Chesnokov S.V., Czarnota P., Gagarina L.V., Frolov I., Himelbrant D., Konoreva L.A., Kubiak D., Kukwa M., **Moisejevs R.**, Stepanchikova I., Suija A., Tagirdzhanova G., Thell A., Tsuryskau A. 2016. Ninety-one species of lichens and allied fungi new for Latvia with a list of additional records from Kurzeme. *Herzogia* 29: 143–163.
- III.** **Moisejevs R.** 2017. New lichens and allied fungi for Latvia. *Folia Cryptogamica Estonica* 54: 9–12.
- IV.** **Moisejevs R.** & Degtjarenko P. 2017. Four species of saxicolous lichens new for Latvia. *Botanica Lithuanica* 23 (1): 68-70.
- V.** **Moisejevs R.**, Degtjarenko P., Motiejūnaitē J., Piterāns A. and Stepanova D. 2019. New lichens and lichenicolous fungi of Latvia, including the first comprehensive list of lichenicolous fungi. *Lindbergia* 42: linbg.01119.
- VI.** Degtjarenko, P. & **Moisejevs, R.** 2020. Revision of the genus *Cetrelia* (Lichenized Ascomycota) in Latvia. *Botanica*. 26 (1): 88–94.
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3. **Moisejevs, R.**, Balalaikins, M., Valainis, U. *9th International Conference on Biodiversity Research (ICBR)*. 2017. Transplantation of rare species from alleys – is it a compromise during road reconstruction activities? case study in Daugavpils, Latvia. (Poster presentation).
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Zolovs M., Jakubāne I., Kirilova J., Kivleniece I., **Moisejevs R.**, Koļesnikova J., Pilāte D. 2020. The potential antifeedant activity of lichen-forming fungal extracts against the invasive Spanish slug (*Arion vulgaris*). *Canadian Journal of Zoology* 98: 195–201.

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1. Aizstāvēts maģistra darbs (Msc) - **Dace Stepanova**. 2022. Pirmais aizsardzības statusa novērtējums *Parmeliaceae* dzimtas Latvijas īpaši aizsargājamo ķērpju sugām, izmantojot IUCN sistēmas kritērijus / First assessment of the threat status of protected lichenized fungi from the family Parmeliaceae based on IUCN system in Latvia. Daugavpils Universitāte. (Maģistra darba vadītājs).
2. Aizstāvēts bakalaura darbs (Bsc) - **Inese Irbe**. 2020. Aizsargājamo ainavu apvidus "Augšdaugava" trīs upju epilītisko ķērpju biota. Bakalaura darbs. Daugavpils Universitāte. (Bakalaura darba vadītājs).
3. Aizstāvēts bakalaura darbs (Bsc) - **Dace Stepanova**. 2019. Epifītisko ķērpju sugu daudzveidība boreālos priežu sausieņu mežos. Degšanas ietekme. Bakalaura darbs. Daugavpils Universitāte. (Bakalaura darba vadītājs).
4. Aizstāvēts bakalaura darbs (Bsc) - **Samanta Incenberga**. 2018. *Cetrelia* ģints sistemātika un izplatība Latvijā. Bakalaura darbs. Daugavpils Universitāte. (Bakalaura darba vadītājs).

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1. INTRODUCTION

Forests of Northern Europe have drastically changed in the past century due to intensive forest management (Esseen et al. 1997; Siitonen 2001). Young and mid-age forest stands have become more frequent in modern forest landscapes of Northern Europe (UNECE & FAO 2011, Vilén et al. 2012). Such forests differ from natural or old-growth forest ecosystems by having reduced quantity and variability of microhabitats (e.g. snags, logs, old trees etc.) (e.g. Lõhmus & Kraut 2010, Table 2), and by disturbed microclimatic conditions. In particular, the availability of dead wood (e.g., standing and downed coarse and fine woody debris) is reduced by logging and it is considered that in European production forests the current volumes of dead wood are less than 10% of what is found in natural forests (Stokland et al. 2012). Meanwhile species dependent on dead wood constitute 20–25% of all forest-dwelling species, as have been estimated for Fennoscandia (Siitonen 2001).

To reduce negative impacts of logging on biodiversity loss and to ensure sustainable forest management policies, and comply with forest certification standard criteria like FSC (Forest Stewardship Council) and PEFC (Programme for the Endorsement of Forest Certification), different management practices have been applied. Retention of live trees (i.e. green tree retention) and dead wood structures in cut sites is one of the most frequently applied practices in northern Europe (Gustafsson et al. 2010; 2012). Recent studies have shown that retention of woody legacies (fine woody debris, snags and logs) can provide suitable microhabitats for species of various organism groups, like saproxylic beetles, polyporoid fungi and epixylic bryophytes (Hautala et al. 2011; Juutilainen et al. 2014; Floren et al. 2015), as well as for lichens (Lõhmus & Lõhmus 2010; Lundström et al. 2013; Runnel et al. 2013; Hämäläinen et al. 2014; Ranius et al. 2014).

It can be expected that dead wood legacies may provide species rich epixylic lichen assemblages during the first post-harvest decade both in boreal and hemiboreal regions of North-Europe (Caruso & Rudolphi 2009; Runnel et al. 2013). However, the difference in wood dwelling lichen assemblages and/or species composition between stands in boreal and hemiboreal regions may differ (e.g., for burned forests see Lõhmus et al. 2018). In the hemiboreal region, lichen assemblages on dead wood have been representatively studied only in Estonia (e.g., Lõhmus & Lõhmus 2001; Lõhmus & Lõhmus 2011; Runnel et al. 2013), so far.

Forest logging is one the biggest threats for biodiversity, and lately at least several publications address the ways how forest logging impact on lichens and other organisms could be mitigated including retentions of green trees and dead wood in post-logging area (Koivula & Vanha-Majamaa 2020). The biota of lichens and allied fungi of Latvia is one of the least studied among Northern European Countries, at the same time the pressure of forest management activities, such as logging, on wood dwelling lichen species has never been studied in Latvia. Therefore, neither true biodiversity of wood-inhabiting lichens is known, nor there is any data on what forest management measures in production forests would be best to maintain the diversity of lichen biota in Latvia.

1.1. Overview of lichen biota studies in Latvia

The first contributor to the knowledge of lichen biota in nowadays territory of Latvia was J. B. Fischer, the student of Carl Linnaeus, who mentioned seven lichen species in the first volume of “Versuch einer Naturgeschichte von Livland” (Fischer 1778). In the second volume, he already listed twenty-seven species of lichens (Fischer 1791). Later, D. H. Grindel, Latvian botanist, chemist and pharmacist, who had a professor and lecturer position in University of Tartu, reported 38 lichen taxa in “Botanisches Taschenbuch für Liv-, Cur-und Ehstland” including the species descriptions and some ecological data (Grindel 1803). Between 1846 and 1869, K. Müllers and K. Heugels, the members of the Riga Naturalist Society (Naturforscher-Verein zu Rīga), recorded 157 lichen species (Heugel & Muller 1847; Heugel 1855; 1857). However, the main contributor to the knowledge on lichen biota of Latvia (and Estonia as well) in 19th century was A. Bruttan, the inspector and school teacher, who reported 362 lichen taxa from nowadays territory of Latvia (Bruttan 1869; 1870; 1889).

In the first part of 20th century K. S. Mereschkowski, profesor in Kazan University, published the list of lichen species for Baltic region (516 species), which also included information on species from nowadays territory of Latvia (Mereschkowski 1913). An outstanding German-Baltic botanist K. Kupfer formed plant herbarium of Baltic and other countries (including lichens, ca 400 specimens, in *Herbarium Balticum* and *Herbarium generale*), which is stored in herbarium of University of Latvia (RIG). Latvian botanist and mycologist K. Starcs also contributed to the collection of lichen herbaria, by collecting ca 1600 lichen specimens from different countries (650 specimens from Latvia). The publications of Latvian botanist and bryologist N. Malta, devoted to the studies of cryptogams, added new information about lichens on sandstones, with twelve species of lichens new for Latvian lichen

biota (Malta 1925; 1926). Botanists H. Skuja and M. Ore (1935) investigated the biology and ecology of *Coenogonium nigrum* (at present *Cystocoleus ebeneus*), and in 1936, published a review of the lichens of Latvia. In 1939, K. Miške (1939) defended a dissertation of Candidate of Science on the *Cladonia* flora of the Riga area.

In the second part of 20th century, botanist and bryologist A. Āboliņa and mycologist E. Vimba (University of Latvia) published first determination keys for forest lichens and bryophytes (Āboliņa & Vimba 1959). Later, A. Piterāns performed an extensive study on lichens in valley of Daugava river, as the part of this territory at that time was planned to flood by building the hydroelectric station. Alfons Piterāns, botany docent (associate professor) of University of Riga and the only active lichenologist at that time in Latvia, carried out an extensive research of lichen biota in different protected territories of Latvia. Other miscellaneous records of species new for territory of Latvia also were published during the whole period of scientific activity. During mentioned period of time lichenological collection (RIG) deposited in University of Latvia has been replenished on more than 10,000 lichen herbaria units. Also, lichen checklists for various protected territories of Latvia were made. In a year 1982 the first comprehensive checklist of lichens for Latvia was published, which included 464 lichen taxa (Piterāns 1982).

In the late 20th and early 21st century, the contributions to the Latvian lichens were made by visiting and local lichenologists (Motiejūnaite & Piterāns 1998; Piterāns 2001; Sundin & Thor 2001; Motiejūnaitė et al. 2006; Mežaka et al. 2008; Czarnota & Kukwa 2010; Mežaka et al. 2012(a;b); 2014; Jurciņš et al. 2014; Motiejūnaitė & Grochowski 2014; Moisejevs 2015; Mežaka 2015; Gerra-Inohosa & Laiviņš 2016; Mežaka et al. 2018; 2019). An updated checklist of lichens was published by A. Piterāns (2001), which contained 503 species, 5 subspecies and 8 varieties of lichens and allied fungi. In year 2015, the second checklist of lichens was published (Āboliņa et al. 2015), listing 588 taxa of lichens and allied fungi. However, total number of recorded lichen and allied fungi taxa still remains rather low in Latvia, comparing to the other Baltic states (Motiejūnaitė 2017; Randlane et al. 2019.). Such obvious dearth in knowledge on lichen biota, comparing with neighboring Baltic countries, can be explained by low activity of Latvian lichenologists in recent decades.

In year 2016, the report on Latvian forest lichens was published by Gerra-Inohosa & Laiviņš (2016), which was based on literature analysis of previously published papers. Authors concluded that approximately 300 species of lichens are related with forest habitats in Latvia.

At the same time, in neighboring Estonia, the meta-analysis of forest lichen biota, done in year 2003, showed that 481 species of lichens grow in forests. The Estonian study findings were based on 13 quantitative studies (with additional data from 4 studies) carried out in period between 1969-2000 (Lõhmus 2003). In case of Latvia, only seven quantitative studies devoted specifically to the forest lichen biota have been performed till year 2016. Most of these studies were carried out in broadleaved stands, and only epiphytic lichens on live broadleaved trees and aspen were studied (Mežaka et al. 2008; 2012(a); Štikāne et al. 2017). Few data on lichen species in boreal forest sites have been also reported (Mežaka et al. 2015; 2018; 2019). Meanwhile information of dead wood dwelling lichens and allied fungi in Latvia is even more sparse than for epiphytic lichens. Also, young and mid-age regenerated forest lichen biota in Latvia has never been studied systematically before.

1.2. Lichens as bioindicators

Lichens (including allied fungi) and their functional traits are widely used as sensitive indicators of climatic conditions, as their poikilohydric physiology depends on light conditions, water availability (Gauslaa 2014), surrounding temperature etc. (Green et al. 2008; Kranner et al. 2008; Giordani et al. 2012). Thus, they are related to environmental changes as land use (Pinho et al. 2012), forest disturbance (Nöske et al. 2008), forest management (Aragón et al. 2010; Nascimbene et al. 2013); fragmentation (Belinchón et al. 2007; Cardós et al. 2016), forest succession (Koch et al. 2013), air pollution and nitrogen deposition (Giordani et al. 2014; McMurray et al. 2015; Degtjarenko et al. 2018a) or climate change (Nascimbene et al. 2016).

In many Northern European countries, as well as in central part of Europe, lichens are widely employed in woodland key habitat evaluation for conservation purposes (Ek et al. 2002; Gustafsson et al. 2004; Motiejūnaitė et al. 2004). Moreover, the indicator species and their values differ among countries where such evaluations are applied, due to different climatic and vegetation conditions and the distributions of individual lichen species (Ek et al. 2002, Motiejūnaitė et al. 2004). For example: *Evernia mesomorpha* and *Mycoblastus sanguinarius* are habitat specialists in Lithuania (at present absent in NE Poland), *Lobaria virens* is an old growth forest indicator in NE Poland (absent in Baltic States), *Calicium trabinellum* is considered as habitat specialist in Lithuania, but not in Estonia and Latvia (Ek et al. 2002; Motiejūnaitė et al. 2004; Āboliņa et al. 2015; Randlane et al. 2019).

Lichens and allied fungi are good model organisms for evaluation of forest ecological functioning and the effectiveness of conservation measures in forest ecosystems (Tibell 1992;

Wolseley et al. 1994; Brunialti et al. 2010; Miller et al. 2020, etc.). However, as it is expected, the ecological preferences and traits of species and their assemblages differ in wide geographical scales (Stofer et al. 2006; Spribile et al. 2008). For instance, recently it was shown the difference between wood dwelling lichen assemblages and/or species composition in burned forests, between stands in European middle-boreal and hemiboreal region (Lõhmus et al. 2018). However, such studies are few and dearth of general information on lichen distribution applicable to a specific region hinders the application of lichens in monitoring systems.

1.3. Functional traits as the measure of diversity

Traditionally the effect of ecological factors on species diversity is measured by changes in species richness, however, such measure might not reflect the function (viability) of assemblages in studied treatment groups nor general forces driving community assembly (Bässler et al. 2016). Functional traits refer to characteristics of the organisms (such as physiological and/or morphological attributes) thought to be relevant to ecosystem functioning and/or its response to the environment (Díaz and Cabido 2001). Lichen species traits (e.g. photobiont type, growth form, reproduction type, etc.) could indicate lichen community adaptation to environmental conditions (Diaz and Cabido 2001; Giordani et al. 2012; Degtjarenko et al. 2018a). For example, photobiont type, growth form, reproductive structure and presence of lichen substances are showed to be directly related to microclimatic factors (e.g. canopy cover and tree age) and abiotic factors, such as humidity, light availability and temperature (Ellis and Coppins 2006; Pinho et al. 2012; Li et al. 2013; Bässler et al. 2016; Prieto et al. 2017,). Moreover, recently was found, that changes in the environment caused by climate warming and eutrophication can shift the composition and functional diversity of lichens in primeval forests (Łubek et al. 2021b). The importance of the flexibility in the photobiont choice was demonstrated by Ertz et al. (2018), as it enables particular species to use a larger range of tree hosts. Such strategies potentially help lichens to withstand changes of environmental conditions, to widen distribution range and to increase the size of population (Kranner et al. 2008; Marini et al. 2011; Hauck et al. 2013). It is also known that different reproductive modes (e.g. type of vegetative propagules, apotecia/peritecia, pycnidia) are responsible for species' spatial pattern in fragmented forest landscapes (Löbel et al., 2006).

Previously it was shown by Giordani et al. (2012) that forest type and light conditions influence the composition of lichen functional traits; the latter is among the main factors

shaping lichen communities, as shown in previous studies done by Humphrey et al. (2002) and Moning et al. (2009). For example, it was also found that fruticose lichens are clearly ecologically separated from other growth forms, requiring higher light availability (Giordani et al. 2012; Gauslaa 2014). In many forest ecosystems, this growth form is known to be restricted to the higher part of the canopy or to open stands, which provide well-lit conditions (Barkman 1958; Aragón et al. 2016). The studies of Łubek et al. (2021a) demonstrated that different types of substrates can affect the grouping of lichens with similar functional traits. At the same time, authors conclude that not forest community itself, but the availability of specific phorophytes and substrates influences the lichen functional traits' composition in different forest communities.

Forest logging (particularly clearcutting) affects several microclimatic conditions, mostly related to light, temperature and moisture (Heithecker & Halpern 2006; 2007). Thus, different logging activities (including preservation of ecological structures) can have different effect on the lichen assemblages, which is related to the availability or absence of particular environmental conditions that were preserved during the process of logging. Moreover, the effect of different forest management activities on lichen functional traits' composition in various forest communities needs to be studied yet.

1.4. Lichens in forest ecosystems

Forest ecosystems contain up to half of total species richness of lichens and allied fungi known in Baltic region. The most species-rich groups of lichens in lowland forest ecosystems of hemiboreal vegetation region are epiphytic and epixylic lichens (Lõhmus 2003; Gerra-Inohosa & Laiviņš 2016). Lichens are considered a valuable component of forest ecosystems, making important part in nutrient turnover, as well as in providing food and habitat for various organisms, etc. (Ellis 2012; Asplund & Wardle 2017). At the same time, lichens are highly sensitive to environmental changes (Kranner et al. 2008; Pinho et al. 2012) because their physiology is strongly coupled to solar radiation, humidity and temperature conditions (Green et al. 2008; Honegger 2009). Old-growth forest ecosystems, with long ecological continuity, hold the unique and species-rich lichen assemblages, that are documented on old trees and dead wood substrates (particularly on coarse woody debris – e.g. logs and snags) in various studies (e.g. Kuusinen and Siitonen 1998; Lie et al. 2009; Moning et al. 2009; Lõhmus and Lõhmus 2011).

Many studies have shown that various stand-scale and particular structure-scale factors are affecting lichen assemblages and species richness in different old-growth and near natural forest ecosystems (Siitonen 2001; Jüriado et al. 2003; Löbel et al. 2006; Jüriado 2009; Mežaka et al. 2012; Lõhmus & Lõhmus 2011; Asplund et al. 2014, etc). Structure-scale factors affecting lichen richness and assemblages are related with microclimatic conditions and physical properties of substrates. It is known that lichen richness and assemblages on living trees in forest ecosystems are driven by following characteristics of phorophytes: physical characteristics of bark (Barkman 1958; Löbel et al. 2006), bark pH (Barkman 1958; Löbel et al. 2006), age of substrate (Ranius et al. 2008; Caruso & Rudolphi 2009). Also, the particular tree species present in a stand constitute an important factor influencing epiphytic lichen assemblages and total stand-scale species richness (Barkman 1958; Jüriado et al. 2003; Mežaka et al. 2008; Mežaka et al. 2012).

A large proportion of forest-dwelling lichens in boreal forests use dead wood as the main or facultative substrate (Spribille et al. 2008). For example, many wood-dependent lichens in boreal pine forests are confined to old hard wood, such as snags, low natural stumps and logs (Lõhmus & Lõhmus 2011; Santaniello et al. 2017). For dead wood dwelling species, structure-scale characteristics such as fraction of dead wood – e.g. coarse woody debris (CWD) or fine woody debris (hereafter FWD) (Hämäläinen et al. 2015), the decay class of wood (Kruys et al. 1999; Humphrey et al. 2002; Caruso & Rudolphi 2009), standing/lying wood (Hämäläinen et al. 2014) are factors affecting lichen assemblages and species richness. The occurrence of additional natural structures in a forest site, such as stones/cliffs, root plates, substrates submerged in the water also support species richness in forests (Ek et al. 2002; Andersson et al. 2005; Lõhmus et al. 2010; Lõhmus & Lõhmus 2011).

The stand scale characteristics, such as stand age and ecological continuity (Jüriado et al. 2003; Mežaka et al. 2012), canopy cover (Esseen et al. 1997), humidity (Gauslaa 2014), light conditions (Trest et al. 2015), topography (Sevgi et al. 2019) also affect positively the formation of lichen communities and species richness in forests. The main human-induced factors that adversely influence lichen diversity in forests are intensive forest management (Thor 1998; Lõhmus et al. 2019), air pollution (Giordani et al. 2014; Degtjarenko et al. 2016;) and climate change (Moning et al. 2009; Ellis 2012; Nascimbene et al. 2013).

1.4.1. Lichens in post-harvest forest sites

The commercial forestry has been proposed as the most important threat causing the decline of many lichen species (Thor 1998; Nascimbene et al. 2013; Lõhmus et al. 2019). It has been shown that intensive forest management methods, such as clear-cut harvesting, cause drastic changes in forests, resulting in modified and impoverished communities of forest biota (Pawson et al. 2006). As a result, silvicultural alternatives have received more and more attention globally (Puettmann et al. 2015) and sustainable management of production forests has been one of the major goals of modern forestry (MCPFE 2020).

Post-harvest sites usually hold large amounts of coarse dead wood (e.g. logs, snags and stumps), in contrast to managed production forests, where such substrates are rather rare (Siitonen 2001). Also, compared to bark on living tree trunks, with a pH from acid to around neutral, decorticated dead wood is generally an acid substrate, and hosts many specialized lichen species (Crites & Dale 1998; Lõhmus and Lõhmus 2001; Humphrey et al. 2002), indicating its importance for conservation of overall lichen diversity.

Retention forestry – one of the tools of sustainable forest management - was first introduced in the Pacific Northwest (USA) in the 1990s and has since become a widely applied practice, particularly in northern Europe and North America (Gustafsson et al. 2012). It refers to the retention of part of the living trees or dead wood on the harvested site (Gustafsson et al. 2012) in order to provide structural legacies similar to those that occur after natural forest disturbances. These legacies increase the structural variability of the sites (Gustafsson et al. 2010; Kruys et al. 2013; Lõhmus et al. 2013; Rudolphi et al. 2014) and enhance habitat connectivity at a landscape scale (Franklin et al. 1997). Sites with retention trees have been shown to maintain higher levels of species richness and abundance of forest-dwelling species than clear-cut sites with no retained trees (Fedrowitz et al. 2014). Various studies have shown that retention of biological legacies, like living and dead trees, during the process of logging, can provide habitats for specialized epiphytic and epixylic lichens (Lõhmus et al. 2006; Rosenvald and Lõhmus 2008; Lohmus & Lohmus 2010; Caruso et al. 2011; Gustafsson et al. 2013; Lundström et al. 2013; Runnel et al. 2013; Hämäläinen et al. 2014; Ranius et al. 2014; Rudolphi et al. 2014).

The retention of dead wood legacies can support the existence of lichen meta populations, providing substrates for colonization and dispersal sources in the regenerating post-harvest stands (Sillett and Goslin 1999; Hedenås and Hedström 2007; Lõhmus and Lõhmus 2010). In previous studies it was found that level/volume of retention trees can affect the species richness (Hämäläinen et al. 2015) and composition in post-harvest sites. The volume of retained CWD structures in cut-over sites is also meaningful factor for conservation

of epixylic lichens (Caruso et al. 2008). However, there are only few studies dealing with the stand scale factors on lichen assemblages in post-harvest sites (Hämäläinen et al. 2014; 2015; Lõhmus et al. 2018). In case of hemiboreal vegetation region, studies devoted to the lichen assemblages in post-harvest sites lack.

1.4.2. The importance of dead wood structure for lichens conservation in boreal post-harvest sites

The dead wood legacies (e.g. snags, logs, stumps and FWD) and retention trees are the main source of substrata for epixylic and epiphytic lichens in young regenerated post-harvest sites of boreal forest types (Caruso 2008; Hämäläinen 2016; Santaniello et al. 2017). The dynamic process of substrate decay and overgrowing by vascular plants and bryophytes provides suitable microhabitat conditions for different lichen assemblages over the course of time (Fałtynowicz 1986, Lõhmus & Lõhmus 2001; Blasy & Ellis 2014).

In the post-harvest ecosystem, stumps are most abundant woody structure, except the cases when stump harvesting is performed. More-over, stumps, similar substrate as snags in old-growth forest ecosystems, can support lichen species rich microhabitats in first decades after felling in Fennoscandia (Hämäläinen et al. 2014; Santaniello et al. 2017), particularly stumps of Norway spruce (*Picea abies* (L.) Karst.; hereafter “spruce”) and pine (Caruso et al. 2008; Svensson et al. 2013; Hämäläinen et al. 2015). Thus stumps provide “buffering” substrata to maintain populations of dead wood lichens in the managed forests at the landscape scale.

Another specific dead wood substrate created during the felling is cutting residues (i.e. alternative for FWD in old stands). For example, in managed boreal spruce forest, lichen species richness is higher on FWD when equal volumes of coarse woody debris (CWD) and FWD are compared (Kruys & Jonsson 1999). Attached dead branches make a significant proportion of all available dead wood in boreal forests, nevertheless such types of substrata mainly support generalist lichen species (Svensson et al. 2014). Downed fine woody debris (including slash) also can provide suitable microhabitats for a relatively large biomass of dead wood dwelling lichens (Kruys & Jonsson 1999; Svensson et al. 2014; Hämäläinen et al. 2015); though several studies have shown that only few species are present on such substrata (Caruso et al. 2008; Svensson et al. 2016). Nevertheless, there is a lack of studies devoted to the green tree retention effect on lichen species assemblages on FWD.

1.5. Actuality, novelty and aims of the study

Compared to the neighboring countries (see 1.1), the diversity of lichens and allied fungi in Latvia is understudied. Therefore, the **first aim** was to increase the general knowledge about the assemblages of lichens and allied fungi in Latvia (**II–VII**), especially the knowledge on wood-inhabiting species (**I–III**, pro parte).

Standardized methodology of thin layer chromatography (TLC) procedure for detecting and comparing lichen metabolites was elaborated by C.F. Culberson already half century back (Culberson 1972) and have been widely used for lichen species delimitation and determination nowadays (Nash III 2008; Smith et al. 2009). But TLC method was not used during revision of materials in Latvian herbarium before. Therefore, based on TLC method, the **second aim** was to revise genus *Cetrelia* in Latvia, because it includes species having similar morphology, but form chemotypically diverse and difficult complex (**VI**). The determination of secondary metabolite complex in lichen thalli can be used for species identification in genus *Cetrelia*, as it was approved by Mark et al. (2019). The genus has high conservation importance in Latvia and Northern Europe, as all *Cetrelia* species are known to be rare and threatened in all Northern European countries where the species are found.

The **third aim** of the study was to explore the effects of sustainable activities in forest management system (i.e. green tree retention) on lichen species assemblages and species richness in various habitat types (**I**). More specifically, the aim was to describe lichen assemblages on pine dead wood structures (stumps, FWD, snags and logs) in post-harvest 4–6 and 9–11 yr. old dry boreal pine stands in Latvia (**I**) and to find, a) what type of substrata and stand-scale characters (e.g., decay stage and proportion of exposed wood of the substrata, green tree retention level, time since harvest, volume of dead wood) are related to the lichen species richness and composition on pine dead wood legacies (**I**), and b) whether the functional traits of lichens and allied fungi on stumps in post-harvest 4–6 and 9–11 yr. old dry boreal pine stands differ (**in current thesis**).

Statements to be defended:

- Lichen diversity of Latvia requires modern treatment and data supplementation;
- Green tree retention level and age since cut in cut-over sites have impact on stump-dwelling and fine woody debris inhabiting lichen species richness;
- Particular lichen functional trait composition on stumps is changing along the time during the first decade after forest harvest.

Tasks of the study:

- Targeted lichen species diversity research in pre-selected areas to reveal existing lichen data deficiency;
- The revision of the lichen genus *Cetrelia* in Latvia;
- The lichen data collection in post-harvest 4–6 and 9–11 yr. old dry boreal pine stands in Latvia;
- The description of lichen assemblages on pine dead wood structures (stumps, FWD, snags and logs) in post-harvest 4–6 and 9–11 yr. old dry boreal pine stands in Latvia;
- The determination of the collected specimens using adequate lichenological methods applied for species determination.

2. MATERIALS AND METHODS

2.1. Background study of lichen diversity

In order to establish general knowledge level of the diversity of lichen and allied fungi diversity and to check the frequency of forest species, several types of habitats (forests, rocks and outcrops, wooded meadows, tree alleys and parks, mires, etc.) were studied in preselected sites (see Fig. 1 and Fig. 2 (black dots)). In every studied habitat lichens and allied fungi were recorded on all substrata up to 2 meters above the ground. All noteworthy (e.g. Red-listed, protected, rare, and previously unknown from Latvia) lichen species were recorded. For species which can not be determined in field, herbaria material was also collected. All lichen records collected by non-professional lichenologists were verified by author of thesis via visiting locations, and/or identification of collected herbaria material (Fig. 2 (white dots)). All noteworthy lichen records are stored in The Nature Data Management System called “OZOLS”, maintained by Nature Conservation Agency (under Ministry of Environmental Protection and Regional Development of the Republic of Latvia).

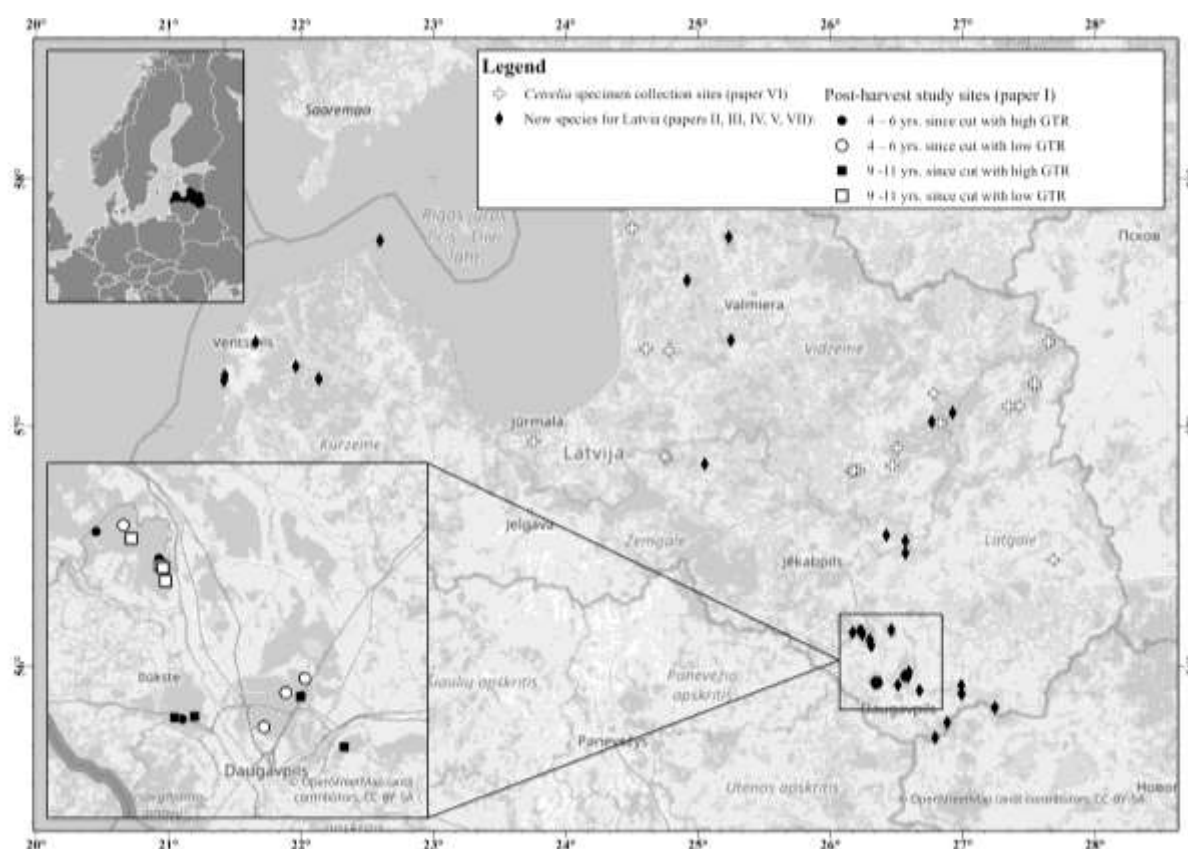


Figure 1. Visited sites where new species for Latvia were recorded (Papers II, III, IV, V and VII), and *Cetrelia* specimen collection sites (from Paper VI). The zoomed map presents study sites of the study I.

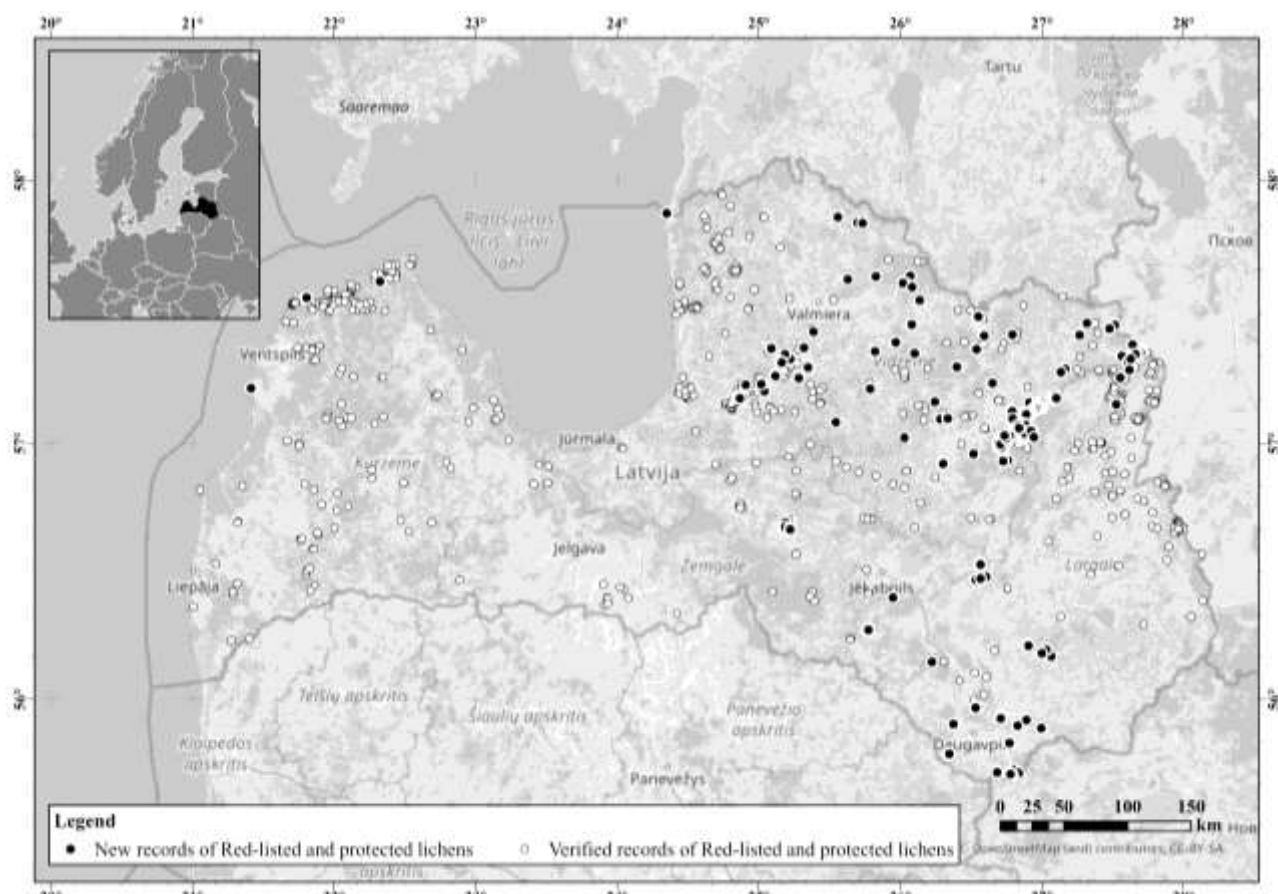


Figure 2. Noteworthy lichen records. Black dots - visited sites for lichen diversity background study, where noteworthy lichen species were recorded. White dots – verified records of noteworthy lichens, visited and/or based on herbaria material.

Spot-test reactions of thalli were checked with sodium hypochlorite solution (commercial bleach) (C) and/or 10% KOH solution (K) in field. Morphology of collected lichen specimens was examined under a dissecting microscope, anatomical structures were studied under a light microscope. At the laboratory, ethanol solution of paraphenylenediamine (PD), Lugol solution (I), commercial bleach) (C), 10% KOH solution (K) and 50% HNO₃ (N) were employed for additional spot tests and anatomical examinations. For the specimens, requiring secondary metabolite examination thin layer chromatography (TLC; solvent systems A and C) was employed (Orange et al. 2001). The collected specimens are stored in the lichenological herbarium of the University of Daugavpils (DAU).

The revision of genus *Cetrelia* in Latvia was based on herbarium material collected in territory of Latvia from 1957 to 2018 and stored in the lichenological herbarium of Daugavpils University (DAU) and the University of Latvia (RIG), as the majority of specimens of genus *Cetrelia* collected in Latvia are stored in mentioned herbarium collections. All specimens were examined for lichen substances by thin-layer chromatography (TLC; solvent A) (Orange et al.

2001). Localities of Latvian specimens were transcribed from the labels and distribution maps of *Cetrelia* species were created using ESRI ArcGIS pro 2.3. (ESRI 2011).

2.2. Lichen assemblages on dead wood in dry boreal post-harvest stands

In Latvia, the proportion of forest land makes about half of the terrestrial area of the country and the proportion of coniferous forests is ca 53%. Pine is the most frequent tree species in Latvia and approximately fourth part of dry boreal pine forests in Latvia are young-regenerated sites (Anonymous 2020). Therefore, study (I) focused on dry boreal post-harvest sites.

The study of lichen assemblages (I) in dry boreal post-harvest sites was performed in the SE part of Latvia, in the hemiboreal forest zone (according to Ahti et al. 1968). The study sites were selected from Vacciniosa type (Bušs 1976), i.e. pine dominated forest on higher fluvio-glacial landforms and till mounds with Podzols (pH 3.5–5.0) having continuous moss cover and species-poor undergrowth (mostly shrubs of *Vaccinium vitis-idaea* L. and *Calluna vulgaris* (L.) Hull.) and a very sparse under-storey. The study sites were selected in 2.5–9 ha size retention-cut stands situated outside of protected areas (Fig. 2, zoomed map).

Among the study sites, four treatment types were designated, based on combinations of time since harvest (4–6 and 9–11 years) and green tree retention level (“high”, 14–20 trees per ha or 30–40 m³/ha, and “low”, ≤6 trees per ha or ≤12 m³/ha). Each of the four treatment combinations (hereafter as “treatment type”) was represented by four sites, 16 sites in total. All sites were standardized to 2.5 ha in size, with 30 m wide buffer zones on each side. In all stands pine saplings have been planted, and all stands were state-owned. Total lichen sampling from woody structures per site were performed from two FWD plots (2 × 8 m) and 4–6 stumps selected along sampling transects and from 1–2 logs and 1–2 snags selected arbitrary over the whole study site (Fig. 3 and Fig.4). Lichen identification was performed employing same methods as described in the subchapter 2.1.



Figure 3. Lichen sampling on well-decayed stump in 9-11 yrs old post-harvest sites (left).

Figure 4. Well decayed stump from 9-11 yrs old post-harvest study site (right).

2.2.1. Assessment of functional traits

For each lichen and allied fungi species found on the stumps (Paper I), ten functional trait types among morphological, anatomical, chemical and reproduction traits were assessed (Table 1). The information related to these traits was obtained from key-books, taxonomic literature and literature about lichen compounds (Tibell 1999; Jørgensen et al. 2007; Smith et al. 2009; Thell & Moberg 2011; Ahti et al. 2013; Nguyen et al. 2013; Wirth et al. 2013).

Table 1. Description of functional trait types and their trait characteristics distinguished.

| Functional trait | Description of the trait levels |
|---------------------|--|
| Thallus type | Following trait groups were distinguished: 1) no thallus (non-lichenized); 2) crustose (subtypes: 2a leprose and granulose without cortex; 2b crustose with cortex; 2c endosubstratic, like have calicioids, |

| Functional trait | Description of the trait levels |
|-------------------------------|--|
| | <p>Micarea spp.); 3) foliose and squamulose; 4) fruticose (subtypes: 4a Cladonia-type thallus; 4b pendulous; 4c) erect fruticose).</p> <p>Foliose and squamous thallus type were combined in the dataset and analyzed as one trait type, due to absence of variation in species with squamous thallus type.</p> |
| Thallus color | <p>The trait of thallus colour had six groups based on thallus colour in dry conditions and on additional information of substances responsible for pigmentation of the thallus: 1) no colour (e.g. endosubstratic thallus); 2) grey thallus (includes atranorin); 3) greenish to greenish grey (thallus do not contain atranorin and usnic acids); 4) yellowish green (includes usnic acids); 5) lemon yellow (includes pulvinic acid derivatives), 6) brownish (do not contain acetone-soluble substances).</p> <p>There were no species with anthraquinone-defined thallus colour. Only one species (<i>Vulpicida pinastri</i>) represented lemon yellow thallus colour and was therefore combined with yellowish green species (group 4) in statistical analysis, but represented separately in results with descriptive statistics. Non-lichenized species do not have current trait value.</p> |
| Lichen substances | <p>Acetone-soluble lichen substances in lichen thalli or apothecia as functional trait was defined based on their 1) absence or 2) presence. Non-lichenized species do not have current trait value.</p> |
| UV-screening compounds | <p>The production of UV screening compounds by lichens, such as phenolic compounds (depsidones, depsides, diphenyl ethers), anthraquinones, xanthonones or shikimic acid derivatives (calycin, mycosporines, scytonemin) etc., serve as an important protective mechanism of lichens (Nguyen et al. 2013).</p> <p>The trait was defined based on the 1) absence or 2) presence of acetone-soluble UV-protective substance(s) in particular lichen species thalli or apothecia.</p> |

| Functional trait | Description of the trait levels |
|--------------------------|---|
| Photobiont type | There was no variation of photobiont types in the dataset, as species with cyanobacteria or <i>Trentepohlioid</i> algae were not presented. Therefore following basic trait values were distinguished: 1) taxa without photobionts (e.g. <i>Sarea resinae</i> , <i>Mycocalicium subtile</i>) and 2) taxa with Chlorococcoid algae (e.g. <i>Treuboxia</i> ; <i>Chlorococcus</i> , <i>Coccomyxa</i> ; <i>Dictyochloropsis</i> , <i>Stichococcus</i>). |
| Reproduction type | Following reproduction types were distinguished: 1) sexual (sub-types: 1a apothecia; 1b stalked apothecia/calicioids); 2) vegetative (sub-types: 2a soresdia; 2b isidia); 3) mixed reproduction (e.g. both vegetative and sexual types occur). There was no variation within asexual (conidia) type of reproduction, same as for species forming perithecia. |
| Spore type | Spore type was distinguished based on the cell numbers forming the ascospore: 1) one-celled spores; 2) two-celled spores; 3) spores with more than 2 cells (including mural spores). |
| Spore length | Spore length trait based on maximal length of spore typical to certain species and the trait was divided into three groups: 1) short spores (<10 μ); 2) medium size spores (10-20 μ); 3) long spores (>20 μ). |
| Spore colour | Spore colour trait based on the presence of the pigmentation and had two groups: 1) hyaline spores and 2) coloured spores. |
| Spore surface | The trait corresponds to the presence of the ornamentation on the spores (facilitates the dispersal); because of low variation in our dataset, the trait had two groups: 1) spores smooth and 2) spores with ornamentation. |

2.3. Statistical data analysis

In the Paper (I) the number of sampled logs and snags among the sites was insufficient for statistical analysis, but the difference of functional area of stumps and volume of CWD and FWD across four treatment combinations were tested with a one-way Welch ANOVA with Games-Howell post hoc analysis (before the analysis two outliers were removed to achieve normal distributions for each treatment). Homogeneity of variances was violated, as assessed by Levene's Test of Homogeneity of Variance ($p < 0.001$). General linear models (GLMs) were used to study the effects of post-harvestage and green tree retention level on total and mean lichen species richness on stumps and on FWD per site.

The fixed factors were age of cut sites (4–6 and 9–11 yr. since cut) and GTR level (≤ 6 trees/ha as “low” and ≥ 14 trees/ha as “high”). Two un-correlated variables were also included per each model: for models (1) and (2) variables average stump functional area and average stump diameter per site, and for model (3) variables volume of FWD and CWD. GLMs were run initially including all factors and interaction between site age and GTR level. Non-significant factors were removed from the model manually one by one, commencing with the least significant factor. Difference of lichen assemblage composition among treatment types (depending on site age and GTR level) were studied for 1) stumps (full unit), 2) stumps vertical and horizontal surfaces separated, and 3) FWD. A multi-response permutation procedure (MRPP) was used to test if Sørensen (Bray-Curtis) dissimilarities among predefined groups exceed those resulting from random assignment of sample units to those classes, and it has the advantage of not requiring distributional assumptions that are seldom met with ecological assemblage data. To correct the p-values for multiple comparisons, a Bonferroni correction was applied.

In the species matrices, frequency score of the species per site (0–4 for stumps and 0–2 for FWD) were used as input. Results based on the full species matrix are presented (i.e. MRPP analysis gave similar results to full and reduced species matrices). The full species matrix was also used to find species specific to treatment type using indicator species analysis (ISA; Dufrêne & Legendre 1997). In that analysis, we distinguished the species with significant ($p < 0.05$) and relatively high ($> 50\%$) indicator values.

For functional traits the assumption of data distribution was assessed by the Shapiro-Wilk test and inspection of the normal Q-Q plots. The assumption of homogeneity of variance was tested by the Levene's test. The two-way ANOVA or aligned ranks transformation

ANOVA (depending on inspection of assumption) was conducted to determine whether there is an interaction effect between two independent trait variables (age since cut and trait type) on a continuous dependent variable (species number). Descriptive statistics presented as mean \pm standard deviation if data met the assumptions of normal distribution and equality of variance, whereas the median is presented if those assumptions were violated.

3. RESULTS

3.1. Background study of lichen diversity

Altogether 69 new lichen and 47 allied fungi species were registered and reported as new for Latvia during the 6-year period (2015 – 2020) of lichen diversity studies presented in the thesis (**I-VII**; Appendix 1). Of these ca. 20% of species were found and/or identified by author of thesis.

Newly recorded species of lichens and allied fungi for Latvia are represented by 86 genera, and 53 families. Of these, Teloschistaceae (10 species) and Verrucariaceae (5 species) are the most represented families among the newly recorded species. Several lichen families are recorded for a first time in Latvia – for example: Carbonicolaceae, Trapeliaceae, Pycnoraceae, Xylographaceae.

Ninety of 116 newly recorded species were found in forest habitats, 14 species in parks/wooded meadows and roadside alleys, eight on rocks and outcrops and four species in disturbed/ruderal habitats (mainly on soil). Most of newly reported taxa were epiphytic and epixylic lichens, but the number of lichenicolous fungi (37 species) was notable too. Out of newly recorded species 39 were found on tree bark, 24 species were found on dead wood; eight species on sand/soil; eight on stones and man-made calcareous substrates, two species on conifer resin, two on roots of windthrows, one on sandstone.

According to The Nature Data Management System called “OZOLS”, maintained by Nature Conservation Agency (under Ministry of Environmental Protection and Regional Development of the Republic of Latvia), 1842 records of 41 noteworthy lichen species were collected in 6-year period (2015-2020), and 754 records of 37 noteworthy lichen species were verified by author of thesis (Anonymous 2021). In addition, c.a. 2000 lichen and allied fungi herbaria units which represent 227 species of lichens and allied fungi are collected and positioned in lichenological herbarium of Daugavpils University (DAU).

3.2. The revision of genus *Cetrelia* in Latvia

In total, 98 specimens (collected between 1957 - 2018) were examined from two herbarium collections (RIG and DAU), as these collections contain the majority of *Cetrelia* specimens collected in Latvia. Of these 19 specimens were found by author of thesis. All examined material was sterile (without apothecia). From the examined specimens *Cetrelia cetrarioides* and *C. olivetorum* accounted for 43% and 42% while *Cetrelia monachorum*

appears to be the rarest – with 15% of the examined specimens. In this study, *Cetrelia monachorum* was reported for the first time for Latvia. The studied specimens of genus *Cetrelia* mostly were found on *Alnus glutinosa*, *Fraxinus excelsior*, *Populus tremula*, *Tilia cordata*, *Quercus robur* and less frequently (7% of specimens) on *A. incana*, *Betula pendula*, *Corylus avelana*, *Padus avium* and *Picea abies* (VI).

3.3. Lichen richness on dead wood in post-harvest sites

Fifty-four species of lichens and allied fungi were recorded in young regenerated *Vacciniosa* forest type from all sampled substrata types (FWD, snags, stumps, logs) from 16 study sites (detailed list see in Appendix 1 Paper I). Most of the species found are common in Latvia, but three were recorded for the first time in the country, namely, *Calicium trabinellum* (on log and stump dead wood), *Pycnora sorophora* (on stump bark and dead wood), and *Scoliciosporum sarothamni* (on fresh FWD) (I, III). Stumps (n = 64) were inhabited by 48 species. FWD (n = 32 subplots) was inhabited by 43 species. On snags (n = 9) 27 species were recorded, and on logs (n = 24) 44 species. In addition, one Red-listed lichen in Latvia – *Cladonia parasitica* – was found in seven sites (all 9–11 yr. post-harvest), growing on horizontal surfaces of stumps and logs (I). However, stump characteristics varied only slightly among cut site treatments (see Table 2); also volumes of FWD and CWD were relatively low and similar among site treatments (p = 0.454 and p = 0.876, respectively; Table 2 and Appendix 2 in Paper I).

Table 2. Characteristics of stumps, downed fine and coarse woody debris (FWD and CWD, respectively) in cut sites of different age and green tree retention (GTR) level (high ≥ 14 and low ≤ 6 trees/ha). Within each site combinations 16 stumps were sampled per study plot.

| Age | 4-6 yr. | | 9-11 yr. | |
|---|-----------------|----------------|----------------|----------------|
| | High (n=4) | Low (n=4) | High (n=4) | Low (n=4) |
| GTR | | | | |
| Stumps | | | | |
| Heigh (cm) | 18.4 \pm 5.6 | 20.4 \pm 2.3 | 20.3 \pm 3.1 | 21.6 \pm 3.2 |
| Diameter (cm) | 37.1 \pm 8.3 | 29.2 \pm 4.0 | 31.4 \pm 5.8 | 28.8 \pm 7 |
| Horizontal area (dm ²) | 11.3 \pm 4.5 | 6.8 \pm 1.7 | 8.0 \pm 2.8 | 6.9 \pm 3.1 |
| Vertical area (dm ²) | 22.1 \pm 10.5 | 18.7 \pm 3.3 | 20.2 \pm 5.7 | 19.6 \pm 5.4 |
| Functional horizontal area (dm ²) | 9.2 \pm 4.1 | 6.0 \pm 1.4 | 5.4 \pm 2.6 | 5.3 \pm 2.6 |

| | | | | |
|---|-----------|------------|-----------|-----------|
| Functional vertical area (dm ²) | 16.4±9.6 | 14.8±3.4 | 11.7±5.1 | 13.2±4.4 |
| Volume (m ³) | 0.02±0.01 | 0.01±0.004 | 0.02±0.01 | 0.02±0.01 |
| FWD | | | | |
| Volume (m ³ /ha) | 0.77±0.22 | 0.91±0.28 | 1.14±0.43 | 0.96±0.27 |
| CWD | | | | |
| Volume (m ³ /ha) | 0.36±0.18 | 0.30±0.18 | 0.32±0.17 | 0.32±0.21 |

Based on the final GLM model, the total and mean species richness on stumps depended on cut site age (higher richness in 9–11 yr. sites) and on GTR level (higher richness in sites with “high” GTR level; Table 2 in Paper (I), Fig. 5 A). The interaction between age and GTR level, as well as stump-scale variables (average stump functional area and diameter per site) were non-significant. Any of tested variables had no significant impact on FWD total species richness (Results in Paper (I) Fig. 5 B).

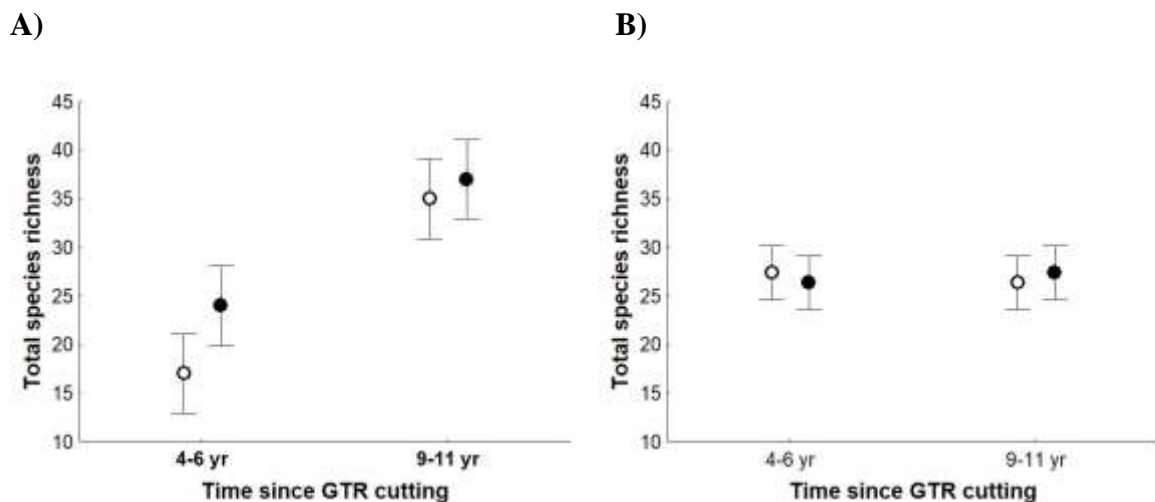


Figure 5. Total species richness on pine stumps(A) and fine woody debris (B) on 4–6 and 9–11 yr. post-harvest sites with low (≤ 6 trees/ha; open symbol) and high (≥ 14 trees/ha; filled symbol) green tree retention (GTR) level. For each treatment of post-harvest sites 16 stumps and eight 2x8m subplots for fine woody debris were surveyed.

3.4. Lichen assemblage composition on stumps and FWD in post-harvest sites

Lichen assemblages on stumps differed between age groups of cuts within the same GTR level (MRPP test; $A = 0.2$ and 0.4 , $p = 0.006$ for both) and between high and low GTR level sites of the same age ($A = 0.2$, $p = 0.007$ for both). Such groupings were slightly visible in the NMS ordination (Fig. 6 A). Among tested environmental variables average functional

area of the stump correlated with the second ordination axis ($r^2 = 0.24$). Lichen assemblages on FWD differed between age groups of cuts within the same GTR level ($A=0.1$, $p=0.01$ for both cases), but high and low level GTR sites had similar assemblages within the same cut age (for 4–6 yr. cut sites $A=0.05$, $p=0.06$ and for 9–11 yr. sites $A=0.04$, $p=0.10$). This pattern is visible also on NMS graph (Fig. 6 B) and none of the tested environmental variables (volume of FWD and CWD) correlated with ordination axes ($r^2 < 0.2$).

Similar assemblage patterns (age difference and dependence on GTR level) emerged also if stump vertical and horizontal surfaces were analyzed separately ($A=0.1–0.2$, $p=0.02–0.01$). In addition, vertical and horizontal surfaces differed from each other within the age \times GTR level treatment (MRPP test $A=0.2–0.5$, $p=0.01$; Fig. 6 B). Focus on horizontal stump surface only (as the case of post-harvest colonization) showed clear difference of assemblages between 4–6 yr. and 9–11 yr. cut sites both on wood and on cut bark around the stump ($A=0.5$, $p<0.001$). Focus on assemblages on wood only revealed the difference between horizontal and vertical surfaces of stumps both on 4–6 yr. ($A=0.4$, $p<0.001$) and 9–11 yr. ($A=0.3$, $p<0.001$) cut sites.

The indicator species analysis (ISA) showed that 19 stump-dwelling species (seven of them from genera *Cladonia*) associated significantly with 9–11 yr. cut sites, four of them showing also the preference for GTR level (like *C. parasitica* on sites with lower GTR level, and *C. chlorophaea* on sites with higher GTR level). At the same time, only five species (*Bryoria fuscescens*, *Platismatia glauca*, *Sarea resinae*, *Scoliciosporum chlorococcum* and *S. sarothamni*) showed affinity for FWD (Paper I, Appendix 3).

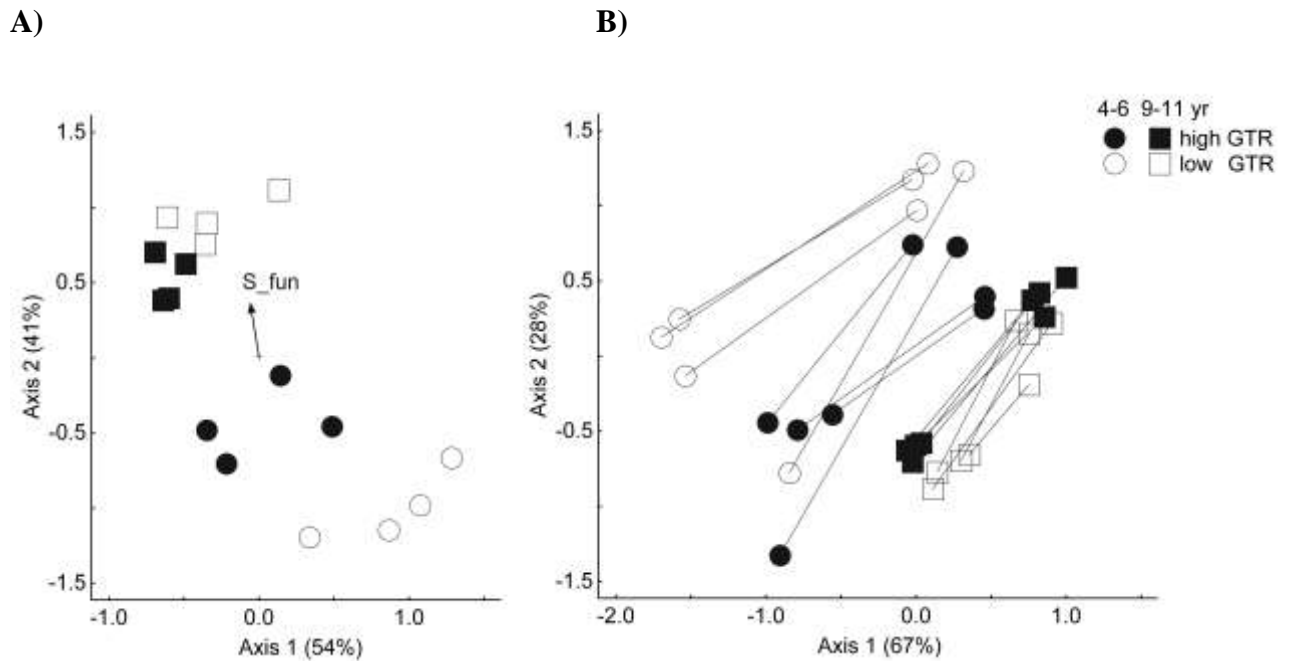


Figure 6. Total composition of lichen species assemblages on pine stumps ($n = 4$ stumps per site) on young and old (4-6 and 9-11 years, respectively) post-harvest sites with high (14-25 trees/ha) and low (≤ 6 trees/ha) green tree retention (GTR) level (A). Vector on the graph indicates that functional area of stumps (i.e. stump area not covered by bryophytes, S_{fun}) is related to the second ordination axis. Graph (B) illustrates lichen assemblages separately on vertical and horizontal surfaces of pine stumps (joint symbols on the left and right, respectively).

3.3.2. Distribution of lichen functional traits in post-harvest sites

The distribution of ten functional trait types and their characteristics within the set of 47 species of lichens and allied fungi found on stumps in young (4-6) and old (9-11) post-harvest sites is presented in the Table 3. Among lichenized species the dataset included only chlorolichens (i.e. no species with cyanobacteria or trentepohlioid algae) and dominated species with thallus that had grey or green to greenish grey colour; most of lichenized species contained also acetone soluble lichen secondary compounds. Prevailing reproduction types within the dataset of stump inhabiting species was vegetative. Among sexually dispersing species spore traits did not have large variation in the dataset: most had one-celled spores and species with hyaline spores dominated over species with coloured spores, as well as more common were species with smooth surface and species with medium size spores (10-20 μ).

Based on results of Two-way ANOVA, the distribution of trait characteristics within trait types “Thallus type”, “Thallus colour” and “UV substances” differed between young (4–6 yr.) and old (9–11 yr.) post-harvest sites ($(F_{(2,42)} = 9.921, p < 0.001, F_{(3,56)} = 11.338, p < 0.001$ and $F_{(1,28)} = 26.468, p < 0.001$ respectively)). Based on Tuckey HSD Post Hoc test, old sites presented higher mean richness of species with crustose and fruticose thallus ($p < 0.001$; Fig. 7 A), species having greenish to greenish grey and yellowish green thallus colour ($p < 0.01$; Fig. 7 B) and species with UV-protective substances ($p < 0.001$; (Fig. 8 B).

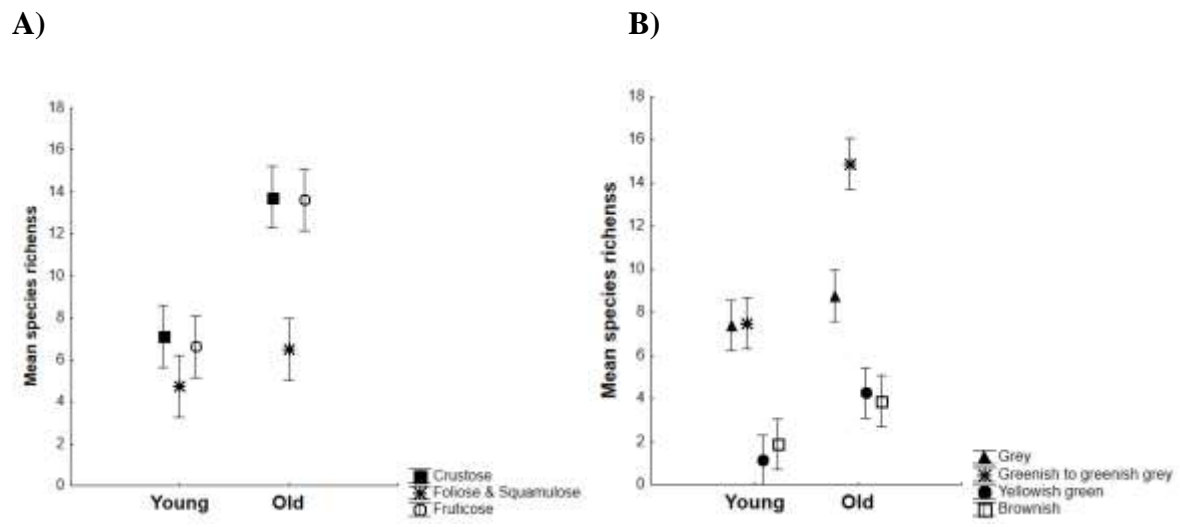


Figure 7. Mean species richness (with standard error) of different thallus type (A) and thallus colour (B) traits on stumps in 4–6 and 9–11 yr. post-harvest sites. Sample sizes see Table 3.

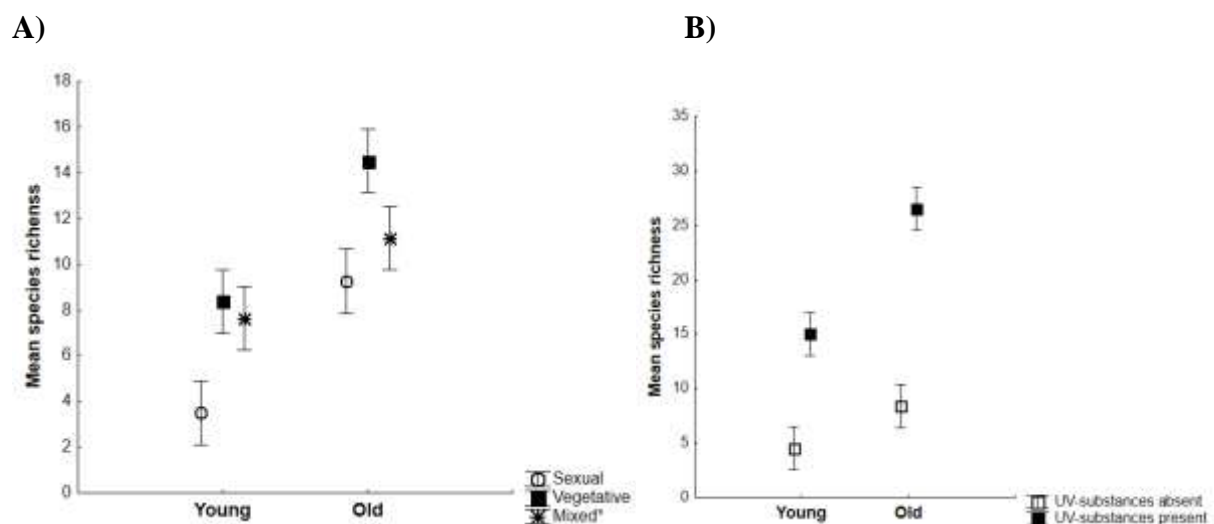


Figure 8. Mean richness of species (with standard error) according to their reproduction traits (A), and presence/absence of UV-protective substance(s) (B) on stumps in 4–6 and 9–11 yr. post-harvest sites. * - Two types of reproduction (vegetative and sexual) are applicable for one species. Sample sizes see Table 3.

Table 3. Studied lichen functional trait types and distribution of their traits in the set of lichen and allied fungi (in total, 47 species) found on Scots pine stumps in young (4-6 yr.) and old (9-11 yr.) post-harvest sites (per age treatment N = 32 stumps from 8 sites).

| Functional trait types and traits | Number of species | | | Mean±SD | |
|--|-------------------|-----|-------|---------|--------|
| | Young | Old | Total | Young | Old |
| Thallus type (N = 47 species) | | | | | |
| Non-lichenized | 1 | 1 | 1 | - | - |
| Crustose | 12 | 18 | 18 | 7±1.6 | 14±2 |
| <i>Leprose and granulous</i> | 4 | 8 | 8 | 3±0.6 | 6±1.1 |
| <i>Crustose with cortex</i> | 5 | 6 | 6 | 2±0.9 | 5±0.7 |
| <i>Endosubstratic (calicioids, Micareas)</i> | 3 | 4 | 4 | 2±0.6 | 3±1 |
| Foliose and squamulose | 8 | 9 | 10 | 5±1 | 7±0.9 |
| Fruticose | 14 | 18 | 18 | 7±3.9 | 14±1.4 |
| <i>Cladonia-type</i> | 9 | 11 | 11 | 4±2.5 | 10±0.5 |
| <i>Pendulous</i> | 0 | 1 | 1 | - | - |
| <i>Erect fruticose</i> | 5 | 6 | 6 | 2±1.5 | 3±1.3 |
| Thallus color (N = 47 species) | | | | | |
| No colour, thallus endosubstratic | 2 | 3 | 3 | - | 2±0.5 |
| Non-lichenized | 1 | 1 | 1 | - | - |
| Grey (atranorin) | 12 | 13 | 14 | 7±1.9 | 9±1.5 |
| Greenish to greenish grey** | 14 | 19 | 19 | 7±3.2 | 15±2.2 |
| Yellowish green (usnic acids) | 3 | 5 | 5 | 1±0.8 | 4±0.7 |
| Lemon yellow (pulvinic acid derivatives) | 1 | 1 | 1 | - | - |
| Anthraquinones | 0 | 0 | 0 | - | - |
| Brownish (acetone-non-soluble substances) | 3 | 5 | 5 | 2±0.4 | 4±0.6 |
| Substances (N = 47 species) | | | | | |
| No acetone-soluble lichen substances | 5 | 5 | 5 | 2±0.9 | 4±0.9 |
| Acetone-soluble lichen substances present | 30 | 41 | 42 | 17±5 | 31±2.7 |
| UV-substances (N = 47 species) | | | | | |
| UV-protective substances absent | 7 | 10 | 11 | 5±1.1 | 8±1.4 |
| UV-protective substances present | 27 | 35 | 36 | 15±4.7 | 27±2.1 |
| Reproductiot type (N = 47 species) | | | | | |
| Sexual | 8 | 12 | 12 | 4±1.3 | 9±1.7 |
| <i>Apothecia</i> | 5 | 8 | 8 | 2±0.9 | 6±1.3 |

| | | | | | |
|---------------------------------------|----|----|----|--------|--------|
| <i>Stalked apothecia (calicioids)</i> | 3 | 4 | 4 | 1 | 3 |
| <i>Perithecia</i> | 0 | 0 | 0 | - | - |
| Asexual (pycnidia) | 0 | 0 | 0 | - | - |
| Vegitative | 15 | 20 | 20 | 8±3.3 | 15±0.9 |
| <i>Soredia</i> | 12 | 17 | 17 | 7±2.6 | 13±1 |
| <i>Isidia</i> | 3 | 3 | 3 | 1 | 1 |
| Mixed *** | 12 | 14 | 15 | 8±2 | 11±1.6 |
| Spore type (N = 47 species) | | | | | |
| One-celled spores | 16 | 20 | 20 | 9±2.5 | 17±1.7 |
| Two-celled spores | 4 | 6 | 7 | 2±1 | 4±1.6 |
| Spores at least three-celled or mural | 0 | 0 | 0 | - | - |
| Spore length (N = 47 species) | | | | | |
| <10 µ | 4 | 4 | 4 | 3±0.6 | 3±1 |
| 10-20 µ | 15 | 21 | 21 | 8±2.4 | 17±2 |
| >20 µ | 1 | 1 | 2 | - | - |
| Spore colour (N = 47 species) | | | | | |
| Hyaline spores | 16 | 21 | 21 | 10±2.3 | 17±2. |
| Colored spores | 4 | 5 | 6 | 1 | 3 |
| Spore surface (N = 47 species) | | | | | |
| Smooth surface | 17 | 22 | 23 | 10±2.4 | 17±2.7 |
| Ornamented surface | 3 | 4 | 4 | 1 | 3 |

**Do not contain atranorin and usnic acids

*** Two types of reproduction (vegetative and sexual) are applicable for one species

4. DISCUSSION

4.1. Contributions to the knowledge on lichen and allied fungi biota of Latvia

The total number of lichen and allied fungi species known in territory of Latvia increased by 116 taxa in 6-year period of the author's (co-)studies (**I-VII**). At present, the total biota of lichens and allied fungi counts ca. 700 species in Latvia. Most of newly recorded species are epiphytic and epixylic lichens –species associated mainly with old-growth forest habitats. Current contributions also indicate the lack of research focused on the diversity of certain species groups in Latvia – e.g. wood dwelling lichens, lichenicolous fungi, saprobic calicioid species, saxicolous lichens etc., as most of newly recorded species belong to the above mentioned groups (Papers **II-V** and **VII**). Out of newly recorded species, several species of lichens are potentially rare and perhaps threatened. For example, such forest species like *Chaenotheca laevigata*, *Cetrelia monachorum*, *Microcalicium ahlneri*, *Parmelia submontana*, are associated with old-growth forests, which are declining in all Baltic countries. Species like *Dermatocarpon miniatum* and *Pilophorus cereolus* grow in unpolluted streams and on undisturbed boulders and sandstones, a habitat type not frequent in the eastern Baltic region. All above mentioned species are rare in Estonia and Lithuania, and considered as threatened (Lõhmus et al. 2019; J. Motiejūnaitė pers. comments). Their situation in Latvia most probably is similar to neighboring countries. At the same time various of newly recorded lichen species like *Leptorhaphis epidermidis*, *Ochrolechia microstictoides*, *Pycnora sorophora*, as well as allied fungi species like *Clypeococcum hypocenomyces* and *Sarea difformis* are probably overlooked in Latvia.

Most of newly recorded species were collected and/or identified by visiting lichenologists (co-authors of Papers **II** and **V**), which is expectable, and demonstrates the importance of foreign specialist involvement with specific knowledge on particular lichen groups. Similar inference was done by Vondrák et al. (2016), who demonstrated that involving more lichenologists with different field experience of lichens in specialized niches is more effective for obtaining more complete species lists. For example, the lichenological meetings with field visits, carried out in Baltic countries in previous years, like Baltic Mycologist and Lichenologist symposium, Nordic Lichen Society meetings, in most cases bring together professionals and amateurs, and it notably raise the knowledge about biodiversity of the region (Motiejūnaitė & Piterāns 1998; Motiejūnaitė et al., 2006; Motiejūnaitė et al. 2012; Paper **II**; Suija et al. 2020, etc).

The general knowledge on several Red-listed, protected, and rare lichen species distribution in the territory of Latvia was noticeably increased. For example: till year 2015, 1052 records of lichens were entered in “OZOLS” while in year 2020 number of Red-listed, protected and rare lichens records was over 4000. Moreover, for several species the knowledge on distribution in territory of Latvia has essentially increased, as an example: only few localities of Red-listed and protected lichen *Mycoblastus sanguinarius* (L.) Norman were known till year 2015, while at the moment of preparation of current thesis (year 2020) over 20 localities were already known. Similar situation is with *Thelotrema lepadinum* (Ach.) Ach. with less than 20 localities known previously, and over 60 localities known at present. However, some rare lichen species like *Bactrospora dryina* (Ach.) A. Massal., *Chaenotheca cinerea* (Pers.) Tibell, *Parmeliella thriptophylla* (Ach.) Müll. Arg., *Protopannaria pezizoides* (Weber) P.M. Jørg. & S. Ekman still have very limited number of known localities, despite the fact that these species can be rather easily recognised in field.

Relative dearth in general knowledge of lichen diversity in Latvia can be explained by the lack of systematical studies of lichens and allied fungi (such as species monitoring programs, targeted ecological studies in different habitat types etc.) and low number of experts capable to identify lichen species in the country. The absence of systematical revisions of lichen herbaria collections, similar as it was done in neighboring Estonia, Lithuania, Belarus (Kukwa & Motiejūnaitė 2012; Motiejūnaitė et al. 2013; Tsurykau et al. 2015; Tsurykau & Golubkov 2015; Motiejūnaitė 2017; Tsurykau et al. 2017; Tsurykau et al. 2018; Randlane et al. 2019; Yatsyna 2020, etc), brings additional lag in knowledge of Latvian lichen biota. Previously, species identification of Latvian lichen material, where also thin-layer chromatography (TLC) methods were applied if needed, was used only in few studies. In most cases, except for the study performed by Mežaka et al. 2012 (b), TLC methods were applied only for miscellaneous collections of visiting experienced lichenologists (e.g., Czarnota & Kukwa 2010). Various lichen genera and groups, in which species determination without TLC methods is problematic, or even impossible, were never revised in Latvia (e.g. *Lepraria*, *Ochrolechia*, *Xanthoparmelia*, various species of genus *Cladonia*, sterile crustose lichens, etc.).

The revision of Latvian herbaria on genus *Cetrelia* (VI) was the first targeted revision of lichens using TLC methods. Until therevision, only two species, *Cetrelia olivetorum* and *C. cetrarioides*, were known in Latvia (Āboliņa et al. 2015), while the revision of herbarium material added the third species – *C. monachorum*. Same three species occur also in neighboring Estonia, Belarus and Lithuania (Kukwa & Motiejūnaitė 2012; Bely et al. 2014;

Degtjarenko et al. 2018b; Randlane et al. 2019). In most countries of Northern Europe all species of genus *Cetelia* are threatened and protected (VI). So far, only *C. olivetorum* and *C. cetrarioides* are strictly protected by local (Latvian) legislation. The results of Paper (VI) show that *C. monachorum* is as rare as *C. cetrarioides* s.str. and *C. olivetorum* s.str. During the upcoming assessment of red-listed species using IUCN criteria, it is suggested to perform the assessment for each species separately, as it was done in case of Estonia (Lõhmus et al. 2019). However, it is suggested to include *C. olivetorum* s.lat. as one conservation unit for legislative lists. Such approach will help to simplify the practical protection of *Cetrelia* species. At the same time, all species of genus *Cetrelia* should be added to monitoring programmes of rare species separately, as it can improve the understanding of the actual situation with each species.

Despite the extensive contribution brought to the general knowledge of lichens and allied fungi during last years, the total number of known lichens and allied fungi species still remains the lowest comparing to the other Baltic States (Motiejūnaitė 2017; Randlane et al. 2019). Current situation indicates the need to include lichens in the monitoring systems of the general species diversity in forest ecosystems, as the knowledge on this group of organisms is very limited in case of Latvia. Also, systematical revisions of lichen herbaria material should be performed using up-to-date literature and methods. Such great amount of work implies the co-operations with specialists from other countries who are familiar with particular groups of lichens and allied fungi, and can contribute in preparation of the new generation of Latvian lichenologists.

4.2. Lichens in post-harvest sites

Various stand- and substrate scale factors, such as volume of dead wood, the level of green tree retention and substrate diversity have an impact on wood-dwelling lichen assemblages, species richness and composition in cut-over sites (e.g. Blasy & Ellis 2014; Ranius et al. 2014; Hämäläinen et al. 2015; Hiron et al. 2017). However, there is a dearth of studies devoted to exploration of forest management activities that shape the richness of wood-dwelling lichens and their assemblages on different substrates in dry boreal post-harvest sites of hemiboreal region.

Altogether 54 species of lichens and non-lichenized fungi were found in dry boreal cut-over sites (I) up to 11 yrs. after cutting. The absolute majority of the species found in studied sites are common lichens of coniferous forests in hemiboreal region. Similar prevalence of forest generalist species on dead wood structures was also observed in studies done by Daniēls

(1983) in nemoral vegetation zone of Netherlands, Fałtynowicz (1986) in dry Scots pine sites in nemoral and hemiboreal vegetation zone of Poland, Spribille et al. (2008) in wide geographical scale with large areas of nemoral vegetation zones in Pacific North-West (North America) to arctic-boreal vegetation in Fennoscandia, Hämäläinen et al. (2015) in dry Scots pine sites in mid-boreal vegetation zone of Finland, Hiron et al. (2017) in conifer dominated mid-boreal vegetation zone of Sweden. Moreover, about two-third of the recorded species in study (I) are the lichens that may also be found on other substrates and habitats apart from pine wood and bark and from dry hemiboreal forests. For example, despite five species turned to inhabit exclusively wood on studied dead-wood structures (e.g. *Calicium glaucellum*, *C. trabinellum*, *Cladonia parasitica*, *Cladonia stellaris* and *Mycocalicium subtile*). Of these, *C. stellaris* is known as a common epigeic species infrequently occurring also on well decayed wood and bark, same as other species of the subgenus *Cladina* (Motiejūnaitė 2002). However, above mentioned calicioid species are known to be obligate epixyles in neighboring Lithuania and Estonia, as well as in Fennoscandia (Lõhmus & Lõhmus 2001; Appendix 1. in Spribille et al. 2008). These two calicioid species are more associated with late successional forest ecosystems, and *C. trabinellum* is considered as indicator species of lichen biocenters in neighboring Lithuania (Motiejūnaitė et al. 2004). *Cladonia parasitica* is Red-listed species in Latvia (similar as in neighboring Estonia and Lithuania) and is commonly associated with lying dead wood in natural and near natural pine forest stands. Surprisingly, in this study, *C. parasitica* was found on horizontal cut surface of stumps, but exclusively in older (9-11 year post-harvest) sites, both in high and low GTR level.

4.2.1. Lichen assemblages on stumps and importance of stumps as lichens habitat

Stumps are known to be a one of the lichen-rich dead wood structures, hosting up to a half of dead wood lichen biota, in natural and managed boreal forest ecosystems (Fałtynowicz 1986; Spribille et al. 2008; Hämäläinen et al. 2015; Svenson et al. 2016; Hiron et al. 2017; Ranius et al. 2019). In all above mentioned studies different sampling methods were used, which makes it difficult to compare lichen species richness on stumps among the studies. For example, in study performed by Fałtynowicz (1986) in Poland, author reports 39 species of lichens recorded on stump surface, while Hämäläinen et al. (2015) found 95 lichen species on same tree species (pine) stumps in Finland. Hiron et al. (2017) found 32 lichen species on pine, spruce and birch stumps in boreal forests of Sweden. The study by Spribille et al. (2008) focused on the obligate lignicolous lichen species in a large-scale geographic area and authors mention that among studied dead wood types, stumps are the structures which host highest

number of obligate lignicolous lichens - 53. Blassy and Ellis (2014) reported 46 species of lichens on pine stumps in north-east Scotland. In case of Latvia (study I) 48 stump-dwelling lichen species were found. However, despite the difference in sampling and geographical location of study sites, in all above mentioned studies, and in the case of Latvia, authors conclude that stumps are among most lichen species rich structures in boreal forests, both, young and old, ranging total species richness on pine stumps at least between 39-48 species.

The results of study (I), similar to the study performed by Hämäläinen et al. (2015), Ranius et al. (2019) in Fennoscandia, and Fałtynowicz (1986) in Poland, showed that pine stumps are lichen species rich substrate even in early-successional stages. Fałtynowicz (1986) found that the vast majority of lichen species on early successional stumps (e.g. 1-5 year after felling) are epiphytes growing on remaining pine bark, which was not yet detached from wood. In addition, Fałtynowicz (1986) also suggested that colonization of cut surface of stumps by lichens is intensive 6-10 years after the logging, and the peak of lichen diversity falls on 11-16 yr. old stumps. Also, the studies of Fałtynowicz (1986) and Humprey et al. (2002) in European temperate forests, and Kruys et al. (1999) and Caruso & Rudolphi (2009) in Fennoscandian boreal forests showed that lichen diversity on conifer dead wood tends to peak at intermediate and late decay classes. The findings of study (I) supports the conclusions of above mentioned studies, showing that species richness of horizontal surface (cut surface) is almost two times higher in 9-11 years old post-harvest sites than in 4-6 years old sites. Also in study (I) it was demonstrated that early-successional (4-6 yr. old) stumps can host up to 18 species of lichens and allied fungi, and most of these are epiphytic species found on the remaining bark. In next five years of succession (9-11 yr. old stumps) species richness can almost double comparing to 4-6 yr. old stumps, and most of species are wood-dwelling, as the bark has already detached, and the variability of decay classes per stump is high. This indicates relatively rapid colonization speed of lichens on the particular substrate within the decade and makes stumps valuable component for lichen diversity conservation in early-successional post-harvest sites.

The results of study (I) for the first time demonstrated clear difference of lichen assemblages on vertical and horizontal surfaces of stumps, showing clear evidence of separate vertical and horizontal assemblages. So far only one study has explored impact of post-harvest age and green tree retention practice on wood-dwelling lichen assemblages in dry boreal pine cut-over sites (Hämäläinen et al. 2015). Comparing the study performed by Hämäläinen et al. (2015), in study (I) lower number of stump-dwelling species was recorded. Such difference in species richness can be explained by distinction of geographical and climatic conditions, as

well as lichen species diversity and composition differences in boreal and hemiboreal vegetation zones. For example, recently such regional differences in lichen assemblages on different wood legacies were demonstrated by Lõhmus et al. (2018), in post-fire sites.

As it was expected, assemblages on younger (4-6 yr. old) and older (9-11 yr. old) stumps significantly differ; this can be explained by predominance of bark-dwelling species on young stumps, and colonization of obligate and facultative epixylic species on older stumps caused by exposure of wood substrate on stumps, similar as it was noticed by Fałtynowicz (1986). Moreover, such ability for rapid changes in lichen assemblages could be explained by evolutionary adaptation of wood-dwelling species to natural disturbances (e.g. windbreaks) which could be similar to logging process, e.g. evolutionary species pool concept suggests that most species are adapted to the naturally most abundant habitats (Taylor et al. 1990). However, the recent study of Ranius et al. (2019) showed that the evolutionary species pool concept did not explain patterns of species' occurrences (including lichens) on different wood substrates in Sweden. However, more empirical studies are needed for each habitat to understand how natural disturbances should be emulated to promote biodiversity in post-harvest sites.

The results of study (I) showed that the cover of bryophytes on stumps increases along the time, shaping the assemblages of lichens. Such pattern potentially can lead to the reduction of available dead wood area on stumps in second decade cut-over sites. This problem potentially could be solved by increasing mean height of stumps, to reduce the speed of bryophyte colonization and increase the available area of dead wood (functional area, not covered by bryophytes) on stumps, in logging sites, similar improvements in boreal forest management were also proposed by Blasy & Ellis (2014).

Stump harvesting, commonly practiced in Fennoscandia (Sweden and Finland), is shown to have a negative impact on wood-dwelling lichen diversity in cut-over sites (Hämäläinen et al. 2015; Persson & Egnell 2018). So far, in Latvia, stump harvesting is not much practiced, and the effects of stump extraction in cut over-sites are still being under study (Lazdiņš et al. 2009; Lībiete et al. 2019). Even more, as study (I) indicated that where there is notable low number and volume of logs and snags in the study sites, stumps may appear to be the main source of dead-wood for lichens during first post-logging decades in cut-over sites of Latvia. For further studies, it is suggested to make a focus dead wood-dwelling lichen assemblages in 30-40 yr old post-harvest sites, as the first forest management activities, such as thinning, are usually performed in this period. At about the same time, dead-wood structures

(including stumps) saved during the harvest, most probably are totally decorticated or overgrown by bryophytes/vascular plants. In such case, mid-aged sites can have even lower amount of available dead wood substrates for lichens, than first two decade regenerated sites.

4.2.2. Lichen functional traits on stumps

Studies that explore functional diversity (e.g. variability of microclimatic conditions in forest structure, species functional trait variability etc.) in young/managed forests are rather recent. Bäcklund (2016) in Sweden, Benítez et al. (2018) in tropical montane forests in southern Ecuador, and Malíček et al. (2019) in Czech Republic showed that analysis of lichen functional traits can contribute to the deeper understanding of the mechanistic processes that control species distribution in managed forest ecosystems. However, the studies that explore the changes in the composition of lichen functional trait groups in regenerated dry boreal forests still lack. In study performed by Bäcklund (2016), non-native phorophytes were also in focus, which makes result comparison problematic. According to Friedl and Büdel (2008) chlorolichens are the most common among lichens. The results of study (I) also showed, that vast majority of the recorded stump-dwelling lichens were chlorolichens or non-lichenized fungi. The predominance of chlorolichens in young cut-over sites is supported by findings of Stofer et al. (2006) in main biogeographic regions of Europe (including boreal Finland) who showed that the proportion of chlorolichens increases with intensification of land use. However, mature and old-growth forests should be analyzed to get deeper understanding about the effect of logging activities on lichen photobiont composition changes in stands.

Stump dataset of the study (I) contained more species with vegetative and mixed reproduction strategy, than species with predominant sexual reproduction. Higher proportions of asexual species in managed rather than in primary forests was previously demonstrated in China and Czech Republic (Li et al. 2013; Malíček et al. 2019). At the same time, the results of other studies performed by Stofer et al. (2006) in main biogeographic regions of Europe (including boreal Finland), and Lundström et al. (2013) in Sweden show an opposite pattern. Meanwhile Bässler et al. (2016) found that asexual reproductive mode facilitate establishment under low temperature conditions. However, all mentioned studies were performed in different forest ecosystem types, outside hemiboreal region. Such pattern differences could be also explained by different study designs and differences in study objects (e.g. stand scale and structure scale variables).

The fungal partner in lichens (mycobiont) reproduces sexually with spores same as other fungi. Most lichen ascospores are small, approximately 1–30 µm, which could be related to the substrate colonization strategy and dispersal over great distances (Hansson et al. 1992). Large ascospores, similar to vegetative propagules are supposed to disperse over a shorter distance, which makes large ascospores important for population turnover at a site-level dispersion, rather than to dispersal of the species over a longer distance (Hansson et al. 1992). The ascospore anatomical and physical characteristics like spore size, colour and cell number etc. - the traits, which can affect the dispersion ability of forest-dwelling lichens, especially in such extreme conditions like post-harvest sites. Based on dataset of study (I) species which inhabit stumps (both young and old ones) in most cases produce small or medium size, hyaline, one celled spores without ornamentation. Species with large spores are extremely uncommon on studied stumps, which does not concur with findings by Hansson et al. (1992). Nevertheless, similar pattern - prevalence of one-celled medium size spore species on dead wood substrates – was found by Pentecost (1981) for lichens distributed in Britain, indicating the ecological importance of this spore trait, for the colonization of short-living substrates (e.g. dead wood). Only several species found in post-harvest sites (both young and old) in study (I), (calicioids and species from genus *Physcia*) have dark spores, and none of species have such combinations of traits as multi-septate dark spores, but older (9-11 yr.) stumps can host more species with unique spore and fruit body traits (mostly calicioids), such as stalked apothecia containing dark ornamented spores. Several reproduction traits were not represented among studied species at all (e.g. perithecia or pycnidia). Also, species with multi-septate (>two-celled) spores were not found. The prevalence of small hyaline spores, probably may explain the prevalence of common species in dataset, species which tend to disperse in large distances. However, most of stump-dwelling lichens colonized stump bark before the harvest, which makes the interpretation of spore trait results problematic.

The colour of the upper surface of lichen thallus is mostly determined by the type of secondary lichen compounds and photobiont type and/or by pigments (such as melanin) of fungal cell walls of cortical hyphae. (Butler & Day 1998; Brodo et al. 2001; Elix & Stocker-Wörgötter 2008). Many secondary lichen compounds and pigments can function as UV-filters (Nguyen et al. 2013). The studies done by Kershaw (1975) indicate that thallus colour appears to modify the temperature of the thallus. Moreover, it was found that light thallus forms are more represented in habitats with high light intensity in lower forest layers (Robinson et al. 1989), while in forest canopy lichen assemblages, dark thallus forms (such as *Bryoria*), which

contain melanin, can be more common than usnic acid containing species – e.g. *Usnea* and *Alectoria* (Färber et al. 2014). The results from stump-dwelling lichens show that almost $\frac{3}{4}$ of recorded species in the studied cut-over sites contain UV-protective substances. Natural dry boreal forests are known to be one of the most sun exposed forest types in hemiboreal region. However, the light intensity in young regenerated forest sites, especially in the first decade, is supposed to be even higher than in mid-aged or old forest ecosystems, except for the cases when natural disturbances such as windbreaks and/or forest fire has affected the ecosystem. Robinson et al. (1989) suggested that the increasing proportion of light-coloured lichens along a light intensity gradient is an expected pattern. The stump dataset analysis of study (I), supports the hypothesis, that the pale-coloured lichen thalli traits are more common in well-lit conditions, than the dark ones. Moreover, it is obvious that most of bark dwelling species found on stumps in study (I) are the species which survived after the unnatural disturbance (forest harvest). Such finding allows to suggest that presence or absence of the UV-protective substances, and color of thalli (which in many cases are defined by lichen substances), are expected to be one of the key traits associated with the survival/extinction of species populations under disturbance conditions, at least in a short-term perspective.

We found that on 9-11 yr. old stumps the proportion of crustose and fruticose thalli (the latter mainly represented by species of genus *Cladonia*), was higher than of species with foliose and squamulose thalli, and it doubled on older stumps compared to the younger (4-6 yr old stumps) ones. Similar prevalence of crustose and/or *Cladonia*-type lichens on conifer dead wood in boreal sites are reported by Faltynovicz (1986); Caruso et al. (2010); Hämäläinen et al. (2015; 2021). Caruso et al. (2010), showed also the rapid increase in occupancy of conifer stumps by crustose species during the first decade. Moreover, stumps in older successional stage were also in focus in studies performed by Faltynovicz (1986) and Caruso et al. (2010), and the replacement of crustose species by species with other thallus traits was observed in both cases. Also, in both of above mentioned cases, authors conclude, that in the end of the second decade after harvest, most of stump wood surface was covered by bryophytes, and was not functioning as a full-scale substrate for lichens.

4.2.3. Fine woody debris

Nowadays, harvesting of stump and fine woody debris for biofuel production is well practiced in Fennoscandia. However, in Baltic countries the harvest of FWD is much more common than stump extraction. Moreover, Baltic States are among largest wood pellet

producers in Europe, and FWD is used as one of the main raw materials for wood pellets production. The potential impact of FWD (including slash) harvest on lichens richness have been assessed to be modest in mid-boreal spruce (Caruso et al. (2008); Svensson et al. (2013)) and pine dominating sites (Hämäläinen et al. (2015)). Hämäläinen et al. (2015) found 69 lichen species on FWD in post-harvest pine sites, while in study (I) 43 species (e.g. approx. 90% of species found on stumps); in both studies most of the recorded lichens were common epiphytes of hemiboreal and boreal forests. Similar to stumps, lichen assemblages on FWD changed and became more distinct in time. This result cannot be considered obvious, as according to Caruso et al. (2008) and Hämäläinen et al. (2015) downed FWD is a less specific lichen habitat and hosts mainly generalist species that are tolerant to environmental change and (or) have good establishment abilities. In case of study (I) five species showed affinity for FWD, of these *Platismatia glauca* and *Bryoria fuscescens* which commonly are associated with lichen communities on tree trunks and twigs in boreal forests, and *Sarea resiniae* which is usually found on resin of *Picea abies*. However, FWD fraction consist of both pre- and post-harvest legacies and thus transition from corticolous to lignicolous communities can take place already on early post-harvest sites. Hiron et al. (2017) estimated the impact at the landscape scale and found that beetles and fungi were more affected by stump extraction, while lichen abundance was more affected by slash extraction. Thus, the large scale study focused on FWD extraction effect on lichen diversity in a landscape-scale, are needed in Latvia as well as in other Baltic countries.

4.2.4. Forest management and lichens

Even though sample size for stumps in our study was considerably lower, tsimilar results as in Hämäläinen et al. (2015) were obtained, i.e. higher retention level favors higher species richness on cut stumps. That can be explained by an additional pool of epiphytic lichens dwelling on retention trees (Hämäläinen et al. (2015)). Altogether the result of study I implies that after the first decade following the harvest, the impact of green tree retention level on lichen species diversity on woody substrata has decreased. However, taking into account the sample size of study (I), the larger-scale study is necessary to test this hypothesis.

The results of the study I showed that dead wood structures in young cut-over sites can host notable lichen and allied fungi species richness during the first decade after the logging. However, the deficiency of dead wood legacies (logs and snags) in studied post-harvest sites was observed. Though snags were considerably lower in numbers, the species richness

recorded on them was notably high and total frequency of some species was even higher than on stumps (e.g. for calicioids *Calicium glaucellum* and *C. trabinellum*). This indicates the importance of snags as valuable habitat for epixylic lichens on harvested sites (see also Runnel et al. 2013; Santaniello et al. 2017), as stumps probably cannot replace spectrum of habitats provided by snags (for example: kelo wood, dry wood above 2 m from ground, etc.), despite they have vertical wood surfaces similar to snags (Hämäläinen et al. 2021). The scarcity of logs and particularly of snags in post-harvest pine forest sites can be explained by local traditions of wood use as a source of firewood for house heating and illegal collection of firewood in recently cut sites (such local tradition of “cleaning” of cutovers have been reported also in Estonia, see Lõhmus et al. 2013). Also, disregard of local laws by logging companies is a possible cause, along with inadequate inspections by responsible environmental institutions.

Given the small sample size of tested treatment groups in the current study, general management recommendations cannot be reliably proposed and the landscape-scale impact of FWD and stump harvest on lichen biodiversity should be studied in future. Still, our study indicates clearly that harvest of pine stumps can increase negative effect to the lichen richness, as stump “life-time” is much longer than of FWD and pine stumps can be inhabited by species of conservation concern, such as *Cladonia parasitica*. According to the results of study (I), the level of green tree retention is supposed to have an effect on lichen species richness on stumps. Also, relatively high lichen species diversity on FWD show that the effect of FWD harvesting should be estimated more precisely in following studies. However, in this study the sample size was relatively low, thus the effect of green tree retention level on dead wood dwelling organisms should be studied in more detail, both in sites with group and dispersed green tree retention.

To follow sustainable forestry certification standards (e.g. FSC, PEFC etc.) forest management companies allowed to retain “high stumps” (live trees cut at height 2–4 m from the ground). Such structures have relatively high value for saproxylic beetles (Abrahamsson & Lindbladh 2006). In studies done by Hämäläinen et al. 2021, authors demonstrate that creation of 3-5 m high “stumps” in cut-over sites has limited value for lichen diversity in c.a. first 20 years. However, authors also mention, that long-term effect of such structure creation should be studied yet. Thus, the true long-term value of high stumps for lichens in post-harvest sites should be explored in more detail, making accents on experimental studies with stump height, both in harvested sites, and in mid-aged production forests as well. The effect on lichens of the measures proposed by Blasy & Ellis (2014) – the retention of stumps with different height has

not yet been studied. Therefore, the retention of stumps with height of about 1 – 1.5 m along with “normal height” stumps can be studied, as an alternative to “high stumps”. Such activities could be combined with group green tree retention, which would reduce the mechanistic problems related to further forest management activities (e.g. planning the technological corridors for forest technique, preparing soil for planting of saplings etc.). The long-term effect of dead wood preservation, on lichen diversity during mid-aged boreal forest management should be studied more detaily, as the number of such studies is very limited. Sustainable forest management planning needs gathering and analysis of more detailed empirical data on biodiversity and its components for modeling the most appropriate management activities (for example species composition and assemblage changes in specific structures, forest types, and forest successional stages, microhabitat quantity, and quality in forest sites, population structure of species etc.). The importance of man-made vertical deadwood (for example “high stumps”, and stumps up to 1.5 m), as well as natural downed structures (e.g. logs) for lichen diversity in boreal cut-over sites, especially in the longer time span post harvest has yet to be studied.

SUMMARY AND CONCLUSIONS

- The findings of this thesis showed that the general knowledge on biota of lichens and allied fungi in Latvia was limited, as extensive contributions done during the last years brought knowledge on 116 taxa, new for Latvian biota. In addition, the revision of genus *Cetrelia* showed the importance of regular revisions of lichen herbaria material using adequate methods of species determination. New records of previously unknown species of lichens and allied fungi in Latvia are expected in near future, as the total number of species in the country still remains the lowest among neighbouring as well as Northern European countries, and large number of lichen groups still need to be revised. Thus, the integration of lichens as organism group in Latvian biodiversity monitoring programs/systems is critically important, as various of newly recorded species are rare and threatened in other Baltic countries or/and in Northern Europe. This brings additional doubts about real knowledge of actual state of lichen biodiversity and possible threats in Latvia.
- In our study it was found that first decade post-harvest forest lichen biota can be relatively diverse, even under the condition that the performed study focused only on dead wood of only one tree species (Scots pine) in particular conditions (*Vacciniosa* type forests) and the number of sampled sites was not high. The higher level (30-40 m³/ha) of green tree retention as a conservation measure in post-cut sites showed the positive effect on stump-dwelling lichen species richness in first years after disturbance (i.e. logging). Still, the overall low amount of natural dead-wood structures (e.g. logs and snags) observed in most harvested study sites, may reduce the positive impact of green tree retention. In addition, low amount of FWD, and at the same time relatively high lichen species diversity show that the effect of FWD harvesting should be estimated more precisely in following studies. Stumps were among the most lichen species rich substrate in our study. It was found that lichen species richness on stumps can almost double in ca 5-year period. Stump dwelling lichen assemblages were distinct among young (4-5 yrs) and old (9-11 yrs) stumps, and also distinct between same age sites with different level of green tree retention. This allows to suggest that level of green tree retention can affect the formation of lichen assemblages in post-harvest sites.

- It was found that lichen assemblages on young stumps formed by lichens with different thallus types, change along the time becoming more represented by crustose and fruticose species on old stumps. Also, both young and old stump lichen assemblages are represented by species which spread by vegetative propagules, or both vegetative propagules and ascospores. Among lichenized species the dataset included only chlorolichens. Moreover, at least $\frac{3}{4}$ of recorded stump-dwelling species in the studied cut-over sites contain acetone-soluble UV-protective substances, and most of the recorded species have light thallus colour, which can be explained as the selection of particular traits by rapidly changing environmental conditions in post-cut forests. The long-term studies are purposed for the estimation of a real effect of forest management on lichen diversity.

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Appendix 1. Summary list of newly recorded lichen and allied fungi taxa for Latvia found and reported between 2015-2020. List of substrata and their abbreviations: Cor = corticolous/epiphytic; Lig = lignicolous/epixylic, Resin = resinicolous; Root = roots of windthrows; Sax = saxicolous/epilithic; SS = sandstone; Ter = terricolous/epigeic. Species found and/or determined by the author of thesis are in **bold**.

| Lichenized fungi | Substrate | Paper No. |
|---|------------------|------------------|
| 1. <i>Anisomeridium polypori</i> (M.B. Ellis & Everh.) M.E.Barr | Cor | II |
| 2. <i>Athallia alnetorum</i> (Giralt, Nimis & Poelt) Arup, Frödén & Søchting | Cor; Sax | II |
| 3. <i>Athallia cerinella</i> (Nyl.) Arup, Frödén & Søchting | Cor | II |
| 4. <i>Athallia holocarpa</i> (Hoffm.) Arup, Frödén & Søchting | Lig; Sax | II |
| 5. <i>Athallia pyracea</i> (Ach.) Arup, Frödén & Søchting | Cor | II |
| 6. <i>Bacidia fraxinea</i> Lönnr. | Cor | II |
| 7. <i>Bacidina delicata</i> (Larbal. ex Leight.) V.Wirth & Vězda | Cor | II |
| 8. <i>Biatora meiocarpa</i> (Nyl.) Arnold | Cor | II |
| 9. <i>Calicium parvum</i> Tibell | Cor | III |
| 10. <i>Calicium pinastri</i> Tibell | Cor | V |
| 11. <i>Calicium trabinellum</i> (Ach.) Ach. | Lig | III |
| 12. <i>Caloplaca chlorina</i> (Flot.) Sandst | Sax | II |
| 13. <i>Caloplaca duplicata</i> (Vain.) H.Olivier | Sax | II |
| 14. <i>Caloplaca ulcerosa</i> Coppins & P.James | Cor | II |
| 15. <i>Candelariella efflorescens</i> R.C.Harris & W.R.Buck | Cor | II |
| 16. <i>Carbonicola anthracophila</i> (Nyl.) Bendiksbj & Timdal | Cor; Lig | III |
| 17. <i>Carbonicola myrmecina</i> (Ach.) Bendiksbj & Timdal | Lig | III |
| 18. <i>Cetrelia monachorum</i> (Zahlbr.) W.L. Culb. & C.F. Culb. | Cor | VI |
| 19. <i>Chaenotheca hispidula</i> (Ach.) Zahlbr | Cor | II |
| 20. <i>Chaenotheca laevigata</i> Nád.v. | Lig | III |

| | | |
|--|----------|-----|
| 21. <i>Chaenotheca xyloxena</i> Nádv. | Lig | II |
| 22. <i>Chrysothrix flavovirens</i> Tønsberg | Cor | II |
| 23. <i>Cresporhaphis wienkampii</i> (J. Lahm ex Hazsl.) M.B.Aguirre | Cor | II |
| 24. <i>Dermatocarpon miniatum</i> (L.) W.Mann. | Sax | IV |
| 25. <i>Diplotomma pharcidium</i> (Ach.) M.Choisy | Cor | II |
| 26. <i>Eopyrenula leucoplaca</i> (Wallr.) R.C.Harris | Cor | II |
| 27. <i>Fuscidea arboricola</i> Coppins & Tønsberg | Cor;Lig | II |
| 28. <i>Gallowayella weberi</i> (S.Y. Kondr. & Kärnefelt) S.Y.Kondr., Fedorenko, S.Stenroos, Kärnefelt, Elix, J.- S.Hur & A.Thell | Cor | II |
| 29. <i>Gregorella humida</i> (Kullh.) Lumbsch | Ter | II |
| 30. <i>Gyalecta derivata</i> (Nyl.) H.Olivier | Cor | II |
| 31. <i>Gyalecta flotowii</i> Körb | Cor | II |
| 32. <i>Lecanora farinaria</i> Borrer | Cor | II |
| 33. <i>Lecanora subcarpineae</i> Szatala | Cor | II |
| 34. <i>Lecania croatica</i> (Zahlbr.) Kotlov | Cor | V |
| 35. <i>Leptogium byssinum</i> (Hoffm.) Nyl. | Ter | II |
| 36. <i>Micarea anterior</i> (Nyl.) Hedl. | Lig | II |
| 37. <i>Micarea byssacea</i> (Th.Fr.) Czarnota, Guzow- Krzemińska & Coppins | Lig; Cor | II |
| 38. <i>Micarea micrococca</i> (Körb.) Gams ex Coppins | Cor | II |
| 39. <i>Micarea peliocarpa</i> (Anzi) Coppins & R.Sant. | Lig | II |
| 40. <i>Normandina acroglypta</i> (Norman) Aptroot | Cor | II |
| 41. <i>Ochrolechia androgyna</i> (Hoffm.) Arnold | Cor | II |
| 42. <i>Ochrolechia microstictoides</i> Räsänen | Lig;Cor | II |
| 43. <i>Parmelia serrana</i> A.Crespo, M.C.Molina & D.Hawksw. | Cor | II |
| 44. <i>Parmelia submontana</i> Nádv. ex Hale | Cor | II |
| 45. <i>Peltigera extenuata</i> (Vainio) Lojka | Ter | III |
| 46. <i>Phaeophyscia endophoenicea</i> (Harm.) Moberg | Cor | II |
| 47. <i>Pilophorus cereolus</i> (Ach.) Th. Fr. | Ter/SS | VII |

| | | |
|---|----------------------------|------------------|
| 48. <i>Polyscaulion phlogina</i> (Ach.) Arup, Frödén & Søbchting | Lig | II |
| 49. <i>Protothelenella sphinctrinoidella</i> (Nyl.) H.Mayrhofer & Poelt | Ter | II |
| 50. <i>Psilolechia clavulifera</i> (Nyl.) Coppins | Root | II |
| 51. <i>Pycnora praestabilis</i> (Nyl.) Hafellner | Lig | V |
| 52. <i>Pycnora sorophora</i> (Vain.) Hafellner | Cor | III |
| 53. <i>Rinodina septentrionalis</i> Malme | Cor | II |
| 54. <i>Rinodina sophodes</i> (Ach.) A.Massal. | Cor | II |
| 55. <i>Schismatomma pericleum</i> (Ach.) Branth & Rostr. | Cor | II |
| 56. <i>Scytinium teretiusculum</i> (Wallr.) Otálora, P.M. Jørg. & Wedin | Cor | VII |
| 57. <i>Scoliciosporum sarothamni</i> (Vain.) Vězda | Cor | I |
| 58. <i>Solitaria chrysophthalma</i> (Degel.) Arup, Frödén & Søbchting | Lig | II |
| 59. <i>Steinia geophana</i> (Nyl.) Stein | Ter | II |
| 60. <i>Strigula jamesii</i> (Swinscow) R.C.Harris | Cor | II |
| 61. <i>Thelocarpon superellum</i> Nyl. | Ter | II |
| 62. <i>Trapelia coarctata</i> (Sm.) M.Choisy | Sax | IV |
| 63. <i>Trapelia corticola</i> Coppins & P.James | Lig | II |
| 64. <i>Trapelia placodioides</i> Coppins & P.James | Sax | IV |
| 65. <i>Umbilicaria hirsuta</i> (Sw. ex Westr.) Hoffm. | Sax | IV |
| 66. <i>Verrucaria tectorum</i> (A.Massal.) Körb. | Sax | II |
| 67. <i>Verrucaria xyloxena</i> Norman | Ter | II |
| 68. <i>Xylographa parallela</i> (Ach.) Fr. | Lig | II |
| 69. <i>Zwackhia soreidiifera</i> (P.James) Ertz | Cor | II |
| Non lichenized, lichenicolous taxa | Host | Paper No. |
| 70. <i>Arthonia epiphyscia</i> Nyl. | <i>Physcia aipolia</i> | II |
| 71. <i>Arthonia molendoi</i> (Heufl. ex Frauenf.) R.Sant. | <i>Xanthoria parietina</i> | II |
| 72. <i>Arthrorhaphis aeruginosa</i> R. Sant. & Tønsberg | <i>Cladonia</i> spp. | V |
| 73. <i>Corticifraga fuckelii</i> (Rehm) D. Hawksw. & R.Sant. | <i>Peltigera neckeri</i> | II |

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|---|---|-----|
| 74. <i>Chaenothecopsis epithallina</i> Tibell | <i>Chaenotheca trichialis</i> | V |
| 75. <i>Clypeococcum hypocenomyces</i> D. Hawksw. | <i>Hypocenomyce scalaris</i> | III |
| 76. <i>Didymocyrtis epiphyscia</i> Ertz & Diederich s. lat. | <i>Xanthoria parietina</i> | II |
| 77. <i>Didymocyrtis ramalinae</i> (Roberge ex Desm.) Ertz, Diederich & Hafellner | <i>Ramalina fraxinea</i> | II |
| 78. <i>Ellisembia lichenicola</i> Heuchert & U.Braun | <i>Ramalina fraxinea</i> | II |
| 79. <i>Epicladonia sandstedei</i> (Zopf) D.Hawksw. | <i>Cladonia coniocraea</i> | II |
| 80. <i>Erythrimum aurantiacum</i> (Lasch) D. Hawksw. & A.Henrici | <i>Physcia</i> spp. | II |
| 81. <i>Graphium aphthosae</i> Alstrup & D.Hawksw. | <i>Peltigera neckeri</i> | II |
| 82. <i>Homostegia piggotii</i> (Berk. & Broome) P.Karst. | <i>Parmelia submontana</i> | II |
| 83. <i>Illosporium carneum</i> Fr. | <i>Peltigera extenuata</i> | III |
| 84. <i>Lichenochora obscuroides</i> (Linds.) Triebel & Rambold | <i>Phaeophyscia orbicularis</i> | II |
| 85. <i>Lichenochora weillii</i> (Werner) Hafellner & R.Sant. | <i>Physconia enteroxantha</i> | II |
| 86. <i>Lichenocodium lecanorae</i> (Jaap) D.Hawksw. | <i>Evernia prunastri</i> | II |
| 87. <i>Lichenocodium pyxidatae</i> (Oudem.) Petr. & Syd. | <i>Cladonia</i> aff. <i>chlorophaea</i> | II |
| 88. <i>Lichenocodium usneae</i> (Anzi) D.Hawksw. | <i>Evernia prunastri</i> | II |
| 89. <i>Lichenocodium xanthoriae</i> M.S.Christ. | <i>Xanthoria parietina</i> | II |

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| 90. <i>Lichenodiplis lecanorae</i> (Vouaux) Dyko & D.Hawksw. | <i>Lecanora</i> <i>aff.hagenii</i> | II |
| 91. <i>Marchandiomyces corallinus</i> (Roberge) Diederich & D.Hawksw. | <i>Physcia tenella</i> | II |
| 92. <i>Phaeopyxis punctum</i> (A. Massal.) Rambold, Triebel & Coppins | <i>Cladonia</i> <i>digitata</i> | II |
| 93. <i>Pyrenochaeta xanthoriae</i> Diederich | <i>Xanthoria</i> <i>parietina</i> | II |
| 94. <i>Reconditella physconiarum</i> Hafellner & Matzer | <i>Physconia</i> <i>distorta</i> | II |
| 95. <i>Refractohilum peltigerae</i> (Keissl.) D.Hawksw. | <i>Peltigera</i> spp. | II |
| 96. <i>Sphinctrina turbinata</i> (Pers.: Fr.) De Not. | <i>Pertusaria</i> <i>pertusa</i> | II |
| 97. <i>Stigmidium microspilum</i> (Körb.) D.Hawksw. | <i>Graphis scripta</i> | II |
| 98. <i>Taeniolella punctata</i> M.S. Christ. & D.Hawksw. | <i>Graphis scripta</i> | II |
| 99. <i>Teloggalla olivieri</i> (Vouaux) Nik.Hoffm. & Hafellner | <i>Xanthoria</i> <i>parietina</i> | II |
| 100. <i>Thelocarpon epibolum</i> Nyl. var. <i>epibolum</i> | <i>Peltigera</i> <i>neckeri</i> | II |
| 101. <i>Tremella lichenicola</i> Diederich | <i>Violella fucata</i> | II |
| 102. <i>Tremella phaeophysciae</i> Diederich & M.S.Christ. | <i>Phaeophyscia</i> <i>orbicularis</i> | II |
| 103. <i>Trichonectria anisospora</i> (Lowen) van den Boom & Diederich | <i>Hypogymnia</i> <i>physodes</i> | II |
| 104. <i>Trichonectria rubefaciens</i> (Ellis & Everh.) Diederich & Schroers | <i>Parmelia</i> <i>sulcata</i> | II |
| 105. <i>Vouauxiomyces santessonii</i> D.Hawksw. | <i>Platismatia</i> <i>glauca</i> | II |
| 106. <i>Xenonectriella leptaleae</i> (J.Steiner) Rossman & Lowen | <i>Physcia aipolia</i> | II |
| Non lichenized, saprobic or resinicolous taxa | Substrate | Paper No. |
| 107. <i>Chaenothecopsis savonica</i> (Räsänen) Tibell | Lig | III |
| 108. <i>Chaenothecopsis pusiola</i> (Ach.) Vain | Lig | VIII |

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| 109. <i>Chaenothecopsis viridireagens</i> (Nádv.) A.F.W.Schmidt | Lig | II |
| 110. <i>Leptorhaphis epidermidis</i> (Ach.) Th.Fr. | Cor | II |
| 111. <i>Microcalicium ahlneri</i> Tibell | Lig | VIII |
| 112. <i>Microcalicium arenarium</i> (Hampe ex A. Massal.) Tibell | Root | III |
| 113. <i>Peridiothelia fuliguncta</i> (Norman) D.Hawksw. | Cor | II |
| 114. <i>Sarea difformis</i> (Fr.) Fr. | Resin | II |
| 115. <i>Sarea resinae</i> (Fr.: Fr.) Kuntze | Resin | II |
| 116. <i>Stictis brunnescens</i> Gilenstram, Döring & Wedin | Lig | III |

ORIGINAL PAPERS

I



Lichen assemblages on Scots pine stumps and fine woody debris in hemiboreal post-harvest sites: the impact of site age and green tree retention

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With 4 figures and 2 tables

Abstract: Retention of live trees and dead wood structures in clear-cut sites is a common silviculture measure for biodiversity purposes. We studied lichen assemblages on pine stumps and fine woody debris (FWD) in 16 post-cut (4–6 and 9–11 yr. old) dry boreal pine stands in Latvia to explore what type of substrata and stand-scale characters (e.g., retention level, time since harvest) are related to lichen species richness and differences in composition. We found 48 lichen species on stumps and 43 species on FWD. Majority of the species (except *Cladonia parasitica*) were common lichens of coniferous forests in hemiboreal regions. Time since harvest and retention level had positive impacts on richness on stumps, but not on FWD. Increase in total species richness on stumps in older post-harvest sites compared to the younger ones was strong and relatively rapid. Notwithstanding species richness, assemblages on FWD and on stumps were distinct between older and younger cut sites. The impact of time also emerged when assemblages on vertical and horizontal stump surfaces were separated. We conclude that pine stumps are important to lichen richness and post-harvest recovery of the epixylic lichen biota, especially in the face of alarming scarcity of snags and logs in cut-over sites in Latvia, where dead wood legacies (particularly snags) are not retained in sufficient amount.

Key words: coarse woody debris; epixylic; species diversity; logging; Latvia

Introduction

Young and mid-age structurally simplified forest stands have become more frequent in modern forest landscapes of Northern Europe (UNECE & FAO 2011, Vilén et al. 2012). In particular, the availability of dead wood (e.g., standing and downed coarse and fine

woody debris) is reduced by logging and it is considered that in European production forests the current volumes of dead wood are less than 10% of what is found in natural forests (Stokland et al. 2012). Meanwhile species dependent on dead wood constitute 20–25% of all forest-dwelling species, as have been estimated for Fennoscandia (Siitonen 2001). To reduce negative impacts of logging on biodiversity loss and to ensure sustainable forest management policies, different management practices have been applied. Retention of live trees (i.e. green tree retention) and dead wood structures in cut sites is one of the most frequently applied practices in northern Europe (Gustafsson et al. 2010, Gustafsson et al. 2012). Recent studies have shown that retention of woody legacies (fine woody debris, snags and logs) can provide suitable microhabitats for species of various organism groups, like saproxylic beetles and polypores (Juutilainen et al. 2014, Floren et al. 2015), as well as for lichens (Löhmus & Löhmus 2010, Lundström et al. 2013, Runnel et al. 2013, Ranius et al. 2014, Hämäläinen et al. 2014).

Lichens are considered a valuable component of forest ecosystems, with important roles in nutrient turnover, water uptake and retention, as food and habitat for various organisms, etc. (Ellis 2012, Asplund & Wardle 2017). A large proportion of forest-dwelling lichens in boreal forests use dead wood as the main or facultative substratum (Spribille et al. 2008). The dynamic process of substratum decay provides suitable microhabitat conditions for different lichen assemblages over the course of time (Fąfryniewicz 1986, Löhmus & Löhmus 2001). For example, many wood-dependent lichens in boreal Scots pine (*Pinus sylvestris* L.; hereafter “pine”) forests are confined to old hard wood, such as snags, low natural stumps and logs (Santaniello et al. 2017). Norway spruce (*Picea abies* (L.) Karst.; hereafter “spruce”) and pine stumps in cut-over sites can support rich epixylic lichen assemblages (Caruso et al. 2008, Svensson et al. 2013, Hämäläinen et al. 2015), thus providing “buffering” substrata to maintain populations of dead wood lichens at the landscape scale. In managed boreal spruce forest, lichen species richness is higher on fine woody debris (FWD) when equal volumes of coarse woody debris (CWD) and FWD are compared (Kruys & Jonsson 1999). Attached dead branches make a significant proportion of all available dead wood in boreal forests, nevertheless such types of substrata mainly support generalist lichen species (Svensson et al. 2014). Downed fine woody debris (including slash) also can provide suitable microhabitats for a relatively large biomass of dead wood dwelling lichens (Kruys & Jonsson 1999, Svensson et al. 2014, Hämäläinen et al. 2015); however, several studies have shown that only few species are present on such substrata (Caruso et al. 2008, Svensson et al. 2016).

It can be expected that dead wood legacies may provide species rich epixylic lichen assemblages during the first post-harvest decade both in boreal and hemiboreal regions of North-Europe (Caruso & Rudolphi 2009, Runnel et al. 2013). However, the difference in wood dwelling lichen assemblages and/or species composition between stands in boreal and hemiboreal regions may differ (e.g., for burned forests see Löhmus et al. 2018). Unfortunately, in the hemiboreal region, lichen assemblages on dead wood have been representatively studied only in Estonia (e.g., Löhmus & Löhmus 2001, Löhmus & Löhmus 2011, Runnel et al. 2013), and not in Latvia, Nordic countries or Russia, so far.

In Latvia, same as in Estonia, the proportion of forest land makes about half of the terrestrial area of the country and the proportion of coniferous forests in both countries is also similar (ca 50%). Pine is the second most frequent tree species in Latvia and dry boreal pine forests cover ca 6% (i.e. about 205 000 ha) of total forest area of Latvia; about one fourth of these forests are young-regenerated. General regulations of Ministry Cabinet of Latvia (Anonymous 2000) require that at least four largest structures per hectare of dead wood should be left as retention in the clear cut sites. However, in post-harvest sites, local people are traditionally allowed collection of dead wood for house heating. Additionally, during the last decade slash and largest fractions of FWD have been collected to produce wood pellets and briquettes in most logging sites. Neither of these processes are regulated or accounted for in forestry statistics and result in comparatively low amounts of dead wood in logged areas of Latvian forests.

The aim of the current study was to describe lichen assemblages on pine dead wood structures (stumps, fine woody debris, hereafter FWD, snags and logs) in post-cut 4–6 and 9–11 yr. old dry boreal pine stands in Latvia. The study was specifically aimed to explore what type of substrata and stand-scale characters (e.g., decay stage, proportion of wood; green tree retention level, time since harvest, volume of dead wood) are related to the lichen species richness and composition on pine dead wood legacies. We also aimed to find out whether the assemblages on vertical and horizontal stump surfaces are distinct.

Materials and Methods

Study area and design

The study was performed in the SE part of Latvia, in the hemiboreal forest zone (according to Ahti et al. 1968). In the study region, the average air temperature is 17.5 °C in July and –6 °C to –7 °C in February; the annual precipitation is 600–650 mm (Turlais 2011). The study sites were selected from *Vacciniosa* type (Bušs 1976), i.e. pine dominated forest on higher fluvioglacial landforms and till mounds with Podzols (pH 3.5–5.0) having continuous moss cover and species-poor undergrowth (mostly shrubs of *Vaccinium vitis-idaea* L. and *Calluna vulgaris* (L.) Hull.) and very sparse under-storey. The study sites were selected in 2.5–9 ha size retention-cut stands situated outside of protected areas (Fig. 1). Among the study sites, four treatment types were designated, based on combinations of time since harvest (4–6 and 9–11 years) and green tree retention level (“high”, 14–20 trees per ha or 30–40 m³/ha, and “low”, ≤6 trees per ha or ≤12 m³/ha). Each of the four treatment combinations (hereafter as “treatment type”) was represented by four sites, 16 sites in total. Minimal distance between the study sites was 1 km. Sites were selected from comparatively homogeneous, flat relief to minimize variation on humidity conditions among the study sites. The primary site selection was based on Latvia’s State Forests data. Prior to fieldwork selected sites were visited to confirm in the field the data of forest age and site type by the first author. All sites were standardized to 2.5 ha in size, with 30 m wide buffer zones on each side. In all stands pine saplings have been planted,

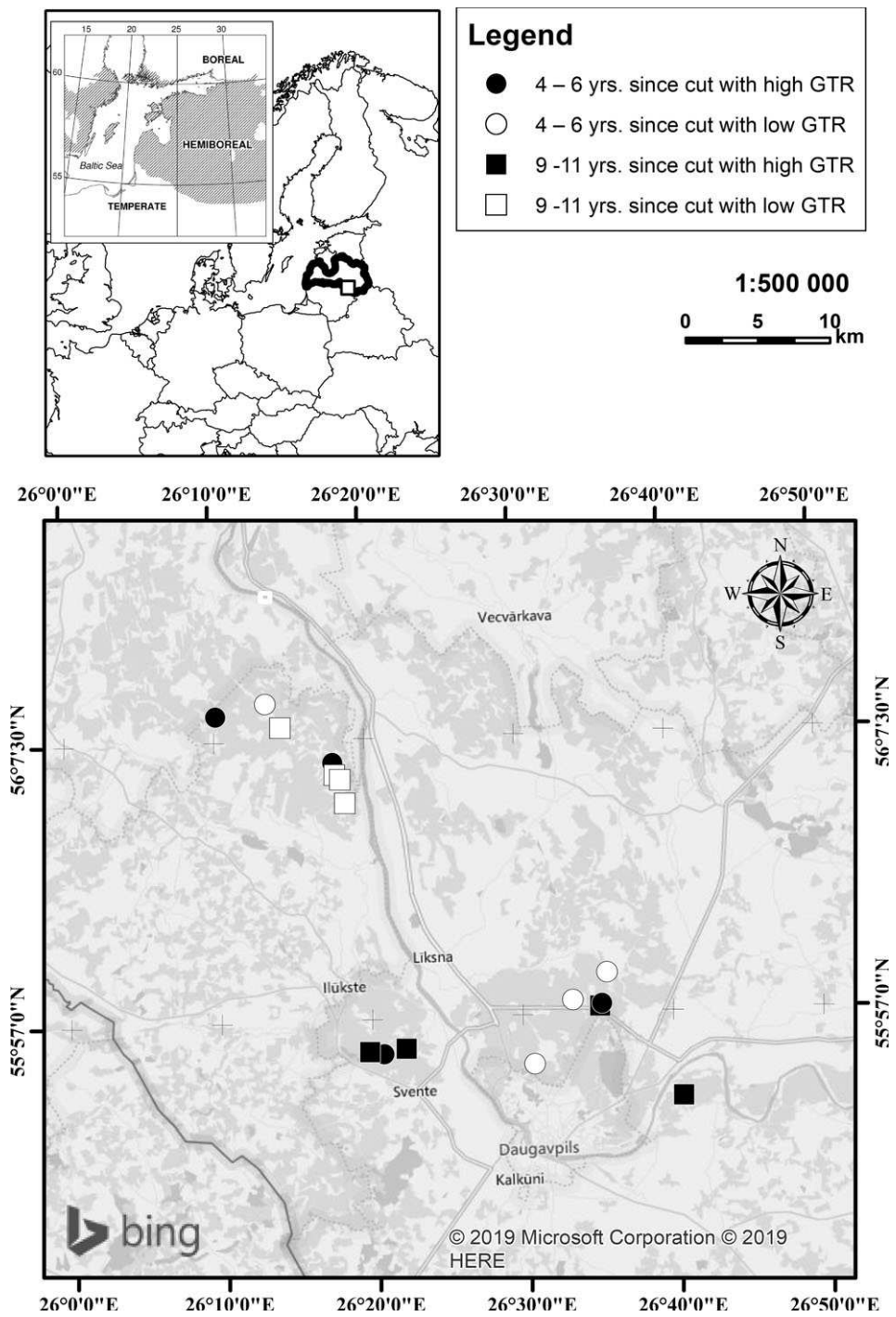


Fig. 1. Location of the study sites in SE part of Latvia. Hemiboreal zone (shaded on the upper left panel) in the Baltic region has been delineated according to Ahti et al. (1968).

and all stands are state-owned. Total lichen sampling from woody structures per site were performed from two FWD plots (2×8 m) and 4–6 stumps selected along sampling transects and from 1–2 logs and 1–2 snags selected arbitrary over the whole study site.

Field work

Field work was performed by the first author from June till September of 2016. In each study site four N–S orientated 50 m long transects (divided to 1 m sections) were arranged in N, S, W and E sides of study site to describe characteristics of pine stumps, pine coarse woody debris (CWD; i.e. logs with diameter ≥ 10 cm) and fine woody debris (FWD; i.e., fallen branches and thin laying trunks of dead wood with diameter 1–9.9 cm). Among FWD, there was occasional ($<1\%$ of all FWD) presumable admixture of other species (spruce and birch). The nearest distance between transect lines was at least 50 m.

The diameter of all FWD and CWD that had diameter ≥ 1 cm at the cross section by transect line were measured and based on collected data, volumes of downed woody debris were calculated based on the method by Van Wagner (1968). To describe lichen assemblages in each study site 1–2 logs were selected arbitrarily among the ones that crossed four transect lines. The length, diameter of log base and top were recorded. In current study most of the sampled logs were in direct contact with soil surface. Over the study site all snags were investigated (only nine snags over all study sites were found). The height of snag and DBH were recorded. For each transect, one stump was selected randomly, so that 4 stumps were surveyed per 2.5 ha study plot area. Stump characteristics, such as height, decay stage of wood at five-point scale (according to Lõhmus & Kraut 2010), and functional habitat area of the stump (i.e. wood and bark not overgrown by bryophytes and vascular plants) were recorded. The decay of vertical and horizontal surfaces was evaluated separately (at four and three different points of the surface, respectively). Data on lichen species occurrence on stumps were recorded separately for vertical and horizontal surfaces, and records on bark and wood were separated as well.

Two FWD sampling plots (2×8 m) were established at each site along transects at the N and S part of the study site. Lichen species growing on FWD were surveyed for 40 minutes per plot. One side of the FWD sampling plot was situated along the 21–29 m of the transect. Lichen species occurrence on arbitrarily selected logs per site was recorded using four 50×20 cm plots (metal frame with grid of 5×5 cm, 40 sheets in total) per log. Plots were placed along the log from the widest end toward the top and with a distance of 20 cm from previous one. First and last plots were at the log's horizontal surface, second on the vertical left side and third on the vertical right side (both about 3–5 cm from the ground). Additional lichen species growing outside log plots were also recorded. On

snags, lichens were sampled using two 50 × 20 cm plots (same frame as for the logs). One of the plots was placed 20 cm from the base and the second plot at the breast height (about 130 cm from the base) of the snag. All plots were positioned at the N side of the snag.

Most lichen species were identified *in situ*. Doubtful specimens were collected and later identified in the laboratory using routine lichenological methods, including spot tests and standardized methods of thin layer chromatography (TLC) (Orange et al. 2001). Species of the genus *Lepraria* Ach., were determined in field using C and K spot tests. For confirmation of species identification 36 samples of *Lepraria* spp. were determined using TLC; in all cases field determinations appeared to be correct. In the case of *Mycocalicium subtile* (Pers.) Szatala, 30 specimens were collected and identified in laboratory for confirmation. In all cases the species was identified correctly in field. All collected data of *Micarea denigrata* s. lat. were treated together in analysis as *Micarea* sp., and juvenile specimens of *Usnea* were treated as *Usnea hirta* (L.) Weber ex F.H. Wigg. The nomenclature of taxa follows Nordin et al. (2011).

Statistical analyses

Difference of functional area of stumps and volume of CWD and FWD across four treatment combinations were tested with a one-way Welch ANOVA with Games-Howell post hoc analysis (before the analysis two outliers were removed to achieve normal distributions for each treatment). Homogeneity of variances was violated, as assessed by Levene's Test of Homogeneity of Variance ($p < 0.001$).

General linear models (GLMs) were used to study the effects of post-cut age and green tree retention level on (1) total and (2) mean lichen species richness on stumps and on fine woody debris (FWD) per site. The number and distribution of logs and snags among the sites was insufficient for statistical analysis. The fixed factors were age of cut sites (4–6 and 9–11 yr. old) and GTR level (≤ 6 trees/ha as “low” and ≥ 14 trees/ha as “high”). Two un-correlated variables were also included per each model: for models (1) and (2) variables average stump functional area and average stump diameter per site, and for model (3) variables volume of FWD and CWD. GLMs were run initially including all factors and interaction between site age and GTR level. Non-significant factors were removed from the model manually one by one, commencing with the least significant factor.

Difference of lichen assemblage composition among treatment types (depending on site age and GTR level) were studied for 1) stumps (full unit), 2) stumps vertical and horizontal surfaces separated, and 3) FWD. A multi-response permutation procedure (MRPP) was used to test if Sørensen (Bray-Curtis) dissimilarities among predefined groups exceed those resulting from random assignment of sample units to those classes, and it has the advantage of not requiring distributional assumptions that are seldom met with ecological assemblage data. To correct the p-values for multiple comparisons, a Bonferroni correction was applied.

In the species matrices, frequency score of the species per site (0–4 for stumps and 0–2 for FWD) were used as input. Results based on the full species matrix are presented (i.e. MRPP analysis gave similar results to full and reduced species matrices). The full species matrix was also used to find species specific to treatment type using indicator species analysis (ISA; Dufrêne & Legendre 1997). In that analysis, we distinguished the species with significant ($p < 0.05$) and relatively high ($> 50\%$) indicator values.

The results of MRPP test were visualised using nonmetric multidimensional scaling (NMS). NMS was run on Sørensen distances and the medium autopilot mode options (Sørensen dissimilarity; two-axes dimension; 200 runs with real data and 100 with randomized data; stability criterion 0.00001). Two-dimensional solutions were acceptable based on mean stress values (< 19.3). Coefficients of determination (r^2) were calculated between distances in the ordination space and dissimilarities in the original space as an expression of variation accounted for by ordination axes. Two continuous factors were added to the analysis to explain assemblage gradients along the NMS axes (for stumps, functional area and average diameter, and for volumes of FWD and CWD per ha). A cut-off value of $r^2 > 0.2$ was used as a criterion for regarding an environmental factor as sufficiently strongly related to an ordination axis to be represented by a vector in the NMD diagram. The MRPP, ISA and NMS analysis were performed with the software package PC-ORD 5.0 (McCune & Mefford 1999).

Results

Altogether 54 species of lichens and allied fungi were recorded from all sampled substrata types from 16 study sites (Appendix 1). Stumps were inhabited by 48 species, of them only one, *Placynthiella dasaea*, was exclusive for this type of substratum. Fine woody debris (FWD) was inhabited by 43 species, including four exclusive species (Appendix 1). On snags ($n = 9$) 27 species were recorded and on logs ($n = 24$) 44 species, adding one new species to the list (*Buellia griseovirens*) (Appendix 1). Most of the species found in the study are common in Latvia, but three were recorded for the first time in the country, namely, *Calicium trabinellum*, *Pycnora sorophora*, and *Scoliciosporum sarothamni*. In addition, one Red-listed lichen in Latvia – *Cladonia parasitica* – was found in seven sites (all 9–11 yr. post-harvest), growing on horizontal surfaces of stumps and logs (Appendix 1). Stump characteristics varied slightly among cut site combinations (Table 1), showing significantly higher horizontal area in 4–6 yr. and 9–11 yr. old sites with high level of GTR (Welch's $F_{(3, 30.838)} = 5.447$, $p = 0.004$) and higher functional area on 4–6 yr. old sites with high level of GTR (Welch's $F_{(3, 30.831)} = 5.012$, $p = 0.006$). Well decayed stumps (classes 3–4) occurred only on old cut sites (Appendix 2). Volumes of FWD and CWD were relatively low and similar among site treatments ($p = 0.454$ and $p = 0.876$, respectively; Table 1).

Based on the final GLM model, the total and mean species richness on stumps depended on cut site age (higher richness in 9–11 yr. sites) and on GTR level (higher richness in sites with “high” GTR level; Table 2, Fig. 2a). The interaction between age and GTR

Table 1. Characteristics of stumps, lying fine and coarse woody debris (FWD and CWD, respectively) in cut sites of different age and green tree retention (GTR) level (high ≥ 14 and low ≤ 6 trees/ha). Within each site combinations 16 stumps were sampled.

| Age GTR | 4–6 yr. | | 9–11 yr. | |
|---|-----------------|------------------|-----------------|-----------------|
| | High | Low | High | Low |
| Stumps | | | | |
| Height (cm) | 18.4 \pm 5.6 | 20.4 \pm 2.3 | 20.3 \pm 3.1 | 21.6 \pm 3.2 |
| Diameter (cm) | 37.1 \pm 8.3 | 29.2 \pm 4.0 | 31.4 \pm 5.8 | 28.8 \pm 7 |
| Horizontal area (dm ²) | 11.3 \pm 4.5 | 6.8 \pm 1.7 | 8.0 \pm 2.8 | 6.9 \pm 3.1 |
| Vertical area (dm ²) | 22.1 \pm 10.5 | 18.7 \pm 3.3 | 20.2 \pm 5.7 | 19.6 \pm 5.4 |
| Functional horizontal area (dm ²) | 9.2 \pm 4.1 | 6.0 \pm 1.4 | 5.4 \pm 2.6 | 5.3 \pm 2.6 |
| Functional vertical area (dm ²) | 16.4 \pm 9.6 | 14.8 \pm 3.4 | 11.7 \pm 5.1 | 13.2 \pm 4.4 |
| Volume (m ³) | 0.02 \pm 0.01 | 0.01 \pm 0.004 | 0.02 \pm 0.01 | 0.02 \pm 0.01 |
| FWD | | | | |
| Volume (m ³ /ha) | 0.77 \pm 0.22 | 0.91 \pm 0.28 | 1.14 \pm 0.43 | 0.96 \pm 0.27 |
| CWD | | | | |
| Volume (m ³ /ha) | 0.36 \pm 0.18 | 0.30 \pm 0.18 | 0.32 \pm 0.17 | 0.32 \pm 0.21 |

Table 2. The final results of general linear model (GLM) analyses for total and mean lichen species richness on stumps depending on the cut site age (Age) and on the level of green tree retention (GTR).

| | Total richness | | | | Mean richness | | | |
|--------------|----------------|------|------|------------------|---------------|------|------|------------------|
| | df | MS | F | p | df | MS | F | p |
| Model | 2 | 521 | 34.7 | <0.001 | 2 | 217 | 77.3 | <0.001 |
| GTR | 1 | 81.0 | 5.40 | 0.037 | 1 | 19.1 | 6.80 | 0.022 |
| Age | 1 | 961 | 64.1 | <0.001 | 1 | 415 | 148 | <0.001 |
| Error | 13 | 15.0 | | | 13 | 2.81 | | |

level, as well as stump-scale variables (average stump functional area and diameter per site) were non-significant. No tested variables had significant impact on FWD total species richness (df = 1, MS = 0.56, F = 0.04, $p > 0.05$; Fig. 2b).

Lichen assemblages on stumps differed between age groups of cuts within the same GTR level (MRPP test; A = 0.2 and 0.4, $p = 0.006$ for both) and between high and low GTR level sites of the same site age (A = 0.2, $p = 0.007$ for both). Such groupings were slightly visible in the NMS ordination (Fig. 3a). Among tested environmental variables average functional area of the stump correlated with the second ordination axis ($r^2 = 0.24$). Indicator species analysis showed that 19 species (seven of them *Cladonia*) associated signifi-

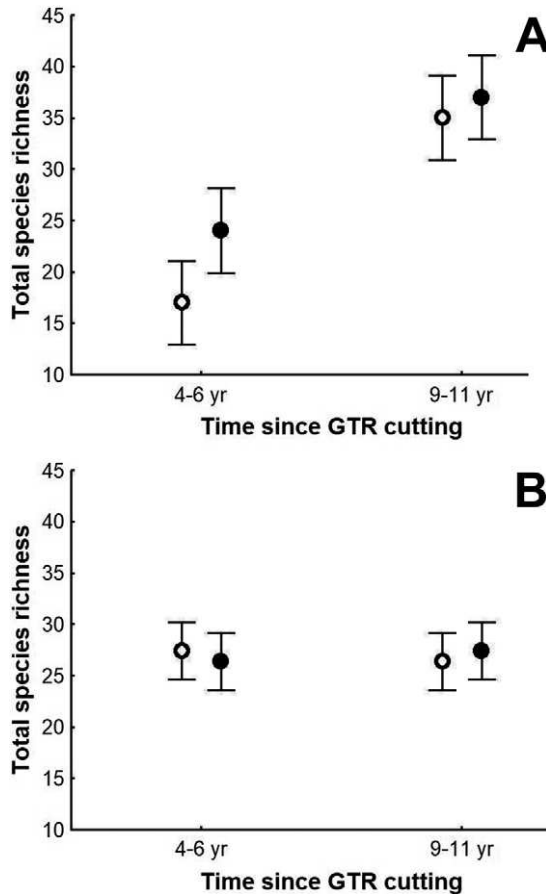


Fig. 2. Total species richness on pine stumps (A) and FWD (B) on 4–6 and 9–11 yr. post-harvest sites with low (≤ 6 trees/ha; open symbol) and high (≥ 14 trees/ha; filled symbol) green tree retention (GTR) level.

cantly with 9–11 yr. cut sites, four of them showing also the preference for GTR level (like *C. parasitica* on sites with lower GTR level) (Appendix 3).

Similar assemblage pattern on age difference emerged also if stump vertical and horizontal surfaces were analyzed separately ($A = 0.2$, $p < 0.001$). In addition, vertical and horizontal surfaces differed from each other within the same age group (MRPP test $A = 0.3$, $p < 0.001$, Fig. 4). A focus on horizontal stump surface only (as the case of post-cut colonization) showed clear difference of assemblages between 4–6 yr. and 9–11 yr. cut sites both on wood and on bark around stump ($A = 0.5$, $p < 0.001$). A focus on assemblages on wood only revealed the difference between horizontal and vertical surfaces of stumps both on 4–6 yr. ($A = 0.4$, $p < 0.001$) and 9–11 yr. ($A = 0.3$, $p < 0.001$) cut sites.

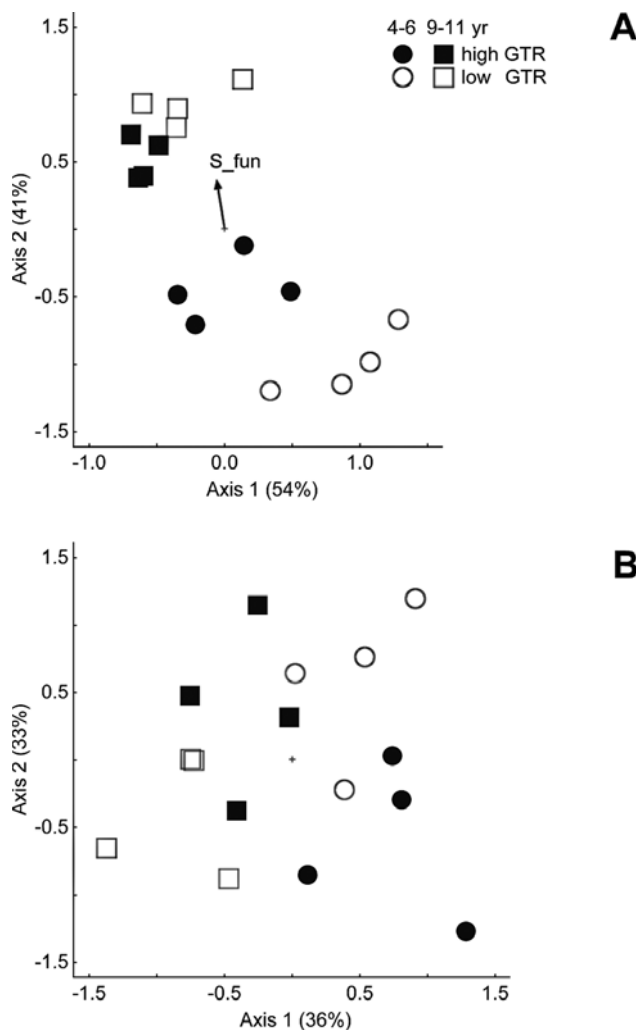


Fig. 3. The composition of lichen species assemblages on pine stumps (A) and FWD (B) depending on years (yr.) since cutting and green tree retention (GTR) level (“high” ≥ 14 trees/ha and “low” ≤ 7 trees/ha). Functional area of stumps (S_{fun}) is related to second ordination axis.

Lichen assemblages on FWD differed near-significantly between age groups of cuts within the same GTR level ($A = 0.1$, $p \geq 0.008$ for both cases), but high and low level GTR sites had similar assemblages within the same cut age (for 4–6 yr. cut sites $A = 0.05$, $p = 0.06$ and for 9–11 yr. sites $A = 0.04$, $p = 0.10$). This pattern is also visible in the NMS ordination (Fig. 3b) and none of the tested environmental variables (volume of FWD and CWD) correlated with the ordination axes ($r^2 < 0.2$). Only three species showed affinity for FWD (*Cladonia fimbriata*, *C. ochrochlora* and *Violella fucata*; Appendix 3).

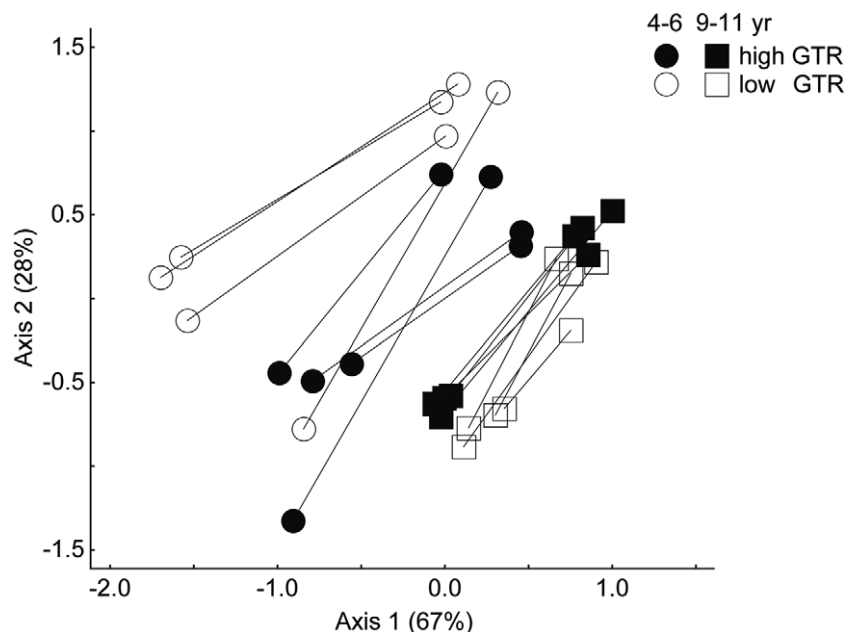


Fig. 4. Lichen assemblages on vertical and horizontal surfaces of pine stumps (symbols on left and right side of the line, respectively) on 4–6 and 9–11 yr. old green tree retention sites with different GTR level (“high” ≥ 14 trees/ha and “low” ≤ 6 trees/ha).

Discussion

The majority of the species found during this study are common lichens of coniferous forests in the hemiboreal region. Even the species newly reported to Latvia reflect the effect of understudy of Latvian lichen diversity than their regional rarity. Moreover, about two-third of the recorded species may also be found on other substrata and habitats apart from pine wood and bark and outside dry hemiboreal forests. Similar studies that examined the lichens on Norway spruce stumps and slash in Sweden (Caruso et al. 2008, Svensson et al. 2013) also reported that the assemblages on the investigated woody substrata were dominated by generalist species. Prevalence of generalist species was also obvious in studies of lichens on stumps of *Pinus sylvestris* in the Netherlands, Poland and Finland (Daniëls 1983, Fałtynowicz 1986, Hämäläinen et al. 2015). One species in our study, however, made an exception: *Cladonia parasitica*, the lichen which is considered as a key woodland habitat indicator species in Latvia (Ek et al. 2002) and was reported from the studied sites for the first time outside of protected habitats in the country.

In addition, our results point to biogeographic differences in lichen diversity and composition between the hemiboreal and middle boreal forests. The list of species found on pine stumps and FWD was shorter than that presented for unburned retention cut-sites in middle boreal forests by Hämäläinen et al. (2015), i.e. 54 vs 77 species, respectively. This

can partly be linked with notably higher sampling effort of stumps per site in Finland, but the low overlap of species list (45%) also indicates differences in assemblage composition on FWD and stumps between middle-boreal and hemiboreal regions. For example, *Evernia prunastri*, *Melanohalea exasperatula*, *Parmelia sulcata*, *Sarea resinae*, *Sclio-sporum chlorococcum* and *Strangospora moriformis* were quite common in our plots, but not in Finland. Meanwhile, *Cetraria sepincola*, *Cladonia bacilliformis*, *Parmeliopsis hyperopta* and *Xylographa vitiligo* were reported as common by Hämäläinen et al. (2015) but not found during our study. Of the latter, *X. vitiligo* is not even known in Latvia and more southern Lithuania. There are very few comparative studies to make strong conclusion about the impact of regional species pools and biogeographic differences in pine forests, however representative study on burned pine forests in Estonia and Eastern Finland showed also that several epiphytes were generally more frequent in the middle-boreal than hemi-boreal region (Lõhmus et al. 2018).

It is known that various stand- and substratum scale factors, such as volumes of dead wood, green tree retention level and diversity of available substrata have an impact on wood-dwelling lichen assemblages, species richness and composition in cut-over sites (e.g., Blasy & Ellis 2014, Ranius et al. 2014, Hämäläinen et al. 2015). The current study showed that among stand-scale factors, green tree retention level (high versus low) had only a slight impact on stump assemblages in early (4–6 yrs.) post-harvest sites, with its importance decreasing in older (9–11 yrs.) sites and showing no effect on the assemblages of downed FWD in the sites of both age groups. Even though sample size for stumps in our study was considerably lower, we obtained similar results as in Hämäläinen et al. (2015), where higher retention level (50 m³/ha vs. 10m³/ha) resulted in slightly higher species richness on cut stumps 12 years after harvest (and was related to positive shade effect of retention trees), but did not affect the assemblages on downed FWD or older stumps (being pre-harvest legacies). Altogether this implies that after the first decade following the harvest, the impact of GTR and its retention level on lichen species diversity on woody substrata has decreased.

Post-harvest age was more important for lichen assemblages than GTR level both on stumps and on FWD. For species richness, it was however statistically important only for stumps. The increase in total species richness on stumps (summed on horizontal and vertical surfaces) was relatively rapid, as richness increased twofold (in low-retention sites) or by one-third (in high-retention sites) between 4–6 to 9–11 yrs. post-harvest. This effect may be related to the development of wood-dependent assemblages (particularly on horizontal stump surfaces) and new substratum qualities developed during the decomposition of stumps (i.e. decay stage, difference between vertical and horizontal wood surfaces; see also Caruso & Rudolphi 2009). In addition, based on Fałtynowicz (1986), in early stages of decay vertical surfaces of pine stumps still retain most of their bark and are largely colonized by the species characteristic to live tree bases; however, after five years, more species including terricolous ones, start to appear, so the communities accumulate diversity by retaining lichens from early successional stages in addition to adding those of later stages. Similar succession was noted by Krüger & Daniels (1998) for cut surfaces of pine stumps, where they noted that stumps aged 10–12 yrs. contained intermediate lichen communities consisting of both early and late stage colonizers. In terms of community

structure, it is difficult to make any comparisons because Fałtynowicz (1986) and Krüger & Daniels (1998) largely neglected crustose species, noting only the most conspicuous and identifiable in field. However, our full assemblage analysis revealed that assemblages on pine stumps become more homogeneous over time (such a trend was also detected if horizontal and vertical surfaces were separated; see Fig. 3a and 4). This might be related to stabilization of late stage colonizers (e.g., *Cladonia* species) and decrease of functional area (i.e. available free space for colonizations) over time.

Similar to stumps, lichen assemblages on FWD change and become more distinct over time. This result cannot be considered obvious, as according to Caruso et al. (2008), Hämäläinen et al. (2015) and the results of our study, downed FWD is a less specific lichen habitat and hosts mainly generalist species that are tolerant to environmental change and (or) have good establishment abilities. In addition, the FWD fraction consists of both pre- and post-cut legacies and thus transitions from corticolous to lignicolous communities can take place already on early post-cut sites. Indeed, several wood-dwelling/terricolous species were found on 4–6 yr. cut sites, like *Cladonia macilenta*, *C. stellaris*, *Micarea denigrata*, *Placynthiella icmalea* and *Trapeliopsis flexuosa*. However, it is possible, that the distinction was shaped by a diminished supply of fallen branches from retention trees, as assemblages on older cut sites were not influenced by GTR level, and the volume of FWD was slightly lower on older sites (which apparently stems from the local tradition of deadwood collection, see below).

In our study, the number of retained logs and snags and their distribution among the sites were insufficient for statistical analysis, so we only included them in general observations of species diversity (Appendix 1). Only one species was found on logs and snags which was not recorded also on stumps and FWD, a common and ubiquitous microlichen *Buellia griseovirens*. Though snags were lower in numbers, the species richness recorded on them was notably high and total occurrence of some species was even higher than on stumps (e.g., *Calicium glaucellum* and *C. trabinellum*). This indicates the importance of snags as valuable habitat for epixylic lichens on harvested sites (see also Runnel et al. 2013). The scarcity of logs and particularly of snags in post-harvest pine forest sites can be explained by local traditions of wood use as a source of firewood for house heating and illegal collection of firewood in cut sites (such local tradition of “cleaning” of cutovers have been reported also in Estonia, see Lõhmus et al. 2013). However, during the last few years the internal rules of Latvia’s State Forests encouraged the practice the retention of “high stumps” (trees cut at height 2–4 m from the ground), that in future potentially can ensure suitable habitats for snag dwelling organisms. This is in agreement with the measures proposed by Blasy & Ellis (2014) and results achieved for saproxylic beetles (Abrahamsson & Lindbladh 2006). However, the impact of such structures on lichen diversity in cut-over sites has yet to be studied.

Given the small sample size of tested treatment types in the current study, general management recommendations cannot be reliably proposed. However, we suggest that in pine dominated *Vacciniosa* type stands, slash and FWD could be harvested from cut over sites in same volumes as it has been up to now; however, the harvest of stumps may negatively affect lichen richness and should be evaluated more thoroughly in Latvia. In

addition, because of alarming scarcity of CWD structures in studied cut-over sites, more CWD structures (particularly snags) should be retained and the post-cut control of retention legacies and the illegal harvest of CWD should be regulated more strictly. We support that leaving high (pine) stumps in cut over sites is useful precautionary management activity for lichen species conservation in industrially exploited forests.

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Appendix 1. The occurrences (observations per treatment or per substratum unit) of lichen species on stumps (A), fine woody debris (B), logs (C) and snags (D) in Latvian green tree retention (GTR) cut sites with different GTR level ("high", ≥ 14 trees/ha and "low", ≤ 6 trees/ha) and time since harvest. Growing on wood (W) and/or bark surface of the stump is distinguished, as well as Red-listed lichens¹ (RL) and species reported as new for Latvia (N; N1 – published already in Moisejevs 2017²). Non-lichenised or saprotrophic calicioid species are marked with asterisk.

| Time since GTR harvest GTR level Type of substratum n = (A-B sites; C-D units) | 4–6 yr. | | | | 9–11 yr. | | | | | |
|---|---------|---|------|---|----------|---|------|---|----|----------------|
| | High | | Low | | High | | Low | | | |
| | A | B | A | B | A | B | A | B | C | D ³ |
| | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 24 | 9 |
| <i>Bryoria fuscescens</i> | – | 2 | – | 2 | – | 1 | – | 2 | – | – |
| <i>Buellia griseovirens</i> | – | – | – | – | – | – | – | – | 14 | 6 |
| <i>Calicium glaucellum</i> | – | – | 1 W | – | 4 W | – | 4 W | – | 15 | 1 |
| <i>Calicium trabinellum</i> ^{N1} | – | – | – | – | 1 W | – | 1 W | – | 10 | 1 |
| <i>Chaenotheca ferruginea</i> | 1 B | 1 | – | 2 | 4 BW | 2 | 2 W | 2 | 1 | 2 |
| <i>Cladonia arbuscula/mitis</i> | 3 BW | – | 1 W | 4 | 3 BW | 3 | 1 BW | 3 | 1 | – |
| <i>Cladonia botrytes</i> | 1 W | 1 | – | – | 4 W | 2 | 4 BW | 2 | 11 | – |
| <i>Cladonia cenotea</i> | 2 BW | – | – | 1 | 4 BW | – | 4 BW | – | 3 | – |
| <i>Cladonia chlorophaea</i> | 1 W | 1 | – | 3 | 4 BW | 3 | – | – | 8 | – |
| <i>Cladonia coniocraea</i> | 4 BW | 3 | 4 BW | 4 | 4 BW | 4 | 4 BW | 3 | 10 | 1 |
| <i>Cladonia cornuta</i> | 2 B | – | – | – | 4 BW | – | 4 BW | – | 3 | – |
| <i>Cladonia digitata</i> | 3 B | 2 | 4 B | 3 | 4 BW | 2 | 3 BW | 2 | 3 | 4 |
| <i>Cladonia fimbriata</i> | 4 B | 2 | – | 3 | 4 BW | 4 | 4 BW | 4 | 9 | 2 |
| <i>Cladonia floerkeana</i> | – | – | – | – | – | 1 | 4 BW | 2 | 1 | – |
| <i>Cladonia juv.</i> | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 23 | 8 |
| <i>Cladonia macilenta</i> | 4 BW | 4 | 1 W | 4 | 4 BW | 4 | 4 BW | 3 | 15 | – |
| <i>Cladonia ochrochlora</i> | 3 BW | – | 1 W | 1 | 4 BW | – | 4 BW | 3 | 5 | – |
| <i>Cladonia parasitica</i> ^{RL} | – | – | – | – | 3 W | – | 4 W | – | 4 | – |
| <i>Cladonia rangiferina</i> | 3 BW | – | 4 B | – | 4 BW | – | 4 BW | – | 3 | – |
| <i>Cladonia stellaris</i> | 3 W | 4 | – | 4 | – | 4 | 1 W | 4 | 1 | – |
| <i>Evernia prunastri</i> | 1 W | 4 | – | 4 | 2 BW | 4 | 2 W | 4 | 4 | 2 |
| <i>Hypocenomyce scalaris</i> | 3 B | 2 | 3 B | 2 | 4 B | 2 | 4 BW | 1 | 10 | 4 |
| <i>Hypogymnia physodes</i> | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 22 | 6 |
| <i>Hypogymnia tubulosa</i> | – | 1 | – | 1 | – | – | 1 BW | 1 | 4 | – |

¹ Anonymous 2013. Regulation of Cabinet of Ministers Nr. 940 (Accepted: 01.01.2013). Noteikumi par mikroliegumu izveidošanas un apsaimniekošanas kārtību, to aizsardzību, kā arī mikroliegumu un to buferzonu noteikšanu. <https://likumi.lv/doc.php?id=253746> [Accessed 20 September 2018]. (in Latvian).

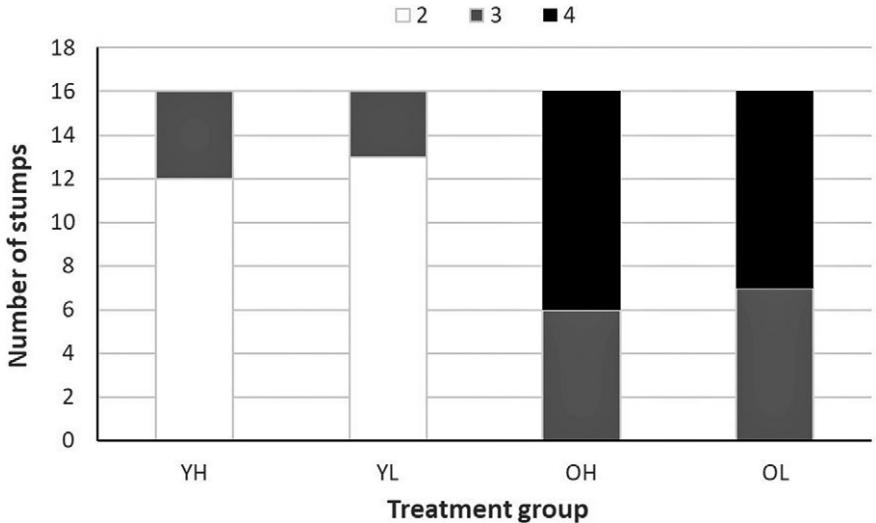
² Moisejevs, R. 2017. Lichens and allied fungi new for Latvia. *Folia Cryptog. Estonica* 54: 9–12.

³ In our study all sampled snags were at least 2.5 m high and had at least 18 cm diameter at breast height (DBH).

Appendix 1. cont.

| Time since GTR harvest GTR level Type of substratum n = (A-B sites; C-D units) | 4–6 yr. | | | | 9–11 yr. | | | | | |
|---|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|----------------|
| | High | | Low | | High | | Low | | C | D ³ |
| | A | B | A | B | A | B | A | B | | |
| | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 24 | 9 |
| <i>Imshaugia aleurites</i> | 2 B | 4 | 4 B | 1 | 2 B | 3 | 4 BW | 3 | 6 | 6 |
| <i>Lecanora pulicaris</i> | 2 B | 2 | – | 3 | 4 B | 4 | 4 BW | 4 | 15 | 7 |
| <i>Lecanora varia</i> | – | – | – | – | 4 W | 1 | 3 BW | – | 17 | 6 |
| <i>Lecidea nylanderii</i> | 4 B | 1 | 2 B | 1 | 4 B | 2 | 3 B | – | 4 | 6 |
| <i>Lecidea turgidula</i> | 3 B | – | 4 B | – | 4 BW | 1 | 1 B | – | – | – |
| <i>Lepraria incana</i> | – | 3 | – | 4 | – | 4 | 3 B | 3 | 5 | 1 |
| <i>Lepraria jackii</i> | 4 BW | – | 4 B | – | 4 BW | – | 4 BW | – | 7 | 1 |
| <i>Melanohalea exasperatula</i> | 1 B | 4 | – | 4 | – | 4 | 3 B | 4 | 8 | 2 |
| <i>Melanohalea olivacea</i> | – | 2 | – | 1 | – | – | – | – | 1 | – |
| <i>Micarea denigrata</i> | – | 2 | 1 W | 3 | 4 BW | 3 | 2 W | – | 7 | – |
| <i>Micarea prasina</i> s.lat | – | 1 | – | – | 3 BW | 1 | 3 W | – | 4 | 2 |
| <i>Mycocalicium subtile</i> * | 4 W | – | 4 W | – | 4 W | – | 4 W | – | 22 | 8 |
| <i>Parmelia sulcata</i> | 3 BW | 4 | 2 BW | 4 | 4 BW | 3 | 4 BW | 3 | 13 | 2 |
| <i>Parmeliopsis ambigua</i> | 3 BW | 4 | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 21 | 2 |
| <i>Physcia adscendens</i> | 1 W | 1 | – | 1 | – | 2 | – | 2 | – | – |
| <i>Physcia tenella</i> | – | – | – | – | 2 BW | – | 1 W | – | 1 | – |
| <i>Placynthiella dasaea</i> | – | – | – | – | 4 W | 1 | 3 W | – | – | – |
| <i>Placynthiella icmalea</i> | 4 W | 4 | 4 W | 4 | 4 W | 4 | 4 BW | 4 | 2 | – |
| <i>Platismatia glauca</i> | – | – | – | 1 | – | 1 | – | 4 | – | – |
| <i>Pseudevernia furfuracea</i> | 4 B | 4 | – | 4 | 2 BW | 4 | – | 4 | 7 | 1 |
| <i>Pycnora sorophora</i> ^{N1} | – | – | – | – | 1 W | – | 1 B | – | 1 | 1 |
| <i>Sarea resinae</i> * | – | 2 | – | 1 | – | 2 | – | 4 | – | – |
| <i>Scoliciosporum chlorococcum</i> | – | 4 | – | 4 | – | 4 | – | 4 | – | – |
| <i>Scoliciosporum sarothamni</i> ^N | – | 1 | – | 1 | – | 2 | – | 2 | – | – |
| <i>Strangospora moriformis</i> | 4 B | 3 | 3 B | 2 | 4 B | 1 | 3 B | 2 | – | – |
| <i>Trapeliopsis flexuosa</i> | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 4 BW | 4 | 2 | – |
| <i>Trapeliopsis granulosa</i> | 3 B | – | 3 W | 1 | 4 BW | 3 | 4 BW | 1 | – | – |
| <i>Tuckermannopsis chlorophylla</i> | – | – | – | 1 | 3 BW | 4 | 2 W | 4 | – | – |
| <i>Usnea hirta</i> | – | 4 | – | 3 | 4 BW | 4 | 2 W | 4 | 3 | 1 |
| <i>Violella fucata</i> | 1 BW | 4 | 1 W | 3 | 3 BW | 4 | 3 W | 4 | 16 | 4 |
| <i>Vulpicida pinastri</i> | 3 BW | 4 | 1 B | 2 | 3 BW | 1 | 4 BW | 3 | 12 | 4 |
| Total | 34 | 35 | 24 | 39 | 42 | 41 | 45 | 36 | 44 | 27 |

Appendix 2. The maximal number of decay classes (five-point scale according to Löhmus & Kraut 2010) per stump among treatment types (YH and YL: 4–6 yr. old sites with high, ≥ 14 trees/ha and low, ≤ 6 trees/ha GTR level, respectively; OH and OL: 9–11 yr. old sites with high and low GTR level, respectively).



Appendix 3. Lichen species association based on indicator species analysis to stumps and fine woody debris (FWD) in pine dominated dry-boreal cut sites with different age and green tree retention (GTR) level. Only species with high indicator value (IV>50%) and statistical significance ($p<0.05$) are given. Cut site treatment types: 1: 4–6 yr. high GTR level (≥ 14 trees/ha); 2: 9–11 yr. high GTR; 3: 9–11 yr. low GTR (≤ 6 trees/ha); 4: 9–11 yr. low GTR (≤ 6 trees/ha).

| Substratum | | | Stump | | FWD | |
|-------------------------------|--------------------|-----------|-------|-----------|--------------------|-----------|
| Treatment | Age × IV(p) GTR | | Age | IV(p) | Age × IV(p) GTR | Age IV(p) |
| <i>Calicium glaucellum</i> | | | 2 | 93(0.001) | | |
| <i>Chaenotheca ferruginea</i> | 3 | 58(0.049) | 2 | 80(0.001) | | |
| <i>Cladonia botrytes</i> | 4 | 57(0.009) | 2 | 93(0.001) | | |
| <i>Cladonia cenotea</i> | | | 2 | 88(0.001) | | |
| <i>Cladonia chlorophaea</i> | 3 | 88(0.006) | | | | |
| <i>Cladonia coniocraea</i> | | | 2 | 63(0.024) | | |
| <i>Cladonia cornuta</i> | 4 | 50(0.055) | 2 | 93(0.001) | | |

Appendix 3. cont.

| Substratum | | | Stump | | FWD | | | |
|-------------------------------------|-----------|------------|-------|---------------|-----------|-----------|-----|-----------|
| Treatment | Age × GTR | IV(p) | Age | IV(p) | Age × GTR | IV(p) | Age | IV(p) |
| <i>Cladonia fimbriata</i> | | | 2 | 77(0.002) | | | 1 | 59(0.024) |
| <i>Cladonia floerkeana</i> | 4 | 100(0.001) | | | | | | |
| <i>Cladonia ochrochlora</i> | | | 2 | 81(0.001) | | | 2 | 68(0.04) |
| <i>Cladonia parasitica</i> | 4 | 67(0.011) | 2 | 88(0.001) | | | | |
| <i>Hypocenomyce scalaris</i> | | | 2 | 68(0.013) | | | | |
| <i>Lecanora pulicaris</i> | 4 | 65(0.001) | 2 | 91(0.001) | | | | |
| <i>Lecanora varia</i> | 3 | 62(0.023) | 2 | 100.0(<0.001) | | | | |
| <i>Lepraria incana</i> | 4 | 75(0.022) | | | | | | |
| <i>Micarea denigrata</i> | 3 | 57(0.050) | 2 | 70(0.015) | | | | |
| <i>Micarea prasina</i> s.lat | | | 2 | 75(0.007) | | | | |
| <i>Placynthiella dasaea</i> | | | 2 | 88(0.001) | | | | |
| <i>Platismatia glauca</i> | | | | | 4 | 57(0.033) | | |
| <i>Pseudevernia furfuracea</i> | 1 | 67(0.036) | | | | | | |
| <i>Trapeliopsis granulosa</i> | | | 2 | 77(0.002) | | | | |
| <i>Tuckermannopsis chlorophylla</i> | | | 2 | 63(0.030) | | | | |
| <i>Usnea hirta</i> | | | 2 | 88(0.003) | | | | |
| <i>Violella fucata</i> | | | 2 | 74(0.020) | | | 2 | 64(0.028) |

II

Ninety-one species of lichens and allied fungi new to Latvia with a list of additional records from Kurzeme

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Abstract: MOTIEJŪNAITĒ, J., CHESNOKOV, S. V., CZARNOTA, P., GAGARINA, L. V., FROLOV, I., HIMELBRANT, D., KONOREVA, L. A., KUBIAK, D., KUKWA, M., MOISEJEVS, R., STEPANCHIKOVA, I., SUIJA, A., TAGIRDZHANOVA, G., THELL, A. & TSURYKAU, A. 2016. Ninety-one species of lichens and allied fungi new to Latvia with a list of additional records from Kurzeme. – *Herzogia* 29: 143–163.

The results of lichenological excursions of the 19th Symposium of Baltic Mycologists and Lichenologists (BMLS) in Latvia, Kurzeme region, 22–26 September 2014, are reported. A list of 290 species is presented, of which 238 are lichenized, 43 lichen-inhabiting, and nine saprotrophic fungi: ninety-one species are new to Latvia, twelve of which (*Caloplaca duplicata*, *Cresporhaphis wienkampii*, *Ellisembia lichenicola*, *Gallowayella weberi*, *Gregorella humida*, *Lichenochora weillii*, *Parmelia serrana*, *Polycauliona phlogina*, *Reconditella physconiarum*, *Stictis brunnescens*, *Thelocarpon superellum*, and *Verrucaria tectorum*) are also new for the Baltic States. *Athallia alnetorum* is reported here for the first time in northern Europe. The presence of *Ochrolechia androgyna* s. str., *Athallia holocarpa* and *A. pyracea* is confirmed for Latvia, and *Parmelia submontana* is reported as a new host for *Homostegia piggottii*.

Zusammenfassung: MOTIEJŪNAITĒ, J., CHESNOKOV, S. V., CZARNOTA, P., GAGARINA, L. V., FROLOV, I., HIMELBRANT, D., KONOREVA, L. A., KUBIAK, D., KUKWA, M., MOISEJEVS, R., STEPANCHIKOVA, I., SUIJA, A., TAGIRDZHANOVA, G., THELL, A. & TSURYKAU, A. 2016. Einundneunzig Flechten, flechtenbewohnende und flechtenähnliche Pilze neu für Lettland, mit einer Liste weiterer Funde aus Kurland. – *Herzogia* 29: 143–163.

Die Ergebnisse der lichenologischen Exkursionen im Rahmen des 19. Symposiums Baltischer Mykologen und Lichenologen (BMLS) in Kurland (Lettland) vom 22.–26. September 2014 werden vorgestellt. Die Liste umfaßt 290 Arten, darunter 238 Flechten, 43 flechtenbewohnende und 9 saprophytische Pilze. 91 Arten sind neu für Lettland, darunter sind 12 (*Caloplaca duplicata*, *Cresporhaphis wienkampii*, *Ellisembia lichenicola*, *Gallowayella weberi*, *Gregorella humida*, *Lichenochora weillii*, *Parmelia serrana*, *Polycauliona phlogina*, *Reconditella physconiarum*, *Stictis brunnescens*, *Thelocarpon superellum* und *Verrucaria tectorum*) neu für die baltischen Staaten. *Athallia alnetorum* wird erstmals für Nordeuropa dokumentiert. Das Vorkommen von *Ochrolechia androgyna* s. str., *Athallia holocarpa* und *A. pyracea* in Lettland wird bestätigt. *Parmelia submontana* ist ein neuer Wirt für *Homostegia piggottii*.

Key words: Lichenized fungi, lichenicolous fungi, biodiversity, protected areas, Kurland, Baltic region.

Introduction

Lichenological investigations in Latvia date back to the 18th century (PĪTERĀNS 1982) and continued to be undertaken until the last decade of the 20th century. However, for the last three decades, new records of lichens and lichenicolous fungi have been mainly reported by partici-

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pants of the Symposia of Baltic Mycologists and Lichenologists (MOTIEJŪNAITĒ & PITERĀNS 1998, MOTIEJŪNAITĒ et al. 2006, CZARNOTA & KUKWA 2010), by fortuitous visiting lichenologists (SUNDIN & THOR 1990, MOTIEJŪNAITĒ & GROCHOWSKI 2014), or as the results of ecological studies (PŁOCIŃA 2007, MEŽAKA 2009, MEŽAKA et al. 2008, 2009, 2012), and only occasionally as targeted lichen diversity studies (PITERĀNS et al. 2005, 2006). Thus, Latvian lichen biota is at present the least studied of the three Baltic States. However, the Kurzeme region is one of the better explored parts of Latvia, especially areas of Nature Reserve Moricsala, Nature Preserve Užava and Slītere National Park (MEŽAKA 2009).

Symposia of the Baltic Mycologists and Lichenologists (BMLS) have been held since 1959 (RANDLANE et al. 2000), during which the attendees have firmly kept a tradition to compile a list of species found during the field trips. A deep knowledge of certain taxonomic groups, field experience and an open mind shared by the participants has always resulted in new species for the country, even for all the Baltic States. The 19th BMLS held in the Kurzeme region (historically known as Kurland) of Latvia proved no exception and as a consequence a detailed species list of lichenized, lichenicolous and allied fungi found by the participants of the Symposium held in Šķēde on 22–26 September 2014 was compiled which significantly contributes to our knowledge of Latvian biota.

Study area

The Kurzeme region, situated in the western part of Latvia, occupies an area of 14,771 km². There are two major industrial cities, Liepāja and Ventspils, and the highest populations are to be found in the districts of Liepājas, Ventspils and Talsi. There is one National Park, over 30 Nature Parks, two Nature Reserves, seven protected dendrological plantations, and over 40 geological and geomorphological protected features (DABAS AIZSARDZĪBAS PĀRVALDE 2015).

The region is bordered by the Baltic Sea to the west and Riga Gulf to the northeast. Highest areas above sea level are in the eastern and southeastern parts of the region, namely the Rietumkursas, Ziemeļkursas and Austrumkursas uplands reaching 189.5 m at Krievukalns. The lowlands are mainly coastal, Piejūras and Kursas, situated in the central part of the region. The largest river is Venta which is 346 km in length and has a catchment area of over 11,000 km² (TURLĀJS 2011). Outcrops of dolomite and sandstone are found along its banks and in the catchment area of the river (RĒRIHA 2013).

The climate of the Kurzeme region is mainly attributed to Atlantic cyclones that bring air masses and precipitation from the west and northwest. The average temperature ranges from –3 to –6°C in January, and from 16.5 to 17°C in July. Annual precipitation ranges from 600 to 850 mm, and snow cover remains for 80–110 days (TURLĀJS 2011, ANONYMOUS 2013).

A number of habitat types included into Annex I of European Union Habitat Directive are found in the region, such as 4110 Northern Atlantic wet heaths with *Erica tetralix*, 6530 *Fennoscandian wooded meadows, 9010 *Western taiga, 9020 *Fennoscandian hemiboreal natural old deciduous forests rich in epiphytes, 9050 Fennoscandian herb-rich forests with *Picea abies*, 9080 *Fennoscandian deciduous swamp woods, 9160 Sub-Atlantic and medio-European oak forests, 9180 Tilio-Acerion forests of slopes, scree and ravines, 5130 *Juniperus communis* formations on heaths or calcareous grasslands, 91E0 *Alluvial forests with *Alnus glutinosa* and *Fraxinus excelsior*, 91D0 Bog woodland, 8220 Siliceous (sandstone) rocky slopes with chasmophytic vegetation, 8210 Calcareous rocky slopes with chasmophytic vegetation (ANONYMOUS 2013).

List of visited localities:

A map of the localities is presented in Fig. 1.

- 1 23.09.2014. Talsi district, Šķēde Forest Research Station, Mežmāja Šķēde, park, 57°14'53.4"N, 22°41'39.9"E, village, gardens, roadside scarps, planted arboretum and some old hardwood trees;
- 2 23.09.2014. Talsi district, vicinity of Roceži, 57°16'17"N, 22°37'31"E, village, broad-leaved trees in the roadside alley and in the village;
- 3 24.09.2014. Talsi district, NE of Odre village, 57°14'03.3"N, 22°43'01.0"E, old *Populus tremula* forest with spruce and birch, old farm place;
- 4 24.09.2014. Talsi district, Nature Park "Talsu pauguraine", SE of Sukturi, 57°14'N, 22°38'E, old planted *Larix decidua* stand with *Corylus avellana*, Sapnu ezers peat-bog;
- 5 24.09.2014. Talsi district, Nature Park "Talsu pauguraine", 57°15'06.9"N, 22°42'51.9"E, old planted *Abies alba* forest with spruce, birch and hardwood trees;
- 6 24.09.2014. Talsi district, Talsi, c. 2 km south-east of the city, 57°14'08.1"N, 22°38'04.2"E, roadside broad-leaved trees;
- 7 25.09.2014. Dundaga district, Slītere National Park, Mazirbe, 57°41'25.4"N, 22°19'02.1"E, the village, trees in the roadside alley, roadside sandy meadow, alder stand, pine forest on seashore sand dunes;
- 8 25.09.2014. Dundaga district, Slītere National Park, Slītere lighthouse environs, along Slītere Nature Trail, 57°37'51.3"N, 22°17'30.8"E, mixed old-growth forest with spruces and broad-leaved trees, roadside trees between the parking area and lighthouse;
- 9 25.09.2014. Dundaga district, Dundaga town, 57°30'38.0"N, 22°21'15.1"E, park of Dundaga Castle: old hardwood trees;
- 10 26.09.2014. Kuldīga district, Kuldīga town, 56°58'12.21"N, 21°58'24.05"E, town park, trees in the roadside alleys.

Material and methods

Air-dried samples of lichens and fungi were examined using stereomicroscopes and light microscopes and the usual spot tests with standard identification methods for lichenized and lichenicolous fungi (SMITH et al. 2009). Thin Layer Chromatography (TLC) analyses (ORANGE et al. 2001) were carried out where required. In some cases, the opinion of experts was requested (see Acknowledgements). To check the identification of *Athallia alnetorum* and *Polycauliona phlogina*, the DNA barcoding was employed (the ITS regions including the 5.8S gene of the nuclear rDNA). The CTAB method (ARAS & CANSARAN 2006) was used for DNA isolations. Primers for PCR amplification of ITS were ITS1F (GARDES & BRUNS 1993) and ITS4 (WHITE et al. 1990). PCR cycling parameters followed EKMAN (2001). The most similar known sequences were found using the BLAST search in the GenBank.

The nomenclature of taxa follows NORDIN et al. (2011), except for some recent nomenclatural changes (ERTZ et al. 2015, SCHULTZ et al. 2015, ZHAO et al. 2016). Species new for Latvia are supplied with the collectors' names and herbaria in which the specimens are deposited.

Abbreviations of collectors: AS = Ave Suija, AT = Arne Thell, ANT = Andrei Tsurykau, DH = Dmitry Himelbrant, DK = Dariusz Kubiak, GT = Gulnara Tagirdzhanova, IF = Ivan Frolov, IS = Irina Stepanchikova, JM = Jurga Motiejūnaitė, LG = Ludmila Gagarina, LK = Ludmila A. Konoreva, MK = Martin Kukwa, PC = Paweł Czarnota, and SC = Sergei V. Chesnokov.

Abbreviations of herbaria: BILAS = Institute of Botany, Vilnius, Lithuania, GPN = Gorce National Park, Poland, GSU = F. Skorina Gomel State University, Belarus, LD = Botanical Museum of Lund University, Sweden, LE = V. L. Komarov Botanical Institute, Saint-Petersburg, Russia, LECB = Saint-Petersburg State University, Russia, OLTC = Department of Mycology, Warmia and Mazury University in Olsztyn, Poland, TU = Botanical and

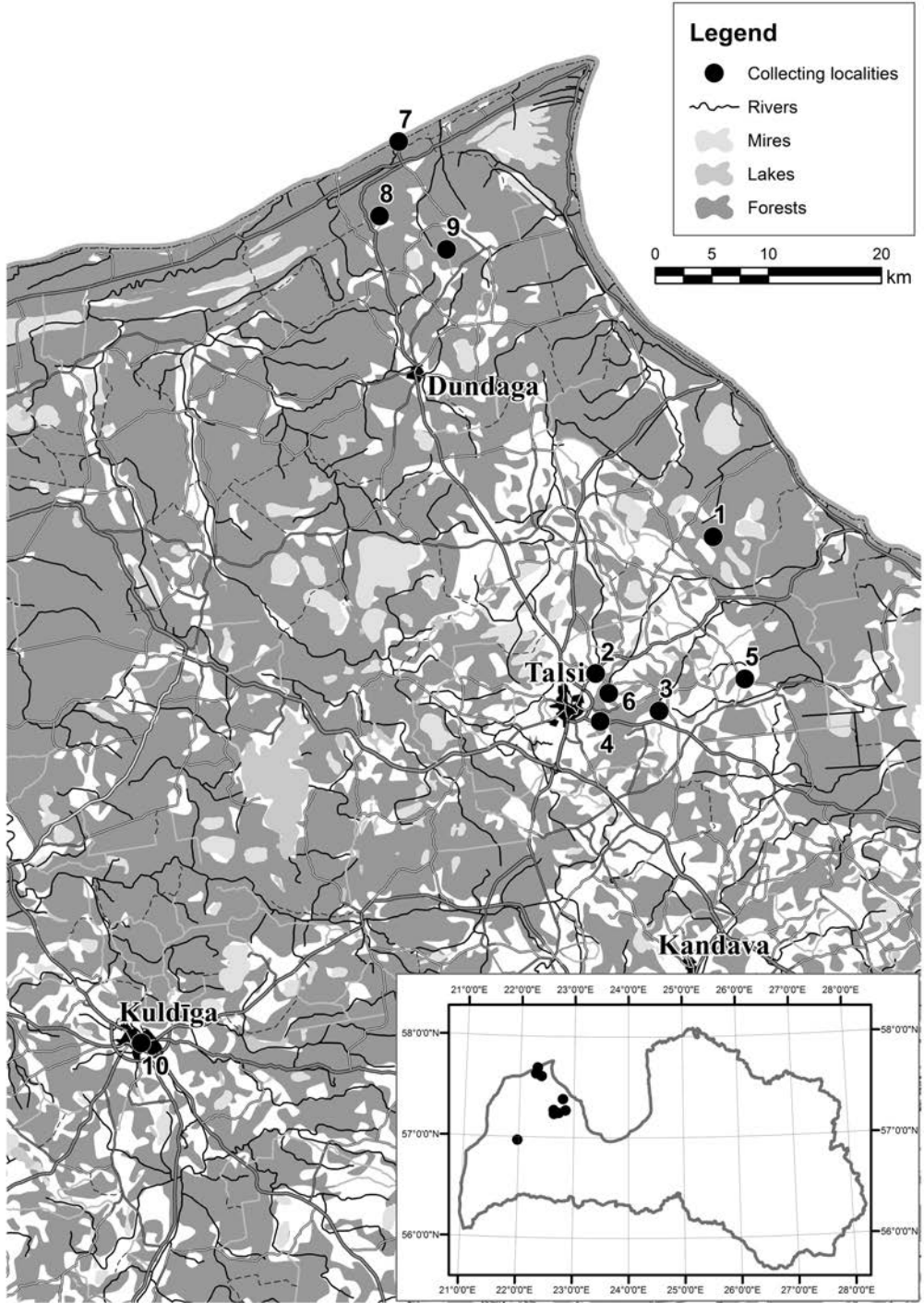


Fig. 1: Visited localities in Kurzeme Region and their situation in Latvia.

Mycological Museum, Natural History Museum of the University of Tartu, Estonia, UGDA = Gdańsk University, Poland, and Herb. IF = Private herbarium of Ivan Frolov.

List of substrata and their abbreviations: Abi = *Abies alba*, Ace = *Acer platanoides*, Aes = *Aesculus hippocastanum*, Aln = *Alnus glutinosa*, Bet = *Betula* spp., Car = *Carpinus betulus*, Coa = *Corylus avellana*, Fag = *Fagus sylvatica*, Fra = *Fraxinus excelsior*, Lar = *Larix decidua*, Mal = *Malus domestica*, Pic = *Picea abies*, Pin = *Pinus sylvestris*, Pop = *Populus tremula*, Pru = *Prunus padus*, Pyr = *Pyrus communis*, Qur = *Quercus robur*, Sal = *Salix* spp., Sor = *Sorbus aucuparia*, Til = *Tilia cordata*, Ulm = *Ulmus* spp.; cal = concrete and other anthropogenic calcareous substrata, cor = corticolous, lig = lignicolous, res = resinicolous, roo = roots of windthrows, sil = siliceous stones, mus = muscicolous (over soil, bark, stone), ter = terricolous (including plant debris).

List of taxa

! = new to Latvia; !! = new to all Baltic States.

Absconditella lignicola Vězda & Pišút: 5, 8 (lig Bet, Pic).

Acarospora fuscata (Schräd.) Th.Fr.: 1 (sil).

Acrocordia cavata (Ach.) R.C.Harris: 3, 4 (cor Coa, Pop, Qur).

Acrocordia gemmata (Ach.) A.Massal.: 1, 3, 4, 8 (cor Fag, Fra, Pop, Qur).

Alyxoria culmigena (Lib.) Ertz: 1 (cor Fag), SC (LE).

Alyxoria varia (Pers.) Ertz & Tehler: 3, 4, 5, 7, 8, 9 (cor Aes, Aln, Fra, Lar, Pop, Ulm).

Amandinea punctata (Hoffm.) Coppins & Scheid.: 1, 4, 7, 8, 9, 10 (cor, lig Aes, Pic, Qur, Til).

Anaptychia ciliaris (L.) Körb.: 1, 4, 5, 8 (cor Qur, Pic, Pop).

! *Anisomeridium polypori* (M.B.Ellis & Everh.) M.E.Barr: 4, 5, 7, 8 (cor Qur, Til), DH (LECB), PC (GPN).

Arthonia didyma Körb.: 1, 3, 4, 5 (cor Abi, Coa, Pop, Sor).

Arthonia leucopellaea (Ach.) Almq.: 8 (cor Bet, Pic, Pin).

Arthonia radiata (Pers.) Ach.: 3, 5 (cor Ace, Coa, Fag).

Arthonia ruana A.Massal.: 3, 4, 5 (cor Aln, Coa).

Arthonia spadicea Leight.: 1, 3, 4, 5, 8 (cor Abi, Aln, Coa, Fag, Qur, Ulm).

!! *Athallia alnetorum* (Giralt, Nimis & Poelt) Arup, Frödén & Søchting: 6, 7 (cor, sil Pru, Sal) IF (Herb. IF).

! *Athallia cerinella* (Nyl.) Arup, Frödén & Søchting: 7 (cor Sal), IF (Herb. IF).

! *Athallia holocarpa* (Hoffm.) Arup, Frödén & Søchting: 7 (lig, sil), DH (LECB), IF (Herb. IF).

! *Athallia pyracea* (Ach.) Arup, Frödén & Søchting: 2, 3, 7 (cor Aln, Fra, Sal), DH (LECB), IF (Herb. IF), MK (UGDA), JM (BILAS).

Bacidia arceutina (Ach.) Arnold: 1 (cor Pyr).

Bacidia bagliettoana (A.Massal. & De Not.) Jatta: 7 (mus, ter).

! *Bacidia fraxinea* Lönnr.: 3 (cor Pop), DH (LECB), PC (GPN).

- Bacidia rubella* (Hoffm.) A.Massal.: 1, 3, 4, 7, 8 (cor, mus Fag, Lar, Pop, Qur, Til).
- Bacidia subincompta* (Nyl.) Arnold: 1, 4 (cor, mus Fag, Pyr).
- Bacidina arnoldiana* (Körb.) V.Wirth & Vězda s. lat.: 3 (lig).
- !*Bacidina delicata* (Larbal. ex Leight.) V.Wirth & Vězda: 3 (cor Sal), PC (GPN).
- Bacidina phacodes* (Körb.) Vězda: 3 (cor Qur).
- Baeomyces rufus* (Huds.) Rebert.: 7 (ter).
- Biatora efflorescens* (Hedl.) Räsänen: 4, 5, 8 (cor, mus Abi, Bet, Pic, Qur).
- Biatora globulosa* (Flörke) Fr.: 1, 4 (cor Qur).
- Biatora helvola* Körb. ex Hellb.: 5, 8 (cor Abi, Pic).
- !*Biatora meiocarpa* (Nyl.) Arnold: 8 (cor Pic), MK (UGDA).
- Biatoridium monasteriense* J.Lahm ex Körb.: 5 (cor Fag).
- Bilimbia sabuletorum* (Schreb.) Arnold: 9 (cal, mus).
- Bryoria fuscescens* (Gyeln.) Brodo & D.Hawksw.: 1 (lig).
- Buellia griseovirens* (Turner & Borrer ex Sm.) Almb.: 1, 3, 4, 5, 7, 8 (cor, lig Abi, Aln, Bet, Coa, Pic, Qur, Sal).
- Buellia schaeferi* De Not.: 8 (cor Qur).
- Calicium adpersum* Pers.: 9 (cor Qur).
- Calicium glaucellum* Ach.: 7 (cor, lig Pin).
- Calicium quercinum* Pers.: 9 (cor Qur).
- Calicium salicinum* Pers.: 1, 9 (cor, lig Til).
- Calicium viride* Pers.: 1, 8, 9 (cor Qur, Til).
- Caloplaca cerina* (Ehrh. ex Hedw.) Th.Fr. var. *cerina*: 3, 4 (cor Ace, Pru, Qur).
- !*Caloplaca chlorina* (Flot.) Sandst.: 7 (sil), IF (Herb. IF).
- !!*Caloplaca duplicata* (Vain.) H.Olivier: 7 (sil), IF (Herb. IF).
- Caloplaca obscurella* (J.Lahm. ex Körb.) Th.Fr.: 1 (cor Pyr).
- !*Caloplaca ulcerosa* Coppins & P.James: 7 (cor Sal), IF (Herb. IF).
- !*Candelariella efflorescens* R.C.Harris & W.R.Buck: 1 (cor Pyr), DH (LECB).
- Candelariella vitellina* (Hoffm.) Müll.Arg.: 7 (lig).
- Candelariella xanthostigma* (Ach.) Lettau: 1, 4, 7, 8, 10 (cor Aes, Pru, Qur, Til).
- Catillaria nigroclavata* (Nyl.) Schuler: 3, 7 (cor Aln, Sal).
- Chaenotheca brachypoda* (Ach.) Tibell: 3, 4, 5, 8 (cor, lig Lar, Pop, Qur).
- Chaenotheca chlorella* (Ach.) Müll.Arg.: 3 (cor, lig Bet, Pic).
- Chaenotheca chrysocephala* (Turner ex Ach.) Th.Fr.: 1, 3, 4, 5 (cor, lig Abi, Bet, Pin, Qur).
- Chaenotheca ferruginea* (Turner ex Sm.) Mig.: 4 (cor Pin).
- Chaenotheca furfuracea* (L.) Tibell: 3, 4, 8 (cor, roo Bet, Pic, Ulm).

- Chaenotheca gracilentia* (Ach.) Mattsson & Middelb.: 8 (cor Ulm).
- ! *Chaenotheca hispidula* (Ach.) Zahlbr.: 9 (cor Til), IS (LECB).
- Chaenotheca phaeocephala* (Turner) Th.Fr.: 3 (lig).
- Chaenotheca stemonea* (Ach.) Müll.Arg.: 1, 3, 9 (cor Pic, Qur, Til).
- Chaenotheca trichialis* (Ach.) Th.Fr.: 1, 3, 4, 7, 8, 9 (cor, lig Lar, Pic, Qur, Til).
- ! *Chaenotheca xyloxena* Nádv.: 8 (lig), PC (GPN).
- Chrysothrix candelaris* (L.) J.R.Laundon: 1, 8, 9 (cor Qur, Til).
- ! *Chrysothrix flavovirens* Tønsberg: 5 (cor Pic, Pin), AS (TU), MK (UGDA, GSU), JM (BILAS).
- Cladonia chlorophaea* (Flörke ex Sommerf.) Spreng.: 5 (cor Pic, Pin).
- Cladonia coniocraea* (Flörke) Spreng.: 5, 8 (cor Abi, Fra, Lar, Pic, Qur, Til).
- Cladonia cornuta* (L.) Hoffm.: 7 (ter).
- Cladonia digitata* (L.) Hoffm.: 7, 8 (cor, lig Aln).
- Cladonia fimbriata* (L.) Fr.: 1 (cor Lar).
- Cladonia furcata* (Huds.) Schrad.: 7 (ter).
- Cladonia gracilis* (L.) Willd. subsp. *gracilis*: 7 (ter).
- Cladonia macilenta* Hoffm.: 8 (cor Aln).
- Cladonia ochrochlora* Flörke: 5 (cor Aln).
- Cladonia pyxidata* (L.) Hoffm.: 7 (ter).
- Cladonia rangiferina* (L.) F.H.Wigg.: 7 (ter).
- Cladonia subulata* (L.) Weber ex F.H.Wigg.: 7 (ter).
- Cliostomum griffithii* (Sm.) Coppins: 1, 4, 5, 7, 9 (cor Abi, Lar, Qur, Sor, Til).
- Coenogonium pineti* (Ach.) Lücking & Lumbsch: 3, 5, 7, 8 (cor, lig Bet, Pic, Pop, Qur).
- Collema limosum* (Ach.) Ach.: 1 (ter).
- Coniocarpon cinnabarinum* DC.: 8 (cor Coa).
- !! *Cresporhaphis wienkampii* (J.Lahm ex Hazsl.) M.B.Aguirre: 4 (cor Sal), DH & IS (LECB).
- ! *Diplotomma pharcidium* (Ach.) M.Choisy: 2 (cor Fra), AS (TU).
- ! *Epyrenula leucoplaca* (Wallr.) R.C.Harris: 8 (cor Til), DH (LECB).
- Evernia prunastri* (L.) Ach.: 1, 4, 5, 10 (cor, lig Abi, Aes, Bet, Qur, Pyr, Til).
- ! *Fuscidea arboricola* Coppins & Tønsberg: 5, 8 (cor, lig Abi, Pic, Pin), MK (UGDA), PC (GPN).
- !! *Gallowayella weberi* (S.Y.Kondr. & Kärnefelt) S.Y.Kondr., Fedorenko, S.Stenroos, Kärnefelt, Elix, J.-S.Hur & A.Thell: 10 (cor Aes), AT (LD).
- Graphis scripta* (L.) Ach.: 3, 4, 5, 8 (cor Aln, Bet, Coa, Fag, Fra, Qur).
- !! *Gregorella humida* (Kullh.) Lumbsch: 1 (ter), DH (LECB), JM (BILAS).
- ! *Gyalecta derivata* (Nyl.) H.Olivier: 3 (cor Qur), JM (BILAS).
- ! *Gyalecta flotowii* Körb.: 4, 8 (cor Ulm), LG (LE), DH & GT (LECB).

- Gyalecta truncigena* (Ach.) Hepp: 4, 8 (cor Pop, Qur).
- Gyalolechia flavorubescens* (Huds.) Söchting, Frödén & Arup: 2, 3, 4 (cor Fra, Pop, Pru, Qur).
- Hypocenomyce scalaris* (Ach.) M.Choisy: 1 (cor Lar, Pin).
- Hypogymnia physodes* (L.) Nyl.: 1–10 (cor, lig Abi, Aln, Bet, Fra, Lar, Pic, Pin, Pru, Qur, Til).
- Hypogymnia tubulosa* (Schaer.) Hav.: 1, 4, 7 (cor, lig Bet, Lar).
- Imshaugia aleurites* (Ach.) S.L.F.Mey.: 8 (cor Pic).
- Lecanactis abietina* (Ach.) Körb.: 1, 3, 4, 5, 8, 9 (cor Abi, Aln, Bet, Lar, Pic, Til).
- Lecania cyrtella* (Ach.) Th.Fr.: 2, 3, 4, 7, 8 (cor Aln, Fra, Pru, Sal).
- Lecania naegelii* (Hepp) Diederich & van den Boom: 1, 3, 4, 7 (cor, lig Ace, Aln, Fag, Fra, Sal).
- Lecanora albellula* (Nyl.) Th.Fr.: 1, 4 (lig).
- Lecanora allophana* (Ach.) Nyl.: 2, 3, 7 (cor Ace, Fra, Pop, Ulm).
- Lecanora argentata* (Ach.) Malme: 4, 7, 8, 9 (cor Qur, Sal, Til).
- Lecanora carpinea* (L.) Vain.: 1, 3, 4, 5, 7, 8 (cor, lig Ace, Aln, Fra, Qur, Sal, Sor, Til).
- Lecanora chlarotera* Nyl.: 1–5, 7 (cor, lig Bet, Fag, Fra, Qur, Sor, Til).
- Lecanora compallens* Herk & Aptroot: 1, 3, 5, 7 (cor Abi, Pic, Pin, Pyr, Til).
- Lecanora expallens* Ach.: 4, 5, 7 (cor Abi, Bet, Qur).
- !Lecanora farinaria* Borrer: 5 (cor Abi), MK (UGDA).
- Lecanora leptyroides* G.B.F.Nilsson: 1–4 (cor Bet, Fra, Qur, Til).
- Lecanora pulicaris* (Pers.) Ach.: 1, 7 (cor Bet, Pin).
- Lecanora rugosella* Zahlbr.: 3, 4, 7 (cor Coa, Fag, Til).
- Lecanora sambuci* (Pers.) Nyl.: 2 (cor Fra).
- !Lecanora subcarpinea* Szatala: 1 (cor Car), MK (UGDA).
- Lecanora symmicta* (Ach.) Ach.: 1, 4, 5 (cor, lig Bet, Pop, Qur).
- Lecanora varia* (Hoffm.) Ach.: 1 (cor Sor).
- Lecidea albohyalina* (Nyl.) Th.Fr.: 1 (cor Abi).
- Lecidea nylanderii* (Anzi) Th.Fr.: 3, 5 (cor Pic, Pin, Qur).
- Lecidella elaeochroma* (Ach.) M.Choisy: 1–5, 7, 8 (cor, lig Aln, Fag, Fra, Pin, Pru, Qur, Sal, Til).
- Lecidella elaeochroma* (Ach.) M.Choisy f. *soralifera* (Erichsen) D.Hawksw.: 8 (cor Til).
- Lecidella euphorea* (Flörke) Hertel: 7 (cor Bet).
- Lecidella flavosorediata* (Vězda) Hertel & Leuckert: 1, 4, 5, 8 (cor, lig Abi, Fra, Qur).
- Lecidella stigmataea* (Ach.) Hertel & Leuckert: 9 (cal).
- Lecidella subviridis* Tønsberg: 4 (cor Qur).
- Lepraria elobata* Tønsberg: 1, 5, 8 (cor Abi, Aln, Qur).
- Lepraria incana* (L.) Ach.: 1, 4, 5, 8, 9 (cor Abi, Bet, Lar, Qur).
- Lepraria lobificans* Nyl.: 4, 5, 7, 8 (cor, roo Abi, Fag, Fra).

- Lepraria rigidula* (de Lesd.) Tønsberg: 1 (cor Pyr).
- ! *Leptogium byssinum* (Hoffm.) Nyl.: 1 (ter), ANT (GSU).
- Lobaria pulmonaria* (L.) Hoffm.: 8, 9 (cor Ace, Qur).
- Loxospora elatina* (Ach.) A.Massal.: 1, 5, 8 (cor, lig Abi, Pic).
- Melanelixia glabrata* (Lamy) Sandler & Arup: 1, 3, 4, 5, 8 (cor, lig Coa, Pru, Sor, Til).
- Melanelixia subargentifera* (Nyl.) O.Blanco, A.Crespo, Divakar, Essl., D.Hawksw. & Lumbsch: 8, 10 (cor Aes, Qur).
- Melanelixia subaurifera* (Nyl.) O.Blanco, A.Crespo, Divakar, Essl., D.Hawksw. & Lumbsch: 1 (cor Mal).
- Melanohalea exasperata* (De Not.) O.Blanco, A.Crespo, Divakar, Essl., D.Hawksw. & Lumbsch: 1, 4, 7, 8 (cor, lig Bet, Qur).
- Melanohalea exasperatula* (Nyl.) O.Blanco, A.Crespo, Divakar, Essl., D.Hawksw. & Lumbsch: 1, 3, 4, 5, 7 (cor, lig Bet, Pic, Qur, Sor).
- ! *Micarea anterior* (Nyl.) Hedl.: 5 (lig), DH (LECB).
- ! *Micarea byssacea* (Th.Fr.) Czarnota, Guzow-Krzemińska & Coppins: 7, 8 (cor, lig Pin), PC (GPN).
- Micarea denigrata* (Fr.) Hedl.: 5, 9 (lig).
- ! *Micarea micrococca* (Körb.) Gams ex Coppins: 8 (cor Pic), PC (GPN).
- Micarea misella* (Nyl.) Hedl.: 5, 7, 8 (lig).
- ! *Micarea peliocarpa* (Anzi) Coppins & R.Sant.: 8 (lig), LK (LE), PC (GPN).
- Micarea prasina* Fr.: 3, 4, 5, 8 (lig).
- Multiclavula mucida* (Pers.) R.H.Petersen: 8 (lig).
- Mycoblastus sanguinarius* (L.) Norman: 8 (cor Bet).
- Myriolecis albescens* (Hoffm.) Śliwa, Zhao Xin & Lumbsch: 7 (lig, dust impregnated).
- Myriolecis dispersa* (Pers.) Śliwa, Zhao Xin & Lumbsch: 9 (sil).
- Myriolecis hagenii* (Ach.) Śliwa, Zhao Xin & Lumbsch: 2, 3, 7, 8, 9 (cor, lig Aln, Fra, Til).
- Myriolecis persimilis* (Th. Fr.) Śliwa, Zhao Xin & Lumbsch: 7 (cor Ace, Pin).
- Naetrocymbe punctiformis* (Pers.) R.C.Harris: 8 (cor Qur).
- ! *Normandina acroglypta* (Norman) Aptroot: 4 (cor, mus Pop), JM (BILAS, GSU).
- ! *Ochrolechia androgyna* (Hoffm.) Arnold: 8 (cor Aln), MK (UGDA), DK (OLTC).
- Ochrolechia arborea* (Kreyer) Almb.: 5 (cor Abi).
- ! *Ochrolechia microstictoides* Räsänen: 1, 3, 4, 7 (cor, lig Bet, Pic, Pin), MK (UGDA), DH & GT (LECB).
- Ochrolechia turneri* (Sm.) Hasselrot: 1 (cor Pyr).
- Opegrapha vermicellifera* (Kunze) J.R.Laundon: 8 (cor Fr, Ulm).
- Opegrapha vulgata* (Ach.) Ach.: 5, 9 (cor Qur, Til).
- Pachyphiale fagicola* (Hepp ex Arnold) Zwackh: 1, 5 (cor Fag).
- Parmelia saxatilis* (L.) Ach.: 8 (cor Bet, Pic).

!!*Parmelia serrana* A.Crespo, M.C.Molina & D.Hawksw.: 8 (cor Coa, Qur), AT (LD, LECB), ANT (GSU), AS (TU), JM (BILAS).

!*Parmelia submontana* Nádv. ex Hale: 1, 4 (cor Ace, Pru), DH (LECB), JM (BILAS).

Parmelia sulcata Taylor: 1, 4, 5, 7 (cor, lig, sil Ace, Bet, Fra, Qur, Sor).

Parmeliopsis ambigua (Wulfen) Nyl.: 4, 5 (cor Bet, Pin).

Peltigera aphthosa (L.) Willd.: 7 (ter).

Peltigera canina (L.) Willd.: 3, 7 (ter, lig).

Peltigera hymenina (Ach.) Delise: 7 (ter).

Peltigera malacea (Ach.) Funck: 7 (ter).

Peltigera neckeri Hepp ex Müll.Arg.: 7 (ter).

Peltigera praetextata (Flörke ex Sommerf.) Zopf: 3, 5, 8 (cor, ter Fra, Lar, Pop).

Peltigera rufescens (Weiss) Humb.: 7, 9 (ter).

Pertusaria albescens (Huds.) M.Choisy & Werner: 1, 3, 5, 9 (cor Ace, Pop, Pyr, Qur, Til).

Pertusaria amara (Ach.) Nyl.: 1, 4, 5, 8, 9 (cor Abi, Bet, Pic, Qur, Sor, Til).

Pertusaria coccodes (Ach.) Nyl.: 1, 4, 7, 8, 9 (cor Ace, Fra, Pop, Pyr, Qur, Til).

Pertusaria coronata (Ach.) Th.Fr.: 1, 4, 8, 9 (cor Fra, Qur, Til).

Pertusaria flavida (DC.) J.R.Laundon: 8, 9 (cor Aln, Fra, Qur).

Pertusaria leioplaca DC.: 3, 4, 8 (cor Coa, Fag, Fra, Ulm).

Pertusaria pertusa (Weigel) Tuck.: 8, 9 (cor Fra, Qur, Til).

Phaeophyscia ciliata (Hoffm.) Moberg: 4 (cor Qur).

!*Phaeophyscia endophoenicea* (Harm.) Moberg: 8 (cor, mus Fra, Til), DH (LECB), PC (GPN).

Phaeophyscia nigricans (Flörke) Moberg: 8 (cor Qur).

Phaeophyscia orbicularis (Neck.) Moberg: 2, 4, 7, 10 (cor Aes, Fra, Pop, Sal).

Phlyctis argena (Ach.) Flot.: 1, 3, 4, 5, 7, 8, 10 (cor Abi, Aes, Aln, Bet, Coa, Lar, Pop, Pru, Qur, Sor, Til).

Physcia adscendens (Fr.) H.Olivier: 1, 3, 8 (cor Aln, Til, Pic, Qur).

Physcia aipolia (Ehrh. ex Humb.) Fürnr.: 1, 2, 3, 5, 7, 8 (cor Aln, Bet, Qur).

Physcia stellaris (L.) Nyl.: 9 (cor Til).

Physcia tenella (Scop.) DC.: 1, 3, 4, 7, 8 (cor, sil Aln, Pic, Pin, Prun, Qur, Til).

Physconia detera (Nyl.) Poelt: 1, 8 (cor Qur, Til).

Physconia distorta (With.) J.R.Laundon: 1, 4, 5, 7, 8 (cor Bet, Pru, Qur).

Physconia enteroxantha (Nyl.) Poelt: 1, 8, 10 (cor Aes, Fag, Pyr, Qur).

Physconia perisidiosa (Erichsen) Moberg: 1, 8, 10 (cor Aes, Qur).

Placynthiella dasaea (Stirt.) Tønsberg: 8 (cor Aln, Ulm).

Placynthiella icmalea (Ach.) Coppins & P.James: 1, 5, 7 (lig).

Platismatia glauca (L.) W.L.Culb. & C.F.Culb.: 1, 4, 5 (cor Bet, Pic, Pin, Qur).

- Pleurosticta acetabulum* (Neck.) Elix & Lumbsch: 1 (cor Qur).
- Polycauliona candelaria* (L.) Frödén, Arup & Söchting: 9 (cor Fra).
- !!*Polycauliona phlogina* (Ach.) Arup, Frödén & Söchting: 7 (lig), IF (Herb. IF), SC (LE).
- Polycauliona polycarpa* (Hoffm.) Frödén, Arup & Söchting: 1, 4, 8 (cor, lig Fra, Qur, Sal).
- !*Protothelenella sphinctrinoidella* (Nyl.) H.Mayrhofer & Poelt: 1 (ter), SC (LE), IS (LECB).
- Pseudevernia furfuracea* (L.) Zopf: 1, 4, 5, 8 (cor Bet, Pic, Pin, Qur, Sor).
- !*Psilolechia clavulifera* (Nyl.) Coppins: 8 (roo Pic), PC (GPN).
- Pseudoschismatomma rufescens* (Pers.) Ertz & Tehler: 1, 4, 7, 8 (cor Fag, Fra, Til).
- Pyrenula nitidella* (Flörke ex Schaer.) Müll.Arg.: 8 (cor Coa, Fra).
- Ramalina baltica* Lettau: 5 (cor Qur).
- Ramalina calicaris* (L.) Fr.: 5 (cor Abi).
- Ramalina farinacea* (L.) Ach.: 1, 3, 4, 5 (cor, lig Bet, Pic, Qur, Sor).
- Ramalina fastigiata* (Pers.) Ach.: 1, 4, 5 (cor Fra, Qur, Til).
- Ramalina fraxinea* (L.) Ach.: 1, 3, 4, 7 (cor Fra, Qur, Sal).
- Ramalina pollinaria* (Westr.) Ach.: 1, 5 (cor Qur).
- Ramalina roesleri* (Hochst. ex Schaer.) Hue: 1 (cor Qur).
- Ramalina sinensis* Jatta: 1 (cor Sor).
- Rinodina pyrina* (Ach.) Arnold: 2, 4, 5 (cor Qur, Sal).
- !*Rinodina septentrionalis* Malme: 7 (cor Bet), LK (LE).
- !*Rinodina sophodes* (Ach.) A.Massal.: 8 (cor Qur), DH (LECB).
- Ropalospora viridis* (Tønsberg) Tønsberg: 4 (cor Bet, Coa, Fra).
- Sarcosagium campestre* (Fr.) Poetsch & Schied.: 1 (ter).
- !*Schismatomma pericleum* (Ach.) Branth & Rostr.: 5 (cor Abi), DH (LECB).
- Sclerophora pallida* (Pers.) Y.J.Yao & Spooner: 8, 9 (cor Fra, Qur).
- Scoliciosporum chlorococcum* (Graewe ex Stenh.) Vězda: 4 (cor Sal).
- !*Scoliciosporum sarothamni* (Vain.) Vězda: 1, 3 (cor, lig Aln, Bet, Pic, Qur), SC (LE), DH & GT (LECB), PC (GPN).
- Scytinium lichenoides* (L.) Otálora, P.M.Jørg. & Wedin: 8 (cor, mus Fra).
- !*Solitaria chrysophthalma* (Degel.) Arup, Frödén & Söchting: 7 (lig), DH (LECB).
- !*Steinia geophana* (Nyl.) Stein: 1 (ter), ANT (GSU), AS (TU).
- !*Strigula jamesii* (Swinscow) R.C.Harris: 3, 8 (cor Coa, Fra), IS, DH (LECB).
- Thelidium zwackhii* (Hepp) A.Massal.: 1 (ter).
- !!*Thelocarpon superellum* Nyl.: 7 (ter), PC (GPN).
- Thelotrema lepadinum* (Ach.) Ach.: 8 (cor Coa, Fra, Qur).
- !*Trapelia corticola* Coppins & P.James: 8 (lig), PC (GPN).

Trapeliopsis flexuosa (Fr.) Coppins & P.James: 1 (lig).

Tuckermannopsis chlorophylla (Willd.) Hale: 1, 8 (cor, lig Bet, Pic, Pin).

Usnea hirta (L.) Weber ex F.H.Wigg.: 1 (lig).

!!*Verrucaria tectorum* (A.Massal.) Körb.: 9 (cal), DH (LECB).

!*Verrucaria xyloxena* Norman: 8 (ter), AS (TU).

Violella fucata (Stirt.) T.Sprib.: 1, 4, 5, 7 (cor Abi, Fag, Pin).

Vulpicida pinastri (Scop.) J.-E.Mattsson & M.J.Lai: 1 (lig).

Xanthocarpia lactea (A.Massal.) A.Massal.: 9 (sil).

Xanthomendoza fallax (Hepp) Søchting, Kärnefelt & S.Y.Kondr.: 10 (cor Aes).

Xanthomendoza fulva (Hoffm.) Søchting, Kärnefelt & S.Y.Kondr.: 8, 10 (cor Aes, Qur).

Xanthoparmelia conspersa (Ach.) Hale: 1, 8 (sil).

Xanthoria parietina (L.) Th.Fr.: 1–10 (cal, cor, lig, sil).

!*Xylographa parallela* (Ach.) Fr.: 7 (lig), AS (TU).

!*Zwackhia soređiifera* (P.James) Ertz: 5 (cor Abi, Qur), MK (UGDA), DH & IS (LECB), JM (BILAS).

Zwackhia viridis (Ach.) Poetsch & Schied.: 8 (cor Pop).

Non lichenized, lichenicolous taxa

!*Arthonia epiphyscia* Nyl.: 7 (on *Physcia aipolia*), AS (TU).

!*Arthonia molendoi* (Heufl. ex Frauenf.) R.Sant.: 2, 4, 7 (on *Xanthoria parietina*), AS (TU), MK (UGDA), IS (LECB), JM (BILAS).

Briancoppinsia cytospora (Vouaux) Diederich, Ertz, Lawrey & van den Boom: 1 (on *Evernia prunastri*).

!*Corticifraga fuckelii* (Rehm) D.Hawksw. & R.Sant.: 7 (on *Peltigera neckeri*), AS (TU), MK (UGDA), PC (GPN).

!*Didymocyrtis epiphyscia* Ertz & Diederich s. lat.: 7 (on *Xanthoria parietina*), MK (UGDA).

! *Didymocyrtis ramalinae* (Roberge ex Desm.) Ertz, Diederich & Hafellner: 8 (on *Ramalina fraxinea*), MK (UGDA).

!!*Ellisembia lichenicola* Heuchert & U.Braun: 7 (on *Ramalina fraxinea*), MK (UGDA, dupl. in BILAS).

!*Epicladonia sandstedei* (Zopf) D.Hawksw.: 5 (on *Cladonia coniocraea*), JM (BILAS).

!*Erythrimum aurantiacum* (Lasch) D.Hawksw. & A.Henrici: 1, 4 (on *Physcia* spp.), MK (UGDA), JM (BILAS).

!*Graphium aphthosae* Alstrup & D.Hawksw.: 7 (on *Peltigera neckeri*), MK (UGDA).

!*Homostegia piggotii* (Berk. & Broome) P.Karst.: 4 (on *Parmelia submontana*), DH (LECB).

Illosporiopsis christiansenii (B.L.Brady & D.Hawksw.) D.Hawksw.: 1, 4, 7 (on *Physcia* spp., *Xanthoria parietina*, unidentified crustose lichen).

!*Lichenochora obscuroides* (Linds.) Triebel & Rambold: 4 (on *Phaeophyscia orbicularis*), MK (UGDA).

!!*Lichenochora weillii* (Werner) Hafellner & R.Sant.: 1 (on *Physconia enteroxantha*), MK (UGDA).

Lichenconium erodens M.S.Christ. & D.Hawksw.: 1, 8 (on *Evernia prunastri*, *Ramalina fraxinea*).

- ! *Lichenocodium lecanorae* (Jaap) D.Hawksw.: 10 (on *Evernia prunastri*), DH (BILAS).
- ! *Lichenocodium pyxidatae* (Oudem.) Petr. & Syd.: 7 (on *Cladonia* aff. *chlorophaea*), JM (BILAS).
- ! *Lichenocodium usneae* (Anzi) D.Hawksw.: 1, 8 (on *Evernia prunastri*), JM (BILAS).
- ! *Lichenocodium xanthoriae* M.S.Christ.: 7 (on *Xanthoria parietina*), DH (BILAS).
- ! *Lichenodiplis lecanorae* (Vouaux) Dyko & D.Hawksw.: 7 (on *Lecanora* aff. *hagenii*, *Caloplaca* sp.), AS (TU), MK (UGDA).
- Lichenosticta alcicornaria* (Linds.) D.Hawksw.: 5 (on *Cladonia coniocraea*, *C. ochrochlora*), AS (TU).
- ! *Marchandiomyces corallinus* (Roberge) Diederich & D.Hawksw.: 1 (on *Physcia tenella*), PC (GPN).
- Monodictys epilepraria* Kukwa & Diederich: 5, 8 (on *Lepraria* spp.).
- ! *Phaeopyxis punctum* (A.Massal.) Rambold, Triebel & Coppins: 8 (on *Cladonia digitata*, *C. macilenta*), AS (TU), JM (BILAS).
- Pronectria xanthoriae* Lowen & Diederich: 4, 7 (on *Xanthoria parietina*).
- ! *Pyrenochaeta xanthoriae* Diederich: 4, 7 (on *Xanthoria parietina*), AS (TU), MK (UGDA).
- !! *Reconditella physconiarum* Hafellner & Matzer: 4 (on *Physconia distorta*), MK (UGDA).
- ! *Refractohilum peltigerae* (Keissl.) D.Hawksw.: 7 (on *Peltigera* spp.), AS (TU), MK (UGDA).
- ! *Sphinctrina turbinata* (Pers.: Fr.) De Not.: 9 (on *Pertusaria pertusa*), DH (LECB).
- ! *Stigmidium microspilum* (Körb.) D.Hawksw.: 8 (on *Graphis scripta*), AS (TU).
- Szygospora physciacearum* Diederich: 1, 4, 7 (on *Physcia* spp.).
- ! *Taeniolella punctata* M.S.Christ. & D.Hawksw.: 3 (on *Graphis scripta*), MK (UGDA).
- ! *Teloggalla olivieri* (Vouaux) Nik.Hoffm. & Hafellner: 1, 2, 3, 4, 7 (on *Xanthoria parietina*), AS (TU), MK (UGDA), JM (BILAS).
- ! *Thelocarpon epibolum* Nyl. var. *epibolum*: 7 (on *Peltigera neckeri*), MK (UGDA).
- Tremella hypogymniae* Diederich & M.S.Christ.: 8 (on *Hypogymnia physodes*).
- ! *Tremella lichenicola* Diederich: 5 (on *Violella fucata*), DH (LECB), PC (GPN).
- ! *Tremella phaeophysciae* Diederich & M.S.Christ.: 4 (on *Phaeophyscia orbicularis*), MK (UGDA).
- ! *Trichonectria anisospora* (Lowen) van den Boom & Diederich: 8 (on *Hypogymnia physodes*), JM (BILAS).
- ! *Trichonectria rubefaciens* (Ellis & Everh.) Diederich & Schroers: 8 (on *Parmelia sulcata*), JM (BILAS).
- Vouauxiella lichenicola* (Linds.) Petr. & Syd.: 1, 5 (on *Lecanora chlorotera*, *L. pulicaris*).
- ! *Vouauxiomyces santessonii* D.Hawksw.: 8 (on *Platismatia glauca*), JM (BILAS).
- Xanthoriicola physciae* (Kalchbr.) D.Hawksw.: 7 (on *Xanthoria parietina*).
- ! *Xenonectriella leptaleae* (J.Steiner) Rossman & Lowen: 4 (on *Physcia aipolia*), MK (UGDA).

Non lichenized, saprobic taxa

- Arthonia punctiformis* Ach.: 8 (cor Qur).
- ! *Chaenothecopsis viridireagens* (Nádv.) A.F.W.Schmidt: 8 (lig), LK (LE).
- ! *Leptorhaphis epidermidis* (Ach.) Th.Fr.: 4 (cor Bet), DH (LECB).

Mycocalicium subtile (Pers.) Szatala: 1, 4, 5, 7 (lig).

! *Peridiothelia fuliguncta* (Norman) D.Hawksw.: 1 (cor Til), DH (LECB).

! *Sarea difformis* (Fr.) Fr.: 3 (res Pic), DH (LECB)

Sarea resinae (Fr.: Fr.) Kuntze: 3 (res Pic).

Stenocybe pullatula (Ach.) Stein: 3 (cor Aln).

!! *Stictis brunnescens* Gilenstram, Döring & Wedin: 4 (lig), IS (LECB).

Discussion

In total, 290 species are reported from the visited localities in Kurzeme, which constitutes 49% of lichen flora previously known from Latvia (591 species according to PITERĀNS [2001] and various later publications, see Introduction). Among the listed species, 238 are lichenized, 43 lichen-habiting and 9 saprotrophic, non-lichenized fungi. There are 91 species new to Latvia, 12 of which (*Caloplaca duplicata*, *Cresporhaphis wienkampii*, *Ellisembia lichenicola*, *Gallowayella weberi*, *Gregorella humida*, *Lichenochora weillii*, *Parmelia serrana*, *Polycauliona phlogina*, *Reconditella physconiarum*, *Stictis brunnescens*, *Thelocarpon superellum* and *Verrucaria tectorum*) are also new for the Baltic States, and *Athallia alnetorum* is reported here for the first time in Northern Europe.

The large number of new records reflects in part the fact that Latvian lichen flora is understudied, when compared to neighbouring Lithuania and Estonia, a large portion of the species known for some time from these two Baltic States, namely *Anisomeridium polypori*, *Bacidia fraxinea*, *Chaenotheca hispidula*, *Chrysothrix flavovirens*, *Epicladonia sandstedei*, *Leptogium byssinum*, *Graphium aphthosae*, *Gyalecta flotowii*, *Leptorhaphis epidermidis*, *Lichenochora obscuroides*, *Lichenocodium lecanorae*, *Lichenocodium pyxidatae*, *Lichenocodium usneae*, *Lichenocodium xanthoriae*, *Lichenodiplis lecanorae*, *Ochrolechia microstictoides*, *Parmelia submontana*, *Phaeophyscia endophoenicea*, *Psilolechia clavulifera*, *Protothelenella sphinctrinoidella*, *Pyrenochaeta xanthoriae*, *Refractohilum peltigerae*, *Rinodina sophodes*, *Sarea difformis*, *Schismatomma pericleum*, *Scoliciosporum sarothamni*, *Steinia geophana*, *Taeniolella punctata*, *Thelocarpon epibolum* var. *epibolum*, *Tremella lichenicola*, *Tremella phaeophysciae*, *Trichonectria anisospora*, *Verrucaria xyloxena*, *Vouauxiomyces santessonii*, *Xylographa parallela* and *Zwackhia sorediifera*. Some of these species are rather rare in Lithuania and Estonia, but others are quite common in both countries, such as *Anisomeridium polypori*, *Leptorhaphis epidermidis*, *Lichenochora obscuroides*, *Lichenocodium* spp., *Ochrolechia microstictoides*, *Steinia geophana*, *Tremella lichenicola*, *T. phaeophysciae* and *Vouauxiomyces santessonii* (J. Motiejūnaitė and A. Suija, pers. comm.). On the other hand, several new records for Latvia are species that were only recently reported from Lithuania and Estonia, namely *Arthonia molendoi*, *Lecanora farinaria*, *Micarea byssacea*, *M. micrococca*, *Trichonectria rubefaciens*, *Xenonectriella leptaleae* (SUIJA et al. 2009, 2010, CZARNOŃ & GUZOW-KRZEMIŃSKA 2010, MOTIEJŪNAITĖ et al. 2010, 2012, MOTIEJŪNAITĖ 2011, LÖHMUS et al. 2012, SUIJA 2013) and several species were previously known from only Lithuania or Estonia: *Alyxoria culmigena*, *Arthonia epiphyscia*, *Caloplaca ulcerosa*, *Chaenothecopsis viridireagens*, *Didymocyrtis epiphyscia*, *Eopyrenula leucoplaca*, *Homostegia piggotii*, *Lecanora subcarpineae*, *Marchandiomyces corallinus*, *Micarea anterior*, *Rinodina septentrionalis*, *Solitaria chrysophthalma*, *Sphinctrina turbinata* and *Teloggalla olivieri* from Estonia (RANDLANE & SAAG 1999, SUIJA et al. 2001, 2009, JÜRIADO et al. 2003, APTROOT et al. 2005, SUIJA 2005), and *Bacidina delicata*, *Biatora meiocarpa*, *Candelariella efflorescens*, *Didymocyrtis ramali-*

nae, *Gyalecta derivata*, *Peridiothelia fuliguncta*, *Strigula jamesii* and *Trapelia corticola* from Lithuania (MOTIEJŪNAITĒ 1999, 2005, 2007, MOTIEJŪNAITĒ et al. 1998, 2008, 2011, 2012).

Six protected species in Latvia, of 34 red-listed lichens (PITERĀNS 2001), were found, namely *Lobaria pulmonaria*, *Pertusaria flavida*, *P. pertusa*, *Pleurosticta acetabulum*, *Thelotrema lepadinum* and *Zwackhia viridis* all of which were recorded in the old-growth forest in Slītere National Park (locality 8); *Pleurosticta acetabulum* was also found near Šķēde Forest Research Station (locality 1), and *Lobaria pulmonaria*, *Pertusaria flavida* and *P. pertusa* also grew in an old park of Dundaga castle (locality 9).

The concerted efforts during the rather short field trip resulted in a high number of records new to Latvia, as well as extensions of known distributions of some species to the eastern Baltic region. Furthermore, presence of certain taxa, previously recorded under species complexes, was clarified. *Ochrolechia androgyna*, known in Latvia since the 1970s (PITERĀNS 1982), was revised according to KUKWA (2009), confirming the presence of *O. androgyna* s. str., and the *Caloplaca holocarpa* complex, known in Latvia first under the name of *C. pyracea* (Ach.) Th.Fr. (PITERĀNS 1982) and later as *C. holocarpa* (Hoffm.) A.E.Wade, was recorded from both bark and stones (PITERĀNS 2001); however, based on the revision by ARUP (2009) two clearly defined entities, belonging to a phylogenetically well-defined group, have been segregated into a newly described genus *Athallia* (ARUP et al. 2013), both of which, *A. holocarpa* and *A. pyracea*, were found in Kurzeme.

Notes on the species new to all Baltic States

Athallia alnetorum is very similar to *Gyalolechia flavorubescens* (Huds.) Söchting, Frödén & Arup differing from the latter in its smaller ascospores. Previously it was known only from the Alps, Southern Europe and Turkey, where it is quite common (GIRALT et al. 1992, VONDRÁK & WIRTH 2013, VONDRÁK et al. 2015). This is the first record for the Baltic region and Northern Europe.

Caloplaca duplicata is characterized by its grey blastidiate thallus and zeorine apothecia with orange disk and proper exciple. It was known only from the type locality in Leningrad region of Russia (TUR-V 07513) (VAINIO 1878, KHODOSOVTSSEV et al. 2004). The status of the taxon has not been fully clarified, but according to molecular data obtained from the type specimen it is very close to *C. atroflava* (Turner) Mong. and *C. soralifera* Vondrák & Hrouzek (REDCHENKO et al. 2012, J. Vondrák, pers. comm.).

Cresporhaphis wienkampii is a rare, probably only facultatively lichenized species confined to broad-leaved trees with rough bark, and scattered throughout Europe (APTROOT 1997, AGUIRRE-HUDSON 2009).

Ellisembia lichenicola is a lichenicolous hyphomycete described from *Physconia distorta*, but also found on other hosts. In the protologue it is characterized as parasitic on lichens, but no infection characteristics are given; furthermore, in all mentioned specimens it was associated with other lichenicolous fungi (HEUCHERT & BRAUN 2006). In our specimen, *Ellisembia lichenicola* was found on blackened, sclerotized parts of its host (*Ramalina fraxinea*), growing together with an unidentified hyphomycete. *Ellisembia lichenicola* was previously known from Denmark, Germany, Italy and Canada (BRACKEL 2014).

Gallowayella weberi can be confused with four other xanthorioid species, *G. fulva* (Hoffm.) S.Y.Kondr. et al., *G. sogdiana* (S.Y.Kondr. & Kärnefelt) S.Y.Kondr. et al., *Honeggeria ros-marieae* (S.Y.Kondr. & Kärnefelt) S.Y.Kondr. et al. and *Oxneria huculica* S.Y.Kondr. It is dis-

tinguished from *G. fulva* by its longer, more branched lobes, and from *O. huculica* by its narrower lobes, a more radiating growth, and lack of large helmet-shaped soralia (KONDRATYUK & KÄRNEFELT 2003, KONDRATYUK et al. 2010), and from the central Asian species *G. sogdiana* by its smaller thallus and thicker ascospore septa, and from the distantly related North American species *Honeggeria rosmariaeae* by a *Jackelixia*-type of true exciple, a higher hymenium and narrower ascospores (KONDRATYUK et al. 2010, LUMBSCH et al. 2011). The known distribution of *Gallowayella weberi* is in northern and central Europe (Germany, Norway, Sweden & Russia) and North America (Canada, USA & Mexico). It typically grows on deciduous roadside trees in moderately polluted settlements.

Gregorella humida, a pioneer species of recently disturbed clayey soils, is scattered and rare in most countries where recorded, but apparently spreading in Central Europe (VONDRÁK et al. 2013). When ill-developed or sterile, *G. humida* can be confused with *Moelleropsis nebulosa* (Hoffm.) Gyeln. or *Leptogium byssinum*.

Lichenochora weillii differs from other *Lichenochora* species confined to Physciaceae by hyaline 2-celled broadly ellipsoid, $10\text{--}12 \times 8\text{--}9.5\mu\text{m}$ ascospores (HAFELLNER 1989). The hosts of the fungus – *Physconia* spp. – occur commonly in the eastern Baltic region.

Parmelia serrana is one of several species discovered through molecular analyses of DNA-sequences in the *P. saxatilis* group and is chemically distinct from *P. saxatilis* s. str. in containing lichesterinic, not lobaric, acid (THELL et al. 2008, OSSOWSKA et al. 2014). Usually it is recognizable morphologically by its rather orbicular thalli with short and broad, often overlapping, lobes with rounded tips, but the longitudinal cracks and ridges on the upper side of the thallus and the clustered isidia, often concentrated to the ridges and the margins, may also occur on *P. saxatilis*. *P. submontana*, occasionally producing isidioid soredia, may be confused with *P. serrana*, but the former has elongated lobes with the margins folded downwards (OSSOWSKA et al. 2014). A specimen recently collected in Sweden, originally determined as *P. serrana* (THELL et al. 2014), with such isidia, but partly with sublinear lobes, was re-determined as *P. saxatilis* after chemical analysis (Ossowska, pers. comm.). It is expected that additional cryptic or semi-cryptic species will be described within the *P. saxatilis* complex (OSSOWSKA et al. 2014). *P. serrana* was believed to have a restricted distribution area (MOLINA et al. 2004), being recorded from southern, central and western Europe and the Canary Islands (THELL et al. 2011).

Polycauliona phlogina is characterized by its mainly sorediate thallus with only occasional small areoles (ARUP 2005). It occurs throughout Europe and is quite common in the southern parts of Scandinavia (ARUP 2005); it is also known from the Leningrad region of Russia (VONDRÁK et al. 2010).

Reconditella physconiarum is characterized by superficial perithecia and one-celled verruculose, $15\text{--}21 \times 6\text{--}10\mu\text{m}$ ascospores (MATZER & HAFELLNER 1990). As with *Lichenochora weillii*, the hosts of the fungus – *Physconia* spp. – are common in eastern Baltic region.

Stictis brunnescens is similar to *S. radiata* Pers., from which differs in the structure of its periphysoid layer, presence of marginal paraphyses and brown pigment in the exciple (WEDIN et al. 2006). It was recently reported from the Leningrad region of Russia (KUZNETSOVA et al. 2012).

Thelocarpon superellum is one of the rarely recorded *Thelocarpon* species, characterized by its rather large ascomata, presence of simple paraphyses, I+ blue asci, and spores which commonly have a pseudoseptum (KOCOURKOVÁ-HORÁKOVÁ 1998).

Verrucaria tectorum is similar to the very common *V. nigrescens* except for the presence of soredia (KRZEWICKA 2012) and is therefore sometimes treated as a morph of the latter species (COPPINS & APTROOT 2008).

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III

Lichens and allied fungi new for Latvia

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Abstract: Six lichenized fungi (*Calicium parvum*, *C. trabinellum*, *Carbonicola anthracophila*, *C. myrmecina*, *Peltigera extenuata*, *Pycnora sorophora*), two lichenicolous (*Clypeococcum hypocenomyces* and *Illosporium carneum*), and two saprobic calicioid fungi (*Chaenothecopsis savonica* and *Microcalicium arenarium*) are reported as new for Latvia.

Keywords: Latvia; lichenicolous; calicioid; saprobic fungi

INTRODUCTION

Several essential contributions for Latvian lichen biota have been presented recently (Āboliņa et al., 2015; Jurciņš et al., 2014; Moisejevs, 2015; Motiejūnaitė et al., 2016). These contributions added over one hundred species of lichens and allied fungi to the Latvian checklist, including re-find of *Lobaria amplissima*, a species which was considered to be extinct in the country. Currently about 690 species of lichenized and allied fungi are recorded in Latvia. This paper further supplements the knowledge on lichens and allied fungi in Latvia.

MATERIAL AND METHODS

All referred specimens were collected by the author in 2015–2016 in different regions of Latvia, during local inventories of protected habitats, according to Latvian interpretation of Annex I EU Habitat Directive in (Auniņš et al., 2013). The collections were determined using the routine lichenological methods (Smith et al., 2009). Specimens were examined under stereomicroscope Nikon SMZ 800. Water mounted hand-made cross sections were examined under light transmission microscope Nikon Eclipse E100. The nomenclature of taxa mainly follows Nordin et al. (2011). Doubtful specimens were compared with the collections at the herbarium of the Institute of Botany, Nature Research Centre, Vilnius (BILAS). For each specimen the protected habitat type and number according to Annex I EU Habitat Directive is added in brackets, if relevant. Species distribution data in neighboring countries (for Russia only Pskov region) are provided. All reference specimens are deposited at the Herbarium of Daugavpils

University, Institute of Life Sciences and Technology, Laboratory of Botany (DAU).

In the species list non-lichenized saprobic fungi are marked with + and lichenicolous fungi with #.

THE SPECIES

CALICIUM PARVUM Tibell – Krāslavas County, Kaplavas District, about 250–300 m off Šilovkas lake center, 55°49'33.32"N, 27°14'37.71"E; in dry old-growth *Pinus sylvestris* forest site, burned in past (evaluated as 9010* Western Taiga), on *Pinus sylvestris* bark; 24 August 2016 (DAU 5018009001).

Note. The species is known in Belarus (A. Tsurykau pers. comm.), Estonia (Löhmus, 1998) and Lithuania (Motiejūnaitė, 1999).

CALICIUM TRABINELLUM (Ach.) Ach. – Krāslavas Co., Ūdrīšu District, Tartaks village, about 600 m E off Tartaks village, 55°52'54.454"N, 26°59'35.746"E; in boggy *Pinus sylvestris* forest site with *Ledum palustre* L. (evaluated as 91D0* Bog woodland), on wood of pine snag; 10 September 2015; DAU518008001; Daugavpils Co., Dvietes District., Kinkausk forest area, 56°8'30.884"N, 26°14'28.676"E; in young dry *Pinus sylvestris* forest, on wood of pine stump; 12 July 2016; DAU518008002.

Note. Known in Belarus (Golubkov & Titov, 1990), Estonia (Löhmus, 1998), Lithuania (Motiejūnaitė, 1999) and Russia (Istomina & Likhacheva, 2014). In Lithuania *C. trabinellum* is treated as an indicator species of undisturbed boggy pine forests (Motiejūnaitė et al., 2004).

CARBONICOLA ANTHRACOPHILA (Nyl.) Bendiksby & Timdal – Gulbenes Co., Stradu District, highway P36, turn to Lazdaga and Kaļņa lakes, about 200–300 m from the highway, 57°3'12.551"N, 26°55'27.059"E; in old-growth dry *Pinus sylvestris* forests (evaluated as 9010* Western taiga), on *Pinus sylvestris* bark; 2 August 2015; DAU552001001; Krāslavas Co., Kaplavas District, Šilovkas lake, about 250–300 m from Šilovka lake central part, 55°49'33.32"N 27°14'37.71"E; in old-growth dry *Pinus sylvestris* forests (evaluated as 9010* Western taiga), on *Pinus sylvestris* bark; 24 August 2016; DAU552001002; Alūksnes Co., Liepnas District, about 2 km to SE from Liepna city, 57°20'29.798"N; 21°39'35.462"E; in old-growth moistured *Pinus sylvestris* forest, on burned pine snag, on wood; 26 September 2016; DAU552001003.

Note. The species is known in Belarus (Yatsyna, 2012), Estonia (Randlane et al., 1999) and Lithuania (Motiejūnaitė, 1999). In all neighboring countries, the species is mainly found in old-growth, previously burned forest sites. In Latvia, in all cases the lichen was found in dry old-growth pine stands impacted by past forest fires.

CARBONICOLA MYRMECINA (Ach.) Bendiksby & Timdal – Alūksnes Co., Liepnas District, about 2 km to SE from Liepna city, 57°20'29.798"N; 21°39'35.462"E; in old-growth moistured *Pinus sylvestris* forest, on burned pine snag, on wood; 26 September 2016; DAU552002001.

Note. Known in Estonia (Randlane et al., 1999) and Lithuania (Motiejūnaitė, 1999).

+CHAENOTHECOPSIS SAVONICA (Räsänen) Tibell – Ventspils Co., Užavas District, Užava town, Nature Preserve Užava, about 600 m SE off Užava lighthouse, 57°12'17.798"N, 21°25'37.462"E; in old-growth birch-spruce-pine forest site (evaluated as 9010* Western taiga), on *Picea abies* snag; 10 September 2015; DAU527005001.

Note. *C. savonica* is known in Estonia (Löhmus, 1998) and Lithuania (Motiejūnaitė, 1999).

#CLYPEOCOCCUM HYPOCENOMYCIS D. Hawksw. – Krāslavas Co., Ūdrīšu District, Borovka village, about 300 m N off the farmstead „Ciemati”, 55°55'9.289"N, 26°59'23.060"E; in a middle-aged *Pinus sylvestris* forest stand, on thalli of

Hypocenomyce scalaris growing on *Pinus sylvestris*; 14 September 2015; DAU646001001; Krustpils Co., Atašienes District, About 700 m S off Teiču Nature Reserve observation tower, 56°31'28.203"N, 26°34'2.132"E; in boggy *P. sylvestris* forest with *Ledum palustre* L., (evaluated as 91D0* Bog woodland), on thalli of *Hypocenomyce scalaris* growing on *Pinus sylvestris*; 17 March 2016; DAU646001002.

Note. It is a common lichenicolous fungus, known in Belarus (Tsurykau & Khramchankova, 2011), Estonia (Randlane et al., 2015) and Lithuania (Motiejūnaitė, 1999).

#ILLOSPORIUM CARNEUM Fr. – Krustpils Co., Atašienes District, about 500 m NE off Teiču Nature reserve observation tower, 56°32'54.332"N, 26°25'25.403"E; on the railroad earthwork, on thalli of *Peltigera extenuata* growing on soil (it was in the state of an anamorph); 26 April 2016; DAU647001001.

Note. *Illosporium carneum* is one of the commonest peltigericolous fungi, it is known in Estonia (Suija, 2005) and Lithuania (Motiejūnaitė, 1999).

+MICROCALICIUM ARENARIUM (Hampe ex A. Massal.) Tibell – Ventspils Co., Užavas District, Užava town, Užava Nature Reserve, about 700 m SE off Užava lighthouse, 57°11'7.195"N, 21°25'16.528"E; in old-growth birch-spruce-pine forest site (evaluated as 9010* Western taiga), on a sand layer in *Picea abies* root plate; 10 Sept 2015; DAU643003001.

Note. Known in Belarus (Golubkov, 1987), Estonia (Löhmus, 1998) and Lithuania (Motiejūnaitė et al., 2013). In all neighboring countries the species is known only from very few localities. In Estonia it has been assessed as vulnerable (Löhmus & Löhmus, 2011). In Lithuania the species was known from literature data and only recently was recorded in one locality (Motiejūnaitė et al., 2013). In Belarus the species is known from a single locality, Belovezhskaya Pushcha National Park, and is considered as very rare (Golubkov, 1987). Regional distribution and sparse occurrence of *M. arenarium* is a good basis to include it to local red lists.

PELTIGERA EXTENUATA (Vainio) Lojka – Krustpils Co., Atašienes District, about 500 m NE off Teiču Nature Reserve observation tower,

56°32'54.332"N, 26°25'25.403"E; on the railroad earthwork, on soil; 26 April 2016; DAU647001001.

Note. The species is recorded in all neighboring countries (Randlane et al., 2015; Motiejūnaitė, 1999).

PYCNORA SOROPHORA (Vain.) Hafellner – Krustpils Co., Atašienes District, Nature preserve “Great Pelečāre bog”, about 300 m to NW from “Bernāni” farmstead, 56°28'22.8"N 26°34'04.9"E; in boggy *Pinus sylvestris* forest with *Ledum palustre* L. (evaluated as 91D0* Bog woodland), on bark of *Pinus sylvestris*; 25 June 2016; DAU 648001001; Ventspils Co., Kolkas District., Slītere National Park, about 300 m to S from Kolka Cape, 57°45'11.73"N 22°35'56.62"E; in dry *Pinus sylvestris* forest (evaluated as 2180 Wooded dunes of the Atlantic, Continental and Boreal region), on bark of *Pinus sylvestris*; 24 July 2016; DAU 648001002.

Note. The species is known in Belarus (Tsurykau et al., 2012), Estonia (Randlane et al., 2015) and Lithuania (Motiejūnaitė, 1999).

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IV

FOUR SPECIES OF SAXICOLOUS LICHENIZED FUNGI NEW TO LATVIA

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Abstract

Moisejevs R., Degtjarenko P., 2017: Four species of saxicolous lichenized fungi new to Latvia [Keturių epilitinių kerpių rūšys naujos Latvijai]. – Bot. Lith., 23(1): 60–62.

Four species of saxicolous and acidophilous lichens – *Dermatocarpon miniatum*, *Trapelia coarctata*, *Trapelia placodioides*, and *Umbilicaria hirsuta* found on different types of granite boulders were reported as new to Latvian lichen biota. Data on substratum geology, accompanying species, microhabitat and distribution in neighbouring to Latvia countries are provided.

Keywords: acidophilous, epilithic, lichens.

In recent years, knowledge about lichens and allied fungi in Latvia has significantly increased (PITERĀNS, 2001; ĀBOLIŅA et al., 2015; MOTIEJŪNAITE et al., 2016). Nevertheless, the saxicolous lichens still remain one of the least studied lichenized fungi group in the Baltics (ĀBOLIŅA et al., 2015; MOTIEJŪNAITE et al., 2016). The current paper complements the knowledge about Latvian biota of lichenized fungi. Four species of saxicolous lichenized fungi were reported as new to Latvia. The species were found in 2016–2017, on rapakivi granite boulders in different regions of Latvia. All of these boulders have probably been transported to Latvia from the southwestern part of Finland during the last glaciation, i.e. Weichselian glaciation (ZELČS & MARKOTS, 2004).

The collections were determined using the routine lichenological methods (SMITH et al., 2009). Spot-test reactions were checked with 10% KOH (K), sodium hypochlorite (C), para-phenylenediamine in ethanol (Pd), and Lugol's solution (I). The nomenclature of taxa follows NORDIN et al. (2011). Doubtful specimens were compared to the collections deposited at the Herbarium of the Institute of Botany, Nature Research Centre, Vilnius (BILAS). The description of

substrata and list of accompanying lichen species is provided. The data on species distribution in neighbouring countries is also provided, except Pskov region for Russia, due to unconfirmed occurrence of all referred species in the region (Irina Stepanchikova, pers.comm). All collected specimens were deposited at the Herbarium of Daugavpils University, Institute of Life Sciences and Technology (DAU).

LIST OF SPECIES:

Dermatocarpon miniatum (L.) W.Mann.

The lichen inhabited well-lit surface of boulder top, on south-facing part. The accompanying species were: *Acarospora fuscata*, *Aspicilia cinerea*, *Prototermeliopsis muralis*, *Rhizocarpon geminatum*, *Phaeophyscia nigricans*, *Phaeophyscia sciastra*, and *Xanthoparmelia conspersa*. *Dermatocarpon miniatum* is known in Belarus (TSURYKAU & KHRAMCHANKOVA, 2011) and Estonia (RANDLANE et al., 1999). So far, this species is only known from literature data in Lithuania (MOTIEJŪNAITE, 1999). *D. miniatum* is included into the Red Data Book of the Republic of Belarus (KHORUZHNIK, 2005).

Specimen examined: Ogres Co., Lauberes Distr., 200 m to the west of Ančiņu House, Ančiņu devil's stone; on the upper surface of exposed granite boulder, rapakivi granite with multiple cracks. Metamorphic rock – coarse-grained gneiss (metamorphosed granite) with dominance of orthoclase feldspars. Leg. et det. Rolands Moisejevs, 24 April 2017; 56°50'36.3" N, 25°03'07.1" E. Ančiņu devil's stone; DAU (438002001).

***Trapelia coarctata* (Sm.) M.Choisy**

The lichen was found on the north-east-facing side of the boulder, on the bottom and central part. The microhabitat can be characterized as shaded with scabrous surface. The accompanying species were: *Candelariella aurella*, *Porpidia crustulata*, *Porpidia macrocarpa*, *Rhizocarpon geminatum*, and *Trapelia placodioides*. *Trapelia coarctata* is known in Lithuania (MOTIEJŪNAITĒ, 1999), Belarus (TSURYKAU & KHRAMCHANKOVA, 2011), and Estonia (RANDLANE et al., 1999).

Specimen examined: Daugavpils Co., Nīcgales Distr., protected landscape area "Nīcgales meži", included into NATURA 2000 protected territory net, 6 km to the east of railway station, Nīcgale Big White Stone; on a well-lit granite boulder; rapakivi granite. Intrusive igneous rock – coarse-grained granite with porphyritic texture. Felsic rock with dominance of plagioclase feldspars. Leg. Rolands Moisejevs, 12 August 2016; 56°09'07.9" N, 26°27'41.9" E. Det. Jurga Motiejūnaitē, 27 August 2016; DAU (546001001).

***Trapelia placodioides* Coppins & P.James**

The lichen grew on the north-east-facing side of the boulder, on the bottom and central part. The microhabitat can be characterized as shaded with scabrous surface. The accompanying species were: *Candelariella aurella*, *Porpidia crustulata*, *Porpidia macrocarpa*, *Rhizocarpon geminatum*, and *Trapelia coarctata*. *Trapelia placodioides* is known in Lithuania (MOTIEJŪNAITĒ, 1999) and Estonia (RANDLANE et al., 1999).

Specimen examined: Daugavpils Co., Nīcgales Distr., protected landscape area "Nīcgales meži", included into NATURA 2000 protected territory net, 6 km to the east of railway station, Nīcgale Big White Stone; on a granite boulder; rapakivi granite. Intrusive igneous rock – coarse-grained granite with

porphyritic texture. Felsic rock with dominance of plagioclase feldspars. Leg. et det. Rolands Moisejevs, 12 August 2016; 56°09'07.9" N, 26°27'41.9" E. DAU (546002001).

***Umbilicaria hirsuta* (Sw. ex Westr.) Hoffm.**

The species was found on the top and east-facing part of boulder surface. The surface of boulder can be described as scabrous with numerous cracks. The accompanying species were: *Acarospora fuscata*, *Aspicilia cinerea*, *Candelariella aurella*, *Protoparmeliopsis muralis*, and *Xanthoparmelia conspersa*. *U. hirsuta* has so far been reported only from Estonia, where it has recently been recorded in the northern part of the country (RANDLANE et al., 2014).

Specimen examined: Ventspils Co., Ugāles Distr., 100 m to the south of the Māteri Cemetery. The Devil's Foot Stone in Māteri is located on top of a Liepkalns Hill, in a clear meadow; on well-lit horizontal surface of a granite boulder; rapakivi granite. Intrusive igneous rock – coarse-grained granite. Felsic rock with dominance of orthoclase feldspars. Leg. Rolands Moisejevs, 12 February 2017; 57°14'35" N, 21°57'47" E. Det. Polina Degtjarenko, 13 February 2017; DAU (635003001).

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KETURIOS EPILITINIŲ KERPIŲ RŪŠYS NAUJOS LATVIJAI

Rolands MOISEJEVS, Polina DEGTJARENKO

Santrauka

Keturios epilitinės acidofilinės kerpių rūšys – *Dermatocarpon miniatum*, *Trapelia coarctata*, *Trapelia placodioides* ir *Umbilicaria hirsuta* aptiktos ant įvairių granito riedulių buvo rastos pirmą kartą

Latvijoje. Straipsnyje pateikiami duomenys apie jų substrato geologiją, mikrobuveinės savybes ir rūšių paplitimą aplinkinėse šalyse.

V



New records of lichens and lichenicolous fungi from Latvia, with a list of lichenicolous fungi reported from Latvia

Authors: Moisejevs, Rolands, Degtjarenko, P., Motiejūnaitė, J., Piterāns, A., and Stepanova, D.

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New records of lichens and lichenicolous fungi from Latvia, with a list of lichenicolous fungi reported from Latvia

Rolands Moisejevs, P. Degtjarenko, J. Motiejūnaitė, A. Piterāns and D. Stepanova

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Four species of lichen-forming fungi (*Calicium pinastri*, *Chaenotheca laevigata*, *Lecania croatica* and *Pycnora praestabilis*) and two lichenicolous fungi (*Arthrorhaphis aeruginosa* and *Chaenothecopsis epithallina*) are reported as new for Latvia. The first comprehensive list of lichenicolous fungi in Latvia is also presented, including their hosts and distribution in Latvia (northern Europe).

Keywords: Baltic countries, distribution, lichenized fungi

Our knowledge of the lichens and allied fungi in Latvia (northern Europe) has been considerably advanced in recent years. Currently ca 640 taxa of lichenized fungi are recorded for Latvia (Āboliņa et al 2015, Moisejevs 2015, 2017, Motiejūnaitė et al. 2016, Moisejevs and Degtjarenko 2017), which is comparable to the 620 taxa known from the geographically similar territory of Lithuania (Motiejūnaitė 2017).

Lichenicolous fungi, on the other hand, have been understudied in Latvia. The first mention of lichenicolous species can be found in the paper by Mereschkowski (1913), who reported *Acolium sessile* from Jaunugulbene (Vidzeme). Later, *Diploschistes muscorum* was reported from the Gauja river valley and Riga city environs (Malta 1926, Vimba 1971). During the 13th International Symposium of Lichenologists and Mycologists of the Baltic States, two more species of lichenicolous fungi were reported (Motiejūnaitė and Piterāns 1998); a further four species were reported in the first annotated Latvian checklist of lichens (Piterāns 2001) and 16 more were added by Motiejūnaitė et al. (2006), Czarnota and Kukwa (2010) and Motiejūnaitė and Grochowski (2014). The second annotated checklist of lichens listed 15 lichenicolous taxa as supplementary data (Āboliņa et al. 2015). Further contribution was made by Motiejūnaitė et al. (2016), with 33 lichenicolous fungi

reported for the first time from Latvia, complemented by the paper of Moisejevs (2017). Since a full and up-to-date list of lichenicolous fungi for Latvia is still lacking, a comprehensive list is provided here.

The current paper reports four new species of lichen-forming and two lichenicolous fungi new to Latvia, together with a list of lichenicolous fungi (61 taxa) known for Latvia, including their hosts and distribution data in the country (Table 1).

Material and methods

The material was determined by means of routine lichenological methods (Smith et al. 2009). Spot-tests were determined with 10% KOH (K), sodium hypochlorite (C), paraphenylenediamine in ethanol (Pd) and Lugol's solution (I), and secondary chemistry by thin layer chromatography (TLC) using solvent C (Orange et al. 2001). Specimens of the newly recorded species, according to the nomenclature of Wirth et al. (2010), are kept in the lichenological herbarium of University of Daugavpils (DAU).

The list of lichenicolous fungi presented below is a combination of published literature data and herbarium collections from DAU and the University of Latvia (RIG), together with those species reported in the current paper. Data on species distribution are derived from literature sources and herbarium collections (DAU and RIG). Regions of Latvia (Fig. 1) are abbreviated in the list as follows: K – Kurzeme (Kurland), V – Vidzeme, L – Latgale, R – Pierīga, Z – Zemgale, LV – all regions of Latvia; # = lichenicolous fungus.

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Table 1. The list of lichenicolous fungi known from Latvia, their distribution and hosts known in the country and the references that mention the species. For the abbreviations see Fig. 1.

| S. No. | Species | Host(s) | Distribution | Reference(s) |
|--------|--|--|--------------|---|
| 1. | <i>Acolium sessile</i> (Pers.) Arnold. | <i>Pertusaria</i> sp. | K, V | Mereschkowski 1913, Piterāns 1982, 2001, Āboliņa et al. 2015 |
| 2. | <i>Arthonia epiphyscia</i> Nyl. | <i>Physcia aipolia</i> | K | Motiejūnaitė et al. 2016 |
| 3. | <i>Arthonia parietinaria</i> Hafellner & A.Fleischhacker | <i>Xanthoria parietina</i> | K | Motiejūnaitė et al. 2016 (as <i>Arthonia molendoi</i> (Heufl. ex Frauenf.) R.Sant.) |
| 4. | <i>Arthrorhaphis aeruginosa</i> R.Sant. & Tønsberg | <i>Cladonia</i> sp. | L | This paper |
| 5. | <i>Bachmanniomyces punctum</i> (A. Massal.) Diederich & Pino-Bodas | <i>Cladonia digitata</i> , <i>C. macilenta</i> | K | Motiejūnaitė et al. 2016 |
| 6. | <i>Biatoropsis usnearum</i> Räsänen | <i>Usnea subfloridana</i> | V, K | Motiejūnaitė et al. 2006, Āboliņa et al. 2015, DAU Herbarium |
| 7. | <i>Briancoppinsia cytospora</i> (Vouaux) Diederich et al. | <i>Evernia prunastri</i> , <i>Melanelixia subaurifera</i> | K, V | Czarnota and Kukwa 2010, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |
| 8. | <i>Chaenothecopsis consociata</i> (Nádv.) A.F.W.Schmidt | <i>Chaenotheca chrysocephala</i> | LV | Piterāns 2001 |
| 9. | <i>Chaenothecopsis epithallina</i> Tibell | <i>Chanotheca trichialis</i> | K | This paper |
| 10. | <i>Chaenothecopsis pusilla</i> (A.Massal.) A.F.W.Schmidt | <i>Chaenotheca</i> sp. | LV | Piterāns 2001 |
| 11. | <i>Clypeococcum cetrariae</i> Hafellner | <i>Cetraria islandica</i> | V | Motiejūnaitė and Piterāns 1998 |
| 12. | <i>Clypeococcum hypocenomycis</i> D.Hawksw. | <i>Hypocenomyce scalaris</i> | LV | Moisejevs 2017 |
| 13. | <i>Corticifraga fuckelii</i> (Rehm) D.Hawksw. & R.Sant. | <i>Peltigera neckeri</i> | K | Motiejūnaitė et al. 2016 |
| 14. | <i>Didymocyrtis epiphyscia</i> Ertz & Diederich | <i>Xanthoria parietina</i> | K | Motiejūnaitė et al. 2016 |
| 15. | <i>Didymocyrtis pseudeverniae</i> (Etayo & Diederich) Ertz & Diederich | <i>Pseudevernia furfuracea</i> | V | Motiejūnaitė and Grochowski 2014, Āboliņa et al. 2015 |
| 16. | <i>Didymocyrtis ramalinae</i> (Roberge ex Desm.) Ertz, Diederich & Hafellner | <i>Ramalina fraxinea</i> | LV | Motiejūnaitė et al. 2016 |
| 17. | <i>Diploschistes muscorum</i> (Scop.) R.Sant. | <i>Cladonia</i> sp. | K, V | Malta 1926, Vimba 1971, Piterāns 1982, 2001 |
| 18. | <i>Ellisembia lichenicola</i> Heuchert & U.Braun | <i>Ramalina fraxinea</i> | K | Motiejūnaitė et al. 2016 |
| 19. | <i>Epicladonia sandstedei</i> (Zopf) D.Hawksw. | <i>Cladonia coniocraea</i> | K | Motiejūnaitė et al. 2016 |
| 20. | <i>Erythrimum aurantiacum</i> (Lasch) D.Hawksw. & A.Henrici | <i>Physcia</i> spp. | K | Motiejūnaitė et al. 2016 |
| 21. | <i>Graphium aphthosae</i> Alstrup & D. Hawksw. | <i>Peltigera neckeri</i> | K | Motiejūnaitė et al. 2016 |
| 22. | <i>Heterocephalacria physciacearum</i> (Diederich) Millanes & Wedin | <i>Physcia</i> spp. | K | Czarnota and Kukwa 2010, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |
| 23. | <i>Homostegia piggotii</i> (Berk. & Broome) P.Karst. | <i>Parmelia submontana</i> | K | Motiejūnaitė et al. 2016 |
| 24. | <i>Illosporopsis christiansenii</i> (B.L.Brady & D.Hawksw.) D.Hawksw. | <i>Physcia</i> spp., <i>Xanthoria parietina</i> | LV | Piterāns 2001, Czarnota and Kukwa 2010 |
| 25. | <i>Lichenochora obscuroides</i> (Linds.) Triebel & Rambold | <i>Phaeophyscia orbicularis</i> | K | Motiejūnaitė et al. 2016 |
| 26. | <i>Lichenochora weillii</i> (Werner) Hafellner & R. Sant. | <i>Physconia enteroxantha</i> | K | Motiejūnaitė et al. 2016 |
| 27. | <i>Lichenocodium erodens</i> M.S.Christ. & D.Hawksw. | <i>Evernia prunastri</i> , <i>Hypogymnia physodes</i> , <i>Parmeliopsis ambigua</i> , <i>Ramalina fraxinea</i> | LV | Motiejūnaitė et al. 2006, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |
| 28. | <i>Lichenocodium lecanorae</i> (Jaap) D.Hawksw. | <i>Evernia prunastri</i> | K | Motiejūnaitė et al. 2016 |
| 29. | <i>Lichenocodium pyxidatae</i> (Oudem.) Petr. & Syd. | <i>Cladonia chlorophaea</i> | K | Motiejūnaitė et al. 2016 |
| 30. | <i>Lichenocodium usneae</i> (Anzi) D.Hawksw. | <i>Evernia prunastri</i> | K | Motiejūnaitė et al. 2016 |
| 31. | <i>Lichenocodium xanthoriae</i> M.S.Christ. | <i>Xanthoria parietina</i> | LV | Motiejūnaitė et al. 2016 |
| 32. | <i>Lichenodiplis lecanorae</i> (Vouaux) Dyko & D.Hawksw. | <i>Caloplaca</i> sp., <i>Myriolecis</i> aff. <i>hagenii</i> | K, L | Motiejūnaitė et al. 2016 |
| 33. | <i>Lichenosticta alcornaria</i> (Linds.) D.Hawksw. | <i>Cladonia coniocraea</i> , <i>C. macilenta</i> , <i>C. ochrochlora</i> | K, L | Czarnota and Kukwa 2010, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |

(Continued)

Table 1. Continued

| S. No. | Species | Host(s) | Distribution | Reference(s) |
|--------|--|--|--------------|---|
| 34. | <i>Marchandiomyces corallinus</i> (Roberge) Diederich & D.Hawksw. | <i>Physcia tenella</i> | K | Motiejūnaitė et al. 2016 |
| 35. | <i>Monodictys epilepraria</i> Kukwa & Diederich | <i>Lepraria</i> spp. | K | Czarnota and Kukwa 2010, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |
| 36. | <i>Muellerella hospitans</i> Stizenb. | <i>Bacidia rubella</i> | L, V | Czarnota and Kukwa 2010, Āboliņa et al. 2015 |
| 37. | <i>Microcalicium disseminatum</i> (Ach.) Vain. | <i>Chaenotheca</i> sp. | LV | Moisejevs 2015, DAU Herbarium |
| 38. | <i>Nectriopsis lecanodes</i> (Ces.) Diederich & Schroers | <i>Peltigera canina</i> , <i>Peltigera rufescens</i> | LV | Motiejūnaitė et al. 2006, Āboliņa et al. 2015, RIGG Herbarium |
| 39. | <i>Nectriopsis rubefaciens</i> (Ellis & Everh.) M.S.Cole & D.Hawksw. | <i>Parmelia sulcata</i> | LV | Motiejūnaitė et al. 2016 |
| 40. | <i>Plectocarpon lichenum</i> (Sommerf.) D.Hawksw. | <i>Lobaria pulmonaria</i> | V | Strazdiņa et al. 2017 |
| 41. | <i>Pronectria anisospora</i> (Lowen) Lowen | <i>Hypogymnia physodes</i> | K, L, V | Motiejūnaitė et al. 2016 |
| 42. | <i>Pronectria leptaleae</i> (J.Steiner) Lowen | <i>Physcia aipolia</i> | K | Motiejūnaitė et al. 2016 |
| 43. | <i>Pronectria robergei</i> (Mont. & Desm.) Lowen | <i>Peltigera didactyla</i> , <i>P. extenuata</i> | K, L | Moisejevs 2017, DAU Herbarium |
| 44. | <i>Pronectria xanthoriae</i> Lowen & Diederich | <i>Xanthoria parietina</i> | LV | Motiejūnaitė et al. 2006, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |
| 45. | <i>Pyrenochaeta xanthoriae</i> Diederich | <i>Xanthoria parietina</i> | LV | Motiejūnaitė et al. 2016 |
| 46. | <i>Reconditella physconiarum</i> Hafellner & Matzer | <i>Physconia distorta</i> | K | Motiejūnaitė et al. 2016 |
| 47. | <i>Refractohilum intermedium</i> Cl.Roux & Etayo | <i>Pachyphiale fagicola</i> | K | Czarnota and Kukwa 2010, Āboliņa et al. 2015 |
| 48. | <i>Refractohilum peltigerae</i> (Keissl.) D.Hawksw | <i>Peltigera</i> spp. | K | Motiejūnaitė et al. 2016 |
| 49. | <i>Sphinctrina turbinata</i> (Pers.) De Not. | <i>Pertusaria pertusa</i> | K, V | Motiejūnaitė et al. 2016 |
| 50. | <i>Stigidium microspilum</i> (Körb.) D.Hawksw. | <i>Graphis scripta</i> | K | Motiejūnaitė et al. 2016 |
| 51. | <i>Taeniolella punctata</i> M.S.Christ. & D.Hawksw. | <i>Graphis scripta</i> | K | Motiejūnaitė et al. 2016 |
| 52. | <i>Teloggalla olivieri</i> (Vouaux) Nik.Hoffm. & Hafellner | <i>Xanthoria parietina</i> | K | Motiejūnaitė et al. 2016 |
| 53. | <i>Thelocarpon epibolum</i> var. <i>epibolum</i> Nyl. | <i>Peltigera neckeri</i> | K | Motiejūnaitė et al. 2016 |
| 54. | <i>Tremella candelariellae</i> Diederich & Etayo | <i>Candelariella</i> sp. | LV | Czarnota and Kukwa 2010, Āboliņa et al. 2015 |
| 55. | <i>Tremella cetrariicola</i> Diederich & Coppins | <i>Nephromopsis chlorophylla</i> | K | Motiejūnaitė et al. 2006, Āboliņa et al. 2015 |
| 56. | <i>Tremella hypogymniae</i> Diederich & M.S.Christ. | <i>Hypogymnia physodes</i> | K | Motiejūnaitė and Grochowski 2014, Motiejūnaitė et al. 2016 |
| 57. | <i>Tremella lichenicola</i> Diederich | <i>Violella fucata</i> | K | Motiejūnaitė et al. 2016 |
| 58. | <i>Tremella phaeophysciae</i> Diederich & M.S.Christ. | <i>Phaeophyscia orbicularis</i> | K | Motiejūnaitė et al. 2016 |
| 59. | <i>Vouauxiella lichenicola</i> (Linds.) Petr. & Sydow | <i>Lecanora chlorotera</i> , <i>L. pulicaris</i> | K, V | Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |
| 60. | <i>Vouauxiomyces santessonii</i> D.Hawksw. | <i>Platismatia glauca</i> | K, V | Motiejūnaitė et al. 2016 |
| 61. | <i>Xanthoriicola physciae</i> (Kalchbr.) D.Hawksw. | <i>Xanthoria parietina</i> | K | Czarnota and Kukwa 2010, Āboliņa et al. 2015, Motiejūnaitė et al. 2016 |

Results

New records for Latvia

#*Arthrorhaphis aeruginosa* R.Sant. & Tønsberg

Distribution

Arthrorhaphis aeruginosa is known from Europe (Wirth et al. 2010, Motiejūnaitė 2017, Tsurykau 2017), including Fennoscandia (Santesson and Tønsberg 1994, Nordin et al. 2011), and also from Greenland (Alstrup et al. 2009), North America (Esslinger 2007), South America (Flakus et al. 2008) and Asia (Sohrabi and Alstrup 2007).

Material examined

Krāslavas Co., Ūdrišu Dist., Nature Park 'Daugavas loki', Tartaka Forest, ca 350 m W of Tartaks village, 55°53'6.9"N, 26°59'18.1"E, 150 m a.s.l., on side of old forest road in boreal forest with *Pinus sylvestris* and *Picea abies*, on primary thallus of *Cladonia* sp., 25 May 2018, leg. & det.: R.Moisejevs (DAU600000910).

Notes

The collected specimen was sterile, but it was recognized by the characteristic colour of the infected host thallus. Only *A. aeruginosa* is known to turn the infected lichen an aeruginose colour and, as stated in the protologue of the

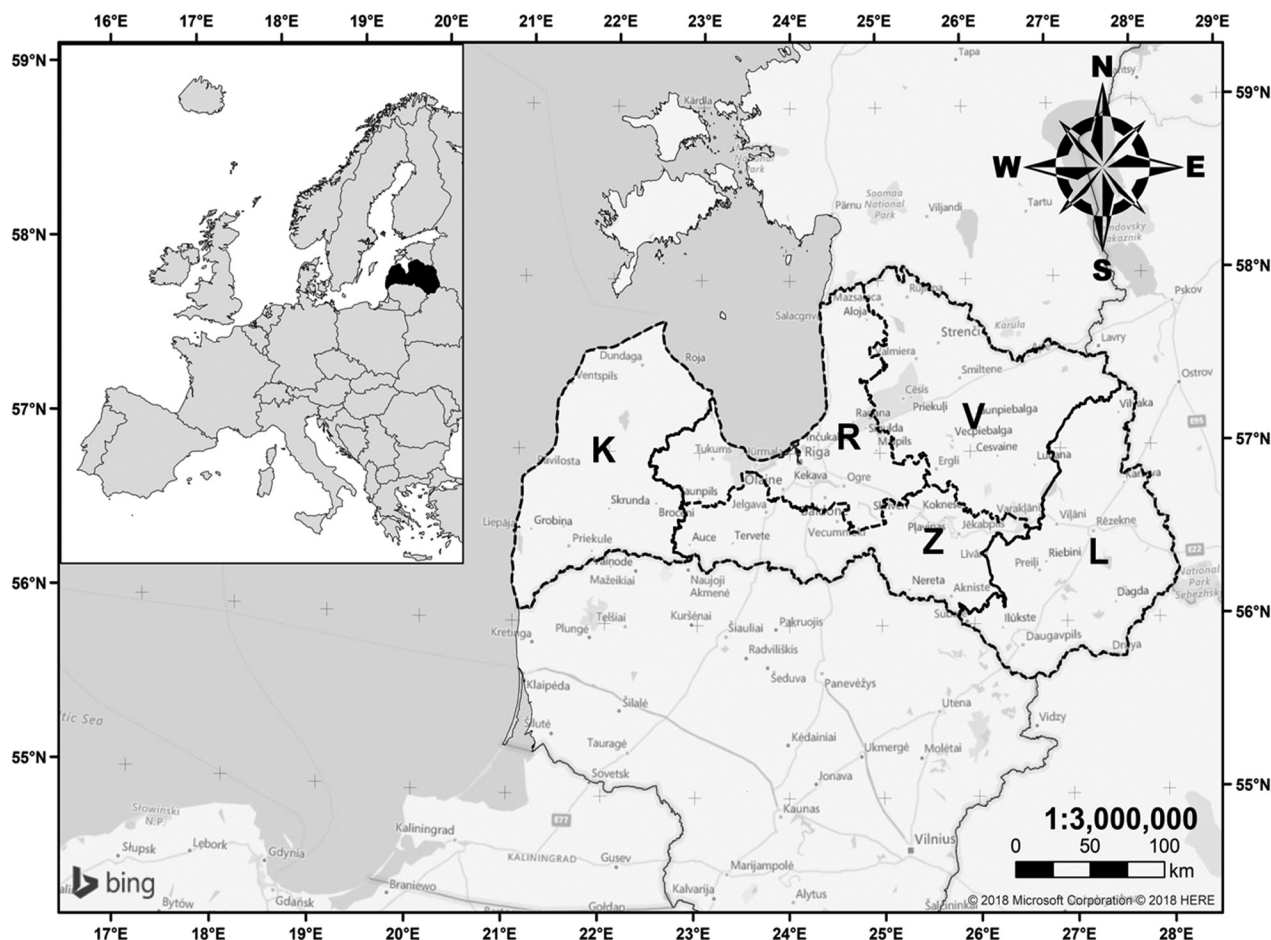


Figure 1. Regions of Latvia: K – Kurzeme (Curland), V – Vidzeme, L – Latgale, R – Pīrija, Z – Zemgale.

species ‘... is therefore easily recognized even when sterile’ (Santesson and Tønberg 1994).

Calicium pinastri Tibell

Distribution

Calicium pinastri is known from Europe (Tibell 1999, Śliwa and Kukwa 2008, Istomina and Likhacheva 2010, Wirth et al. 2010, Nordin et al. 2011) and North America (Hardman et al. 2017).

Material examined

Krāslava Co., Ūdrišu Dist., Nature Park ‘Daugavas loki’, Tartaka forest, ca 500 m SW of Tartaks village, 55°53’7.6”N, 26°59’30.8”E, 130 m a.s.l., old-growth dry boreal forest, on bark of *P. sylvestris*, 20 June 2018, leg. & det.: R.Moisejevs (DAU600000911).

Notes

The lichen was found growing close to *Calicium parvum*, a species that resembles *C. pinastri*, but has clavate asci, while *C. pinastri* has cylindrical asci when mature.

Chaenotheca laevigata Nád.

Distribution

Chaenotheca laevigata is a rare lichen with a wide distribution in Northern Hemisphere, being known from Europe

(Wirth et al. 2010, Nordin et al. 2011), Asia (Titov 2000) and North America (Hardman et al. 2017).

Material examined

Ventspils Co., Usmas Dist., Nature Reserve ‘Moricsala’, ca 400 m NE of guest house, 57°11’28.6”N, 22°8’12.0”E, 25 m a.s.l., in a humid old-growth deciduous forest with *Picea abies*, on the bark of *P. abies*, 9 July 2018, leg. & det.: R.Moisejevs (DAU600000917).

Notes

Chaenotheca laevigata can be confused with *Chaenotheca chlorella*, from which it differs by its immersed thallus, ellipsoid to short cylindrical ascospores and longer ascomata.

#Chaenothecopsis epithallina Tibell

Distribution

Chaenothecopsis epithallina is distributed in central Europe, Fennoscandia (Tibell 1975, Wirth et al. 2010, Nordin et al. 2011, Tsurykau 2017) and North America (Hardman et al. 2017).

Material examined

(1) Ventspils Co., Usmas Dist., Nature Reserve ‘Moricsala’, ca 400 m NE of guesthouse, 57°11’33.5”N, 22°8’12.3”E, 25 m a.s.l., in an old-growth deciduous forest, on thallus of *Chaenotheca trichialis* growing on the bark of *Quercus robur*,

9 July 2018, leg. & det.: R. Moisejevs; (2) Kocēnu Co., Dikļu Dist., ca 200 m S of Rāķis Lake, 57°35'34.1"N, 24°55'6.2"E, 120 m a.s.l., in a deciduous forest, on thallus of *C. trichialis* growing on the bark of old *Q. robur*, 23 March 2018, leg.: M. Kalniņš, det.: R. Moisejevs (DAU600000912).

Notes

Chaenothecopsis epithallina differs from the similar species *Chaenothecopsis nigra* by its association with *C. trichialis*, darker ascospores with less contrasting septum and dark green hypothecium.

Lecania croatica (Zahlbr.) Kotlov

Distribution

Lecania croatica is known from Europe (Printzen 1995, Mrak et al. 2004, Hafellner et al. 2005, Eichler et al. 2010, Vondrák et al. 2010, Kukwa et al. 2012, Motiejūnaitė 2017, Tsurukau 2017) and North America (Tønsberg 2004, Harris and Lendemer 2010).

Material examined

Daugavpils Co., Skrudalienas Dist., Nature Park 'Silene', ca 500 m N of Ilgas manor house, 55°41'54.5"N, 26°47'34.5"E, m a.s.l., in a deciduous forest with *Tilia cordata*, *Populus tremula* and *P. abies*, on the bark of *T. cordata*, 27 May 2018, leg. & det.: R. Moisejevs (DAU600000913).

Notes

The collected sterile specimen was checked using TLC, but no secondary compounds were found. The species was distinguished from species with a similar morphology and chemistry following the same characters as employed by Motiejūnaitė et al. (2012) and Tsurukau (2017).

Pycnora praestabilis (Nyl.) Hafellner

Distribution

Pycnora praestabilis is known in North America (Hodkinson 2009), Europe (Śliwa and Kukwa 2012, Randlane et al. 2016, Motiejūnaitė 2017), including Fennoscandia (Bendiksby and Timdal 2013).

Material examined

Daugavpils Co., Skrudalienas Dist., ca 3.5 km E of Silene town, 55°45'41.60"N, 26°52'58.08"E, 130 m a.s.l., in periphery of raised bog, on dry wood (snag) of *P. sylvestris*, 20 July 2017, leg. & det.: R. Moisejevs (DAU600000918).

Notes

From similar species of *Pycnora* and several morphologically similar species from *Xylopsora* genus, *P. praestabilis* differs in its lack of soredia, normally abundant pycnidia up to 0.3 mm diam., typical spot test reactions and presence of alectorialic acid.

Discussion

According to our data, 61 species of lichenicolous fungi have been recorded from Latvia. *Acolium inquinans*

mentioned by Āboliņa and Vimba (1959), who described it as 'a parasitic lichen, growing on thalli of other lichens', has been excluded since it is a lichenized species that lacks a lichenicolous habit (Tibell 1999). Specimen on which the record was based is lacking, therefore it is impossible to check its identity. Furthermore, Piterāns (1982) did not mention *A. inquinans* in his list of Latvian lichens and Āboliņa et al. (2015) described the species as an epiphytic lichen; therefore, it can be assumed that the aforementioned report was based on misidentification. Specimens of *Biatoropsis usnearum* reported by Motiejūnaitė et al. (2006) (both on *Usnea subfloridana*) are housed in the herbaria of the Institute of Botany, Nature Research Centre (BILAS) and University of Tartu (TU); the BILAS specimen and DAU specimens were checked in accordance with the description of *B. usnearum* (s. str.) given by Millanes et al. (2016), so it is assumed that only one species of *Biatoropsis* is known from Latvia.

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VI

REVISION OF THE GENUS *CETRELIA* (LICHENISED ASCOMYCOTA) IN LATVIA

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Abstract

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All available specimens (98) of the genus *Cetrelia* from Latvia (Northern Europe) in the Herbaria DAU and RIG were revised. *Cetrelia cetrarioides*, *C. olivetorum* and *C. monachorum* were confirmed to occur in the country. The last taxon is new to Latvia. Distribution maps and habitat preferences of all three species in Latvia were presented, and their conservation status was discussed.

Keywords: cetrarioid lichens, *Cetrelia monachorum*, chemotypes, conservation, new record, Parmeliaceae.

INTRODUCTION

The genus *Cetrelia* W.L. Culb. & C.F. Culb. (Parmeliaceae, Lecanorales, Ascomycota) has been described by CULBERSON & CULBERSON (1968). The genus is characterised by a large foliose, loosely attached greyish-green thallus with rounded marginal lobes and laminal pseudocyphellae on an upper cortex, at least partly black lower surface with sparse rhizines, ellipsoid ascospores, and atranorin as the main cortical substance with different orcinol depsides and depsidones as diagnostic medullary substances (CULBERSON & CULBERSON, 1968, 1976; OBERMAYER & MAYRHOFER, 2007; MARK et al., 2019). The European *Cetrelia* species are characterised by the presence of marginal soralia. In Europe, the *Cetrelia* species are mostly epiphytic and found in old natural or seminatural forests on tree bark or occasionally on mossy rocks (OBERMAYER & MAYRHOFER, 2007; KUKWA et al., 2012; DEGTJARENKO et al., 2018). The genus includes 18 currently accepted species (THELL et al., 2012; RANDLANE et al., 2013), and most species are restricted to eastern and south-eastern Asia (CULBERSON & CULBERSON, 1968; RANDLANE & SAAG, 1991). In Europe, only four species are

known: *Cetrelia cetrarioides* (Delise) W.L. Culb. & C.F. Culb., *Cetrelia chicitae* (W.L. Culb.) W.L. Culb. & C.F. Culb., *Cetrelia monachorum* (Zahlbr.) W.L. Culb. & C.F. Culb., and *Cetrelia olivetorum* (Nyl.) W.L. Culb. & C.F. Culb. (CULBERSON & CULBERSON, 1968; HAWKSWORTH et al., 2008, 2011). The taxonomic affinities of these species have been discussed over the years. Several studies have suggested accepting only one species in the group with four chemotypes, separated according to the major medullary substances, which can be visualised by the standard thin-layer chromatography methods (TLC) (CLERC, 2004; SANTESSON et al., 2004; GILBERT & PURVIS, 2009), while others have considered the different chemotypes as separate species (RANDLANE & SAAG, 1991; WIRTH et al., 2013). Recently, MARK et al. (2019) have shown that these chemotypes are correlated with phylogenetic clades. Recognition of the species based on TLC results is, therefore, confirmed to be possible. The *Cetrelia* synthesise specific sets of related substances – chemosyndromes, where the same product may occur in several species, but in different quantities, acting either as a major or a minor compound – specific to its taxonomic affiliation (CULBERSON & CULBERSON, 1968). *Cetrelia monacho-*

rum contains imbricatic acid as the major medullary substance and, additionally perlatolic, 4-O-demethylimbricatic, and anziaic acids in minor or trace amounts in the medulla (OBERMAYER & MAYRHOFER, 2007; KUKWA & MOTIEJŪNAITĒ, 2012). *Cetrelia cetrarioides* contains perlatolic acid as the major and imbricatic acid as minor medullary substance (OBERMAYER & MAYRHOFER, 2007; KUKWA & MOTIEJŪNAITĒ, 2012). *Cetrelia olivetorum* can be separated from the morphologically similar species *C. cetrarioides* and *C. monachorum* by a strong sanguineous/red reaction with the spot-test reaction with solution of sodium hypochlorite (C), which is caused by the presence of olivetoric acid in medulla (with trace of anziaic acid and 4-O-de-methylmicrophyllinic acid) (OBERMAYER & MAYRHOFER, 2007; GOLUBKOV et al., 2015). *Cetrelia chicitae* produces alectoronic acid and a-colatolic acid, including physodic acid and 4-O-methylphysodic acid in traces (OBERMAYER & MAYRHOFER, 2007). Also, all discussed *Cetrelia* species contain atranorin in the cortex.

In Latvia, two species, *C. olivetorum* and *C. cetrarioides*, have been recorded and distinguished by morphology and a spot-test reaction with solution of sodium hypochlorite (C) (ĀBOLIŅA et al., 2015). In the neighbouring countries, Estonia, Lithuania and Belarus, three species of the genus *Cetrelia* have been confirmed by TLC (KUKWA & MOTIEJŪNAITĒ, 2012; GOLUBKOV et al., 2015; DEGTJARENKO et al., 2018). To date, no particular chemical studies using TLC and distributional survey of the genus *Cetrelia* have been performed in Latvia. This study aimed to revise the *Cetrelia* taxa from available herbarium material to determine the species, their distribution and habitat preferences according to the European Union Habitats Directive in Latvia.

MATERIALS AND METHODS

The present study was based on herbarium material collected in Latvia from 1957 to 2018. A total of 98 specimens from the Lichenological Herbarium of Daugavpils University (DAU) and the University of Latvia (RIG) were studied. Morphology of the *Cetrelia* specimens was examined using a stereomicroscope. The specimens were tested for lichen substances by thin-layer chromatography (TLC; solvent A) (ORANGE et al., 2001). Localities of Latvian specimens and their

substrates were transcribed from the labels. Distribution maps of the *Cetrelia* species were created using ESRI ArcGIS pro 2.3. (ESRI, 2011).

RESULTS AND DISCUSSION

Heretofore, only two species, *Cetrelia olivetorum* and *C. cetrarioides*, had been known in Latvia (ĀBOLIŅA et al., 2015), *Cetrelia monachorum* was reported new to Latvia. *Cetrelia cetrarioides* and *C. olivetorum* accounted for 43% and 42% of the examined specimens, respectively. Amongst other *Cetrelia* species in Latvia, *Cetrelia monachorum* appeared to be the rarest – with 15% of the examined specimens. All studied material was sterile (without apothecia).

Cetrelia cetrarioides and *C. olivetorum* were first recorded in Latvia in 1957 (ANDRUŠAITIS et al., 1996) in the Numerne Forest near Donikava town. Since then, it has been recorded from other parts of Latvia as well (PITERĀNS, 2001; ĀBOLIŅA et al., 2015). Most collections of the *Cetrelia* are from the north-eastern and central parts of Latvia (Figs 1, 2). There are no records from the western part of the country. The distribution of *C. monachorum* is congruent with the distribution of other *Cetrelia* species (Fig. 3).

The three *Cetrelia* taxa are morphologically very similar, and difficult to distinguish based on morphology alone (RANDLANE & SAAG, 1991; OBERMAYER & MAYRHOFER, 2007; MARK et al., 2019). In our study, twenty-one *C. cetrarioides* specimens were previously determined as *C. olivetorum*, and seven *C. olivetorum* specimens as *C. cetrarioides*. A possible explanation

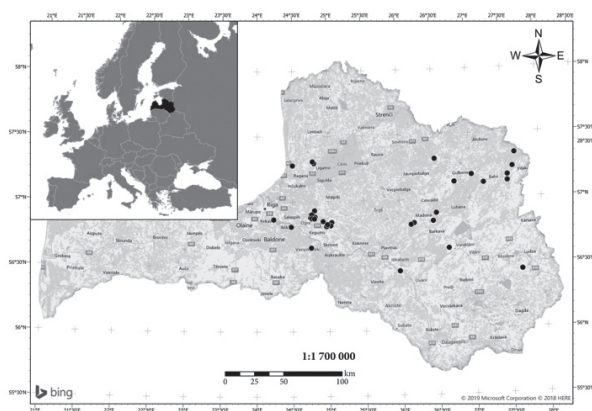


Fig. 1. Map of *Cetrelia cetrarioides* distribution in Latvia (marked with black circles); based on the herbarium collection from 1957 to 2018

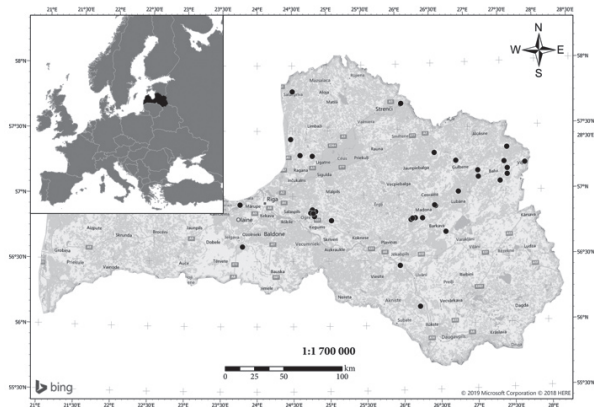


Fig. 2. Map of *Cetrelia olivetorum* distribution in Latvia (marked with black circles); based on the herbarium collection from 1957 to 2018

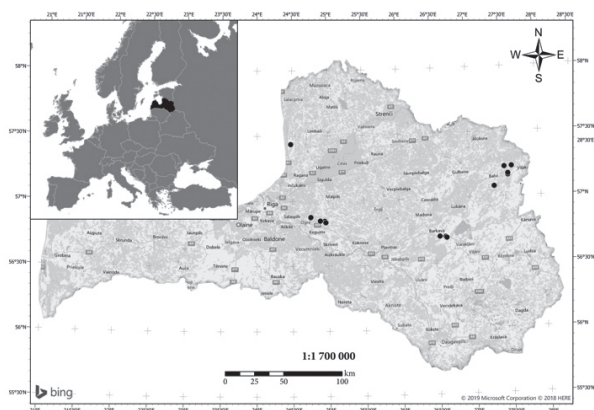


Fig. 3. Map of *Cetrelia monachorum* distribution in Latvia (marked with black circles); based on the herbarium collection from 1957 to 2018

for this might be that *C. olivetorum* and *C. cetrarioides* had been treated in Latvia as one species until 2001 (PITERĀNS, 2001). Since 2001, the determination of these taxa continued to be mainly based on morphology and spot-test reactions which are unreliable. Fifteen specimens were determined as *C. monachorum* (previously determined as *C. cetrarioides* and *C. olivetorum* and kept in the herbarium collections under *C. cetrarioides* or *C. olivetorum* names). The determination of lichen secondary substances using TLC proved to be a reliable tool to distinguish between these taxa (OBERMAYER & MAYRHOFER, 2007; KUKWA et al., 2012; MARK et al., 2019).

In the continental parts of Europe, the *Cetrelia* species grow mostly in old deciduous or mixed forests, in localities of high humidity (OBERMAYER & MAYRHOFER, 2007), especially in swamp forests or

forests surrounded by a lake or river or on hillsides near lakes or streams (RANDLANE & SAAG, 1991; KUKWA et al., 2012; KUKWA & MOTIEJŪNAITĒ, 2012). The *Cetrelia* habitat requirements in Latvia according to the European Union Habitats Directive (AUNIŅŠ et al., 2013) were based only on 33 recorded specimens from DAU and RIG. Data on habitat requirements according to the EU Habitats Directive were not used for other 65 specimens due to insufficient data on the specimen labels and lack of knowledge on the EU habitat distribution in Latvia. The examined *Cetrelia* specimens were mainly found at the sites, which correspond to the habitat type “Fennoscandian hemiboreal natural old broad-leaved deciduous forests (*Quercus*, *Tilia*, *Acer*, *Fraxinus* or *Ulmus*) rich in epiphytes” (habitat type code “9020”, 31% of the examined records), “Fennoscandian deciduous swamp forests” (habitat type code “9080*”, 24% of the examined records), and “Alluvial forests with *Alnus glutinosa* and *Fraxinus excelsior*” (habitat type code “91E0*”, 24% of the examined records).

Elsewhere, the *Cetrelia* species usually occur on deciduous trees and are very rarely found on coniferous trees (OBERMAYER & MAYRHOFER, 2007; KUKWA et al., 2012; KUKWA & MOTIEJŪNAITĒ, 2012). In Latvia, the *Cetrelia* species prefer mostly *Alnus glutinosa*, *Fraxinus excelsior*, *Populus tremula*, *Tilia cordata*, *Quercus robur* and less frequently *A. incana*, *Betula pendula*, *Corylus avellana*, *Padus avium* and *Picea abies* (Table 1). The habitat preferences for 18 specimens were not recorded on the specimen labels (Table 1).

In the Baltic countries, the *Cetrelia* species are rare and threatened. In Estonia, *C. olivetorum* and *C. cetrarioides* are considered as endangered (EN) and *C. monachorum* as critically endangered (CR; LÖHMUS et al., 2019). In Lithuania, all species of the genus *Cetrelia* are considered as EN (Jurga Motiejūnaitė pers. comm.). In Fennoscandia, for example, in Sweden, *C. olivetorum* s. lat. is treated as CR (NORDIN et al., 2011; ARTFAKTA, 2019), and in Finland, *C. olivetorum* – as EN (HYVÄRINEN et al., 2019). In neighbouring Belarus, *C. olivetorum* and *C. monachorum* are treated as EN, and *C. cetrarioides* – as CR (BELY et al., 2014). *Cetrelia olivetorum* is classified in Category I of the Latvian Red Data Book; the assessment of status was based on only seven herbarium specimens and one historical record (ANDRUŠAITIS et al., 1996). The status of newly reported *C. monachorum* has not been

Table 1. Substrate requirements for *Cetrelia cetrarioides*, *Cetrelia monachorum* and *Cetrelia olivetorum* in Latvia; based on the herbarium collection from DAU and RIG

| Substrate | <i>C. cetrarioides</i> | <i>C. monachorum</i> | <i>C. olivetorum</i> | Total |
|---------------------------|------------------------|----------------------|----------------------|-------|
| <i>Alnus glutinosa</i> | 13 | 4 | 12 | 29 |
| <i>Alnus incana</i> | 2 | 1 | 1 | 4 |
| <i>Betula spp.</i> | 2 | 0 | 1 | 3 |
| <i>Corylus avelana</i> | 0 | 0 | 1 | 1 |
| <i>Populus tremula</i> | 2 | 0 | 3 | 5 |
| <i>Fraxinus excelsior</i> | 11 | 2 | 7 | 20 |
| <i>Picea abies</i> | 1 | 0 | 0 | 1 |
| <i>Tilia cordata</i> | 1 | 3 | 5 | 9 |
| <i>Quercus robur</i> | 0 | 3 | 3 | 6 |
| <i>Padus avium</i> | 1 | 1 | 0 | 2 |
| Unknown | 8 | 2 | 8 | 18 |

determined yet. *C. olivetorum* is also included into the Legislative List (Rules of Ministry Cabinet) of Legally Protected Lichens in Latvia (REGULATIONS, 2013). For the current species, micro-reserves of up to 30 ha might be established. The assessment of a threat status of Latvian *Cetrelia* species based on IUCN system (STANDARDS..., 2006) has not been performed yet. Further work needs to be carried out to assess the threat status of all three *Cetrelia* species in Latvia according to the IUCN system and reconsider their conservation status for the Latvian Legislative List of Legally Protected Lichens. It is also recommendable to include *C. olivetorum*, *C. monachorum* and *C. cetrarioides* into a species monitoring system (BARONIŅA, 2017) to estimate the size of species populations and potential threats for all three species.

Record new to Latvia

Cetrelia monachorum (Zahlbr.) W.L. Culb. & C.F. Culb.

C. monachorum is morphologically similar to other sorediate species of the genus *Cetrelia* and is characterised by the presence of (mainly) marginal soredia. The soralia of *C. monachorum* are often coarse and irregular in shape with soredia usually exceeding 40 µm in diam. On the upper part of cortex, typically small and raised pseudocyphellae are found, sometimes associated with large ones, the lower cortex often lacks pseudocyphellae (CULBERSON & CULBERSON, 1968, 1976; OBERMAYER & MAYRHOFER, 2007). Morphologically similar sorediate *Cetrelia* species (*C. cetrarioides*, *C. chicitae* and *C. olivetorum*) are distinguished from *C. monachorum* by their different content sec-

ondary metabolites profile (OBERMAYER & MAYRHOFER, 2007; RANDLANE & SAAG, 1991; MARK et al., 2019). The species is known in the following European countries: Austria (OBERMAYER & MAYRHOFER, 2007), Belarus (GOLUBKOV et al., 2015), Bosnia-Herzegovina (BILOVITZ & MAYRHOFER, 2011), the Czech Republic (OBERMAYER & MAYRHOFER, 2007), Estonia (DEGTJARENKO et al., 2018), France, Germany, Italy (OBERMAYER & MAYRHOFER, 2007), Lithuania (KUKWA & MOTIEJŪNAITE, 2012), Montenegro (OBERMAYER & MAYRHOFER, 2007; KNEZEVIC & MAYRHOFER, 2009), Norway (OBERMAYER & MAYRHOFER, 2007), Poland (Kukwa et al., 2012), Portugal (LLIMONA & HLADUN, 2001), Romania (OBERMAYER & MAYRHOFER, 2007), Russia (European part, Tver Oblast, STEPANCHIKOVA et al., 2011), Slovakia, Slovenia, (OBERMAYER & MAYRHOFER, 2007), Spain (BARBERO et al., 1995), Sweden (NORDIN et al., 2011), Switzerland (OBERMAYER & MAYRHOFER, 2007), Ukraine (KONDRATYUK et al., 2003), the United Kingdom (OBERMAYER & MAYRHOFER, 2007). It is also known from the Republic of Armenia, the Republic of Azerbaijan (SOHRABI et al., 2007), China (WEI, 1991), Georgia (OBERMAYER & MAYRHOFER, 2007; SOHRABI & ALSTRUP, 2007), Hawaii (ELIX & MCCARTHY, 1998), India (MISHRA & UPRETI, 2015), Japan (KUROKAWA, 2003), North American mainland (OBERMAYER & MAYRHOFER, 2007), and from the Asian part of Russia (URBANAVICHUS & ANDREEV, 2010).

Specimens examined. Ogres distr., Suntažu Forest, forest square – 88, forest site 5, on unknown substrate, 16 Oct. 2002, I. Grīsele (RIG15910). Viļakas distr., Žīguru Forest, forest square – 504, forest site 2, on *A. glutinosa*, 2 Oct. 2012, V. Lārmanis (RIG16132).

Ogres distr., Suntažu Forestry, forest square – 256, forest site 12, on *A. glutinosa*, 16 Oct. 2002, I. Grīšle (RIG15925). Ogres distr., Suntažu Forestry, forest square – 79, forest site 8, on *A. incana*, 16 Oct 2002, I. Grīšle (RIG15917). Balvu distr., Kupravas Forest on *T. cordata*, 3 Oct. 1996, A. Piterāns (RIG14313). Balvu distr., Tilža, Dominava Forest of high humidity on *F. excelsior*, 13 Aug. 1957, A. Piterāns (RIG2038). Balvu distr., Kupravas Forest on *T. cordata*, 3 Oct. 1996, A. Piterāns (without number). Ogres distr., Suntažu Forestry, forest square – 274, forest site 10 on unknown substrate, in 2002, I. Grīšle (RIG15905). Ogres distr., Suntažu Forestry, forest square – 157, forest site 8, on *A. glutinosa*, 1 Dec. 2002, I. Grīšle (RIG15908). Madonas distr., Barkavas Rural Municipality, on *Q. robur*, 22 May 1997, A. Piterāns (RIG15353). Madonas distr., on *Q. robur*, 22 May 1997, A. Piterāns (RIG15352). Madonas distr., Nature ‘Barkavs ozolu audze’, *Q. robur* forest, 22 May 1997, A. Piterāns (RIG15245). Balvu distr., Vīksnas Rural Municipality, Nature Reserve ‘Kupravas liepu audze’, *T. cordata* forest, 26 May 1976, A. Piterāns (RIG15244). Salacgrīvas distr., 2 km to the east of Mustkalni town, on *A. glutinosa* in alluvial swamp forest (91E0*), 1 July 2017, R. Moisejevs (DAU600000477). Krimulda distr., about 3 km of Lēdurga town to the east, on *F. excelsior*, in old-growth broadleaved forest dominated by *Q. robur*, 19 Sep. 2016, R. Moisejevs (DAU600000476).

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CETRELIA GENTIES (LICHENIZUOTI AUKŠLIAGRYBŪNAI) REVIZIJA LATVIJOJE

Polina DEGTJARENKO, Rolands MOISEJEVS

Santrauka

Buvo patikrinti visi DAU ir RIG herbariumuose laikomi *Cetrelia* pavyzdžiai (iš viso 98), surinkti Latvijoje. Inventorizacija patvirtino, kad šalyje aptinkamos trys *Cetrelia* genties rūšys: *Cetrelia cetrarioides*,

C. olivetorum ir *C. monachorum*. Pastaroji rūšis yra nauja Latvijos lichenobiotai. Straipsnyje pateikiami visų trijų rūšių paplitimo žemėlapiai bei jų buveinės Latvijoje, aptariamas rūšių apsaugos statusas.

VII

SOME NEW TO LATVIA LICHENS AND ALLIED FUNGI

Rolands Moisejevs

Moisejevs R. 2015. Some new to Latvia lichens and allied fungi. *Acta Biol. Univ. Daugavp.*, 15 (2): 285 – 292.

The paper presents data on five lichenized, lichenicolous and saprobic fungus species that are new to Latvia. *Chaenothecopsis pusiola*, *Scytinium teretiusculum*, *Microcalicium ahlneri*, *M. disseminatum* and *Pilophorus cereolus* were found in Vidzeme region in year 2015. Of them, *Scytinium teretiusculum* and *Pilophorus cereolus* are rare in all Baltic States.

Key words: lichenized fungi, lichenicolous fungi, non-lichenized fungi.

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INTRODUCTION

Lichens are poorly studied group of organisms in Latvia. Current total number of lichen and allied fungus species found in Latvia is the smallest in Baltic States – 682 in total (Motiejūnaitė et al., in prep.). During last 25 year new records of lichens and allied fungi were reported as a fieldwork result of the Symposia of Baltic Mycologists and Lichenologists (Motiejūnaitė & Piterāns 1998, Motiejūnaitė et al. 2006, Czarnota & Kukwa 2010), by visiting lichenologists from other countries (Sundin & Thor 1990, Motiejūnaitė & Grochowski 2014), local ecological studies (Plociņa 2007, Mežaka 2009, Mežaka et al. 2008, 2009, 2012), and some targeted lichen diversity studies (Piterāns et.al. 2005, 2006). Current paper supplements data on biota of lichens and allied fungi in Latvia and expands knowledge on distribution of some rare and understudied species.

MATERIAL AND METHODS

Specimens of lichenized and allied fungi were collected by the author of present paper in year

2015, in two Natura 2000 and one industrial forest territories in Vidzeme region (Fig 1.). Collected specimens were identified using the routine lichenological methods (Smith et al. 2009). Specimens were examined under stereomicroscope Nikon SMZ 800. Watermounted hand-made cross sections were examined under light transmission microscope Nikon Eclipse E100. Measurements of ascospores and other structures were made in water. For determination of collected specimens following references were employed: Ahti & Steinross (2013), Jørgensen (2007), Smith et al. (2009), Tibell (1999). The nomenclature of taxa mainly follows Nordin et al. (2011). Some specimens were compared with the collections at the herbarium of the Institute of Botany, Nature Research Centre, Vilnius (BILAS). Species distribution in neighboring countries (except Russia) is provided. Non-lichenized fungi are marked with (+) and lichenicolous fungi with (#). All lichen specimens that are presented in this paper are deposited at the Herbarium of Daugavpils University, Institute of Life Sciences and Technology, Laboratory of Botany (DAU).

Study site

Vidzeme is situated in northeast part of Latvia and occupies area of 25 683 km² (including Riga and Pierīga districts), with highest population in the districts of Riga, Jūrmala, Valmiera and Cēsis. (Turlājs 2011). There are 7 Nature Parks, 4 Protected Landscape Areas, 1 Biosphere Preserve, 1 National Park and over 70 Nature Preserves. (Dabas aizsardzības pārvalde 2015). The region is bordered by Riga Gulf in the west. Highest areas above the sea level are Vidzemes and Alūksnes uplands with highest points „Gaiziņkalns” 311,6 m and „Dēliņkalns” 271,5. Largest river in Vidzeme is Gauja with total length of 452 km and catchment area over 7.700 km² in Latvia. Sandstone outcrops are found along the banks and in the catchment of the river (Turlājs 2011).

Climate of Vidzeme region is mainly formed by Atlantic cyclones that bring air masses and precipitation from west and northwest. Average temperature in January ranges between -7 and -3 °C, in July +16 and +18°C. Average yearly precipitation ranges between 600 and 850 mm,

snow cover remains for 80–110 days (Turlājs 2011).

A number of forest and outcrop habitat types included into Annex I of European Union Habitat Directive are found in the region, such as 6530 *Fennoscandian wooded meadows, 9010 *Western taiga, 9020 *Fennoscandian hemiboreal natural old broad-leaved deciduous forests rich in epiphytes, 9050 Fennoscandian herb-rich forests with *Picea abies*, 9080 *Fennoscandian deciduous swamp woods, 9160 Sub-Atlantic and medio-European oak forests, 9180 *Tilio-Acerion* forests of slopes, screes and ravines, 91E0 *Alluvial forests with *Alnus glutinosa* and *Fraxinus excelsior*, 91D0 Bog woodland, 8220 Siliceous (sandstone) rocky slopes with chasmophytic vegetation, 8210 Calcareous rocky slopes with chasmophytic vegetation and some others (Lārmanis 2013, Rēriha 2013).

RESULTS AND DISCUSSION

All referred specimens were collected during the year 2015 in Vidzeme. *Chaenothecopsis pusiola*,

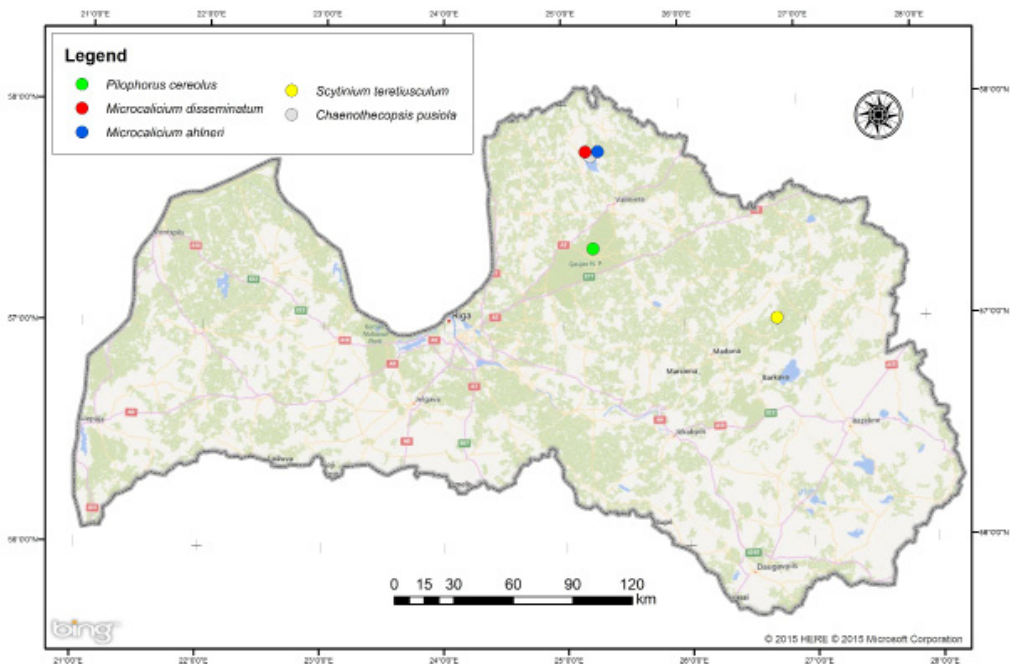


Fig 1. Study areas.

Scytinium teretiusculum, *Microcalicium ahlneri*, *Microcalicium disseminatum*, *Pilophorus cereolus* are reported here for the first time in Latvia. Of the five newly recorded species, *Scytinium teretiusculum* and *Pilophorus cereolus* are rare and red-listed in adjacent countries whenever they are known (Randlane & Saag 2000, Randlane et al. 2008, Rašomavičius 2007).

SPECIES LIST

+ *Chaenothecopsis pusiola* (Ach.) Vain.

Morphology: Apothecia 0,3-0,55 mm tall, head 0,16- 0,2 mm diam. Ascospores 6-7x 2-2,4 µm, 1-septate, septum pale. Excipulum K+ bright red.

Species distribution and ecology: Widely distributed in boreal and montane zones of N.Hemisphere, occurs on lignum of conifers, less often on lignum of deciduous trees (Tibell 1999). Species is known from Lithuania and Estonia (Motiejūnaitė 1999; Löhmus 1998).

Material examined: Latvia, Valmieras region, Burtnieku district, Nature preserve „Vidusburtnieks”, (57°46'N, 25°14'E). On decaying *Quercus* wood in old shaded woodland with high humidity. May 16, 2015. (DAU Nr. 027004001)

Scytinium teretiusculum (Wallr.) Otálora et al.

Morphology: Thallus dark grey. Lobes 0,3-0,8x 0,1-0,2 mm. Isidia long, cylindrical 40- 65 µm diam and 0,6-0,8 mm long. Apothecia in Latvian material not found.

Species distribution and ecology: Europe, Macaronesia, N. and C. America, Asia. Corticolous on coarse-barked trees, especially *Populus tremula* or *Juniperus*. (Gilbert & Jørgensen 2009, Jørgensen 2007). Species known from Lithuania and Estonia (Motiejūnaitė et.al. 2005; Randlane & Saag 1999).

Material examined: Latvia, Lubānas region, Daukstas district, About 5km West off Lubāna



Fig 2. Old woodland from study area.

City (57°1'N 26°46'E). Broad-leaved forest with *Populus tremula*. On trunk of *Tillia cordata*. Old shaded woodland with high humidity (Fig 2). July 22, 2015 (DAU Nr. 064007001).

+ *Microcalicium ahlneri* Tibell (Fig 4.)

Morphology: Apothecia 0,4-0,5 mm tall. Stalk black, head narrowly conical. Ascospore mass green-black with sclerotized hyphae. Ascospores 5-6,5x 2-2,4 µm.

Species distribution and ecology: Known in Europe, N. America, Asia. On decorticated stumps heavily attacked by brown rot fungi, usually of *Pinus sylvestris*, more rarely on lignum of *Picea abies*, and occasionally on oak wood. Mainly in humid locations (Tibell 1999). Recorded from Lithuania and Estonia (Motiejūnaitė 2007; Randlane & Saag 1999).

Material examined: Latvia, Valmieras region, Burtnieku district, Nature preserve „Vidusburtnieks”, (57°46' N 25°14'E). On decaying wood of *Quercus robur*. Old shaded

woodland with high humidity. May 16, 2015 (DAU Nr. 143001001).

Microcalicium disseminatum (Ach.) Vain. (Fig. 5)

Morphology: On thalli of *Chaenotheca trichialis* (Ach.) Th. Fr. Apothecia 0,1-0,3 mm diam., 0,1-0,2 tall, sessile. True exciple aeruginose in section, K+ brown. Ascospores 11-13x 3-3,7 µm, 1- to 3-septate.

Species distribution and ecology: Reported from Europe, N. America, Asia. On lignum and bark of both deciduous trees and conifers in moderately shaded situations. Often parasitic on calicioid lichens, particularly *Chaenotheca* species. Also on free-living algal colonies and apparently also occurring saprobically on wood (Tibell 1999). Species known from Lithuania and Estonia. (Motiejūnaitė 2003; Randlane & Saag 1999).

Material examined: Latvia, Valmieras region, Burtnieku district, Nature preserve „Vidusburtnieks”, (57°46' N 25°14'E). On trunk



Fig. 3. Habitat of *Pilophorus cereolus*.



Fig. 4. *Microcalicium ahlneri*.

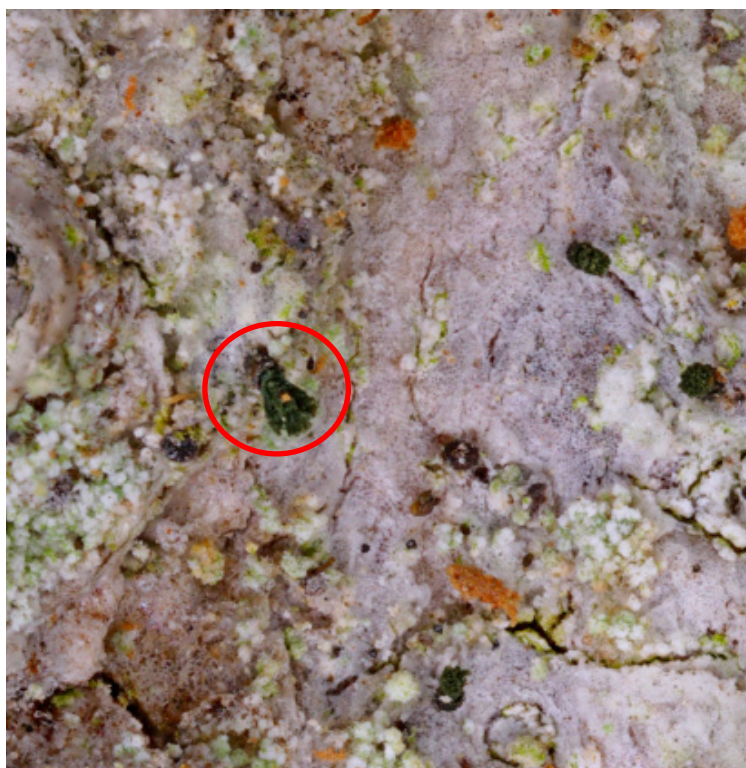


Fig 5. *Microcalicium disseminatum*.

of *Quecus robur*. Old shaded woodland with high humidity. May 16, 2015. (DAU Nr. 143002001)

***Pilophorus cereolus* (Ach.) Hellb.**

Morphology: Crustose, thallus minutely granular, containing soredia and cephalodia. Cephalodia with *Nostoc*. Pseudopodetia 3-5,7 mm high, containing black pycnidia at apices. K+ yellow.

Species distribution and ecology: Reported from Scandinavia, Estonia and Central Europe, Greenland, W and E coast of North America. Grows on siliceous stones and rocks in very shaded, moist places. (Ahti & Stenroos 2013). *P. cereolus* is rare in Baltic region and was found only in Estonia previously (Randlane et al. 2008). In Estonia species is red-listed.

Material examined: Latvia, Cēsu region, Priekule district, Gauja National Park, Ērgļu (Ērgļu) Klintis, (57°21'N 25°15'E). On sandstone outcrops, about 2-3 m from the base (Fig 3).. April 17, 2015 (DAU 144001001)

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